

FINAL REPORT

Metals and pH TMDLs for the Paint Creek Watershed, West Virginia

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For:

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1.0 Introduction

Section 303(d)(1)(A) of the Federal Clean Water Act (CWA) states:

“Each state shall identify those waters within its boundaries for which the effluent limitations required by section 301(b)(1)(B) are not stringent enough to implement any water quality standards applicable to such waters. The State shall establish a priority ranking for such waters taking into account the severity of the pollution and the uses to be made of such waters.”

Further, section 303(d)(1)(C) states:

Each state shall establish for the waters identified in paragraph (1)(A) of this subsection, and in accordance with the priority ranking, the total maximum daily load, for those pollutants which the Administrator identifies under section 304(a)(2) as suitable for such calculations. Such load shall be established at a level necessary to implement the applicable water quality standards with seasonal variations and a margin of safety which takes into account any lack of knowledge concerning the relationship between effluent limitations and water quality.

Two main stem segments and sixteen tributaries have been placed on West Virginia's section 303(d) list of impacted waters (Table 1-1). The objective of this TMDL Report is to develop pH and metals TMDLs for the streams in the Paint Creek watershed impaired by mine drainage. Two main stem segments and fourteen tributaries are listed for the pollutants pH and/or metals with the impairment attributed to mine drainage. Three additional tributaries are listed for biological impairment. The causative pollutant(s) and source(s) of the biologically impaired streams are not identified.

1.1 Problem Understanding

Paint Creek, a tributary of the Kanawha River, flows in a northerly direction through parts of Raleigh, Fayette and Kanawha counties in south-central West Virginia (Figure 1-1) and has a drainage area of 318 km² (123 mi²). For the past 90 years, surface and deep coal mines have operated in the watershed. Before the implementation of the West Virginia Surface Coal Mining and Reclamation Act (WVSCMRA) and Surface Mining Control and Reclamation Act (SMCRA), little consideration was given to the environmental degradation that resulted from these activities. Currently, the quality of Paint Creek and its tributaries are being negatively impacted by the acidic drainage from those mines that were abandoned prior to the environmental regulations. The environmental impact of

this mine drainage is being manifested in depressed stream pH and elevated concentrations of iron, manganese and aluminum.

Sycamore Branch, Hickory Camp Branch and South Sand Branch are the Paint Creek tributaries that are listed as biologically impaired. Hickory Camp Branch is also listed for pH and metals with impairment attributed to mine drainage. Implementation of this TMDL should restore the biological integrity of Hickory Camp Branch. The biological impairment of South Sand Branch and Sycamore Branch cannot be attributed to active or historical mining activities. Further evaluation and TMDL development is needed to address these impaired waters.

Table 1-1. Paint Creek Watershed Segment on Section 303(d) list.

Stream Segment Name	Segment ID	Length (miles)	Year Listed	Trout Stream	pH	Fe	Mn	Al	BC ¹
Jones Branch	K-65-C	1.43	1996	No		X	X	X	
Packs Branch	K-65-DD	3.80	1996	No		X	X	X	
Big Fork of Packs Branch	K-65-DD-2	1.24	1996	No		X	X	X	
Sycamore Branch ²	K-65-L	1.49	1998	No					X
Ten Mile Fork ³	K-65-M	2.44	1996	No	X	X	X	X	
Long Br. of Ten Mile Fork	K-65-M-1	1.43	1996	No	X	X	X	X	
Hickory Camp Branch ⁴	K-65-P	3.80	1996	No	X	X	X	X	X
Cedar Creek	K-65-Q	1.20	1998	No	X				
Unnamed Tributary #1	K-65-Q.3	0.36	1998	No	X	X	X	X	
Unnamed Tributary #2	K-65-Q.5	0.44	1998	No	X	X	X	X	
Fifteen Mile Creek	K-65-R	1.24	1996	No		X	X	X	
Spring Branch	K-65-S	1.30	1998	No	X				
Skitter Creek	K-65-T	1.48	1998	No		X	X	X	
Lykins Creek	K-65-W	4.62	1996	No	X	X	X	X	
Long Branch of Mossy Creek	K-65-Y-2	2.43	1996	No		X	X	X	
South Sand Branch ²	K-65-HH-2	3.97	1998	No					X
Paint Creek (Mouth - 16.8)	K-65	16.80	1996	Partially				X	
Paint Creek (MP 11.0 - 16.8)	K-65	5.80	1996	Partially	X				

¹Biological Criteria.

²Length of entire stream given.

³Corrected length of Ten Mile Fork from 303(d) list.

⁴Length of listed stream segment obtained from AMD portion of the 303(d) list.

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¹Biological Criteria.

²Length of entire stream given.

³Corrected length of Ten Mile Fork from 303(d) list.

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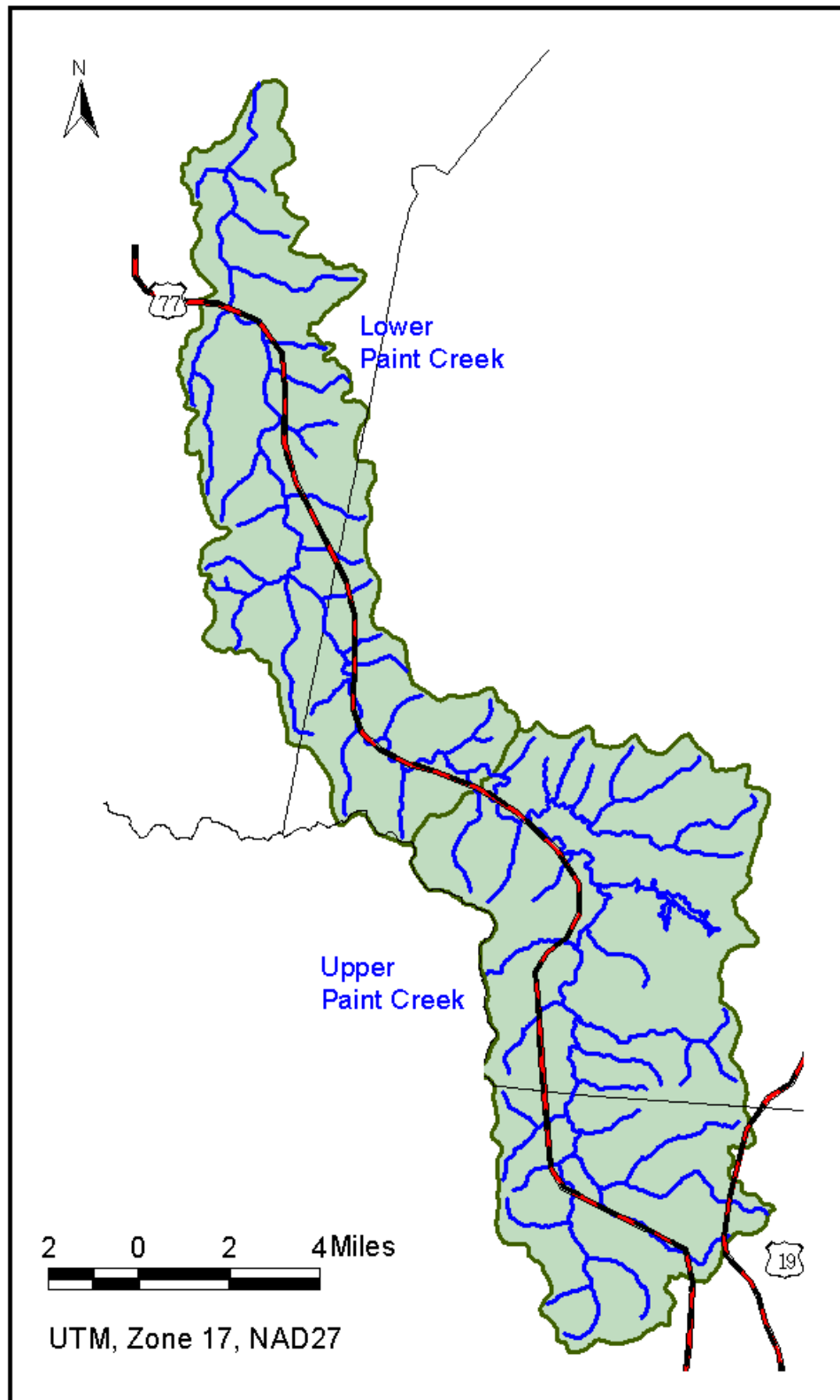


Figure 1-1. Paint Creek Watershed

2.0 Water Quality Standards

The State of West Virginia's *Requirements Governing Water Quality Standards* (WVWQS, 1999) have defined water quality criteria for surface waters as a numeric constituent concentration or a narrative statement representing a quality of water that supports a designated use or uses of the body of water. Aluminum, iron and manganese concentration and stream pH are given numeric criteria under the Aquatic Life and the Human Health use designation categories (Table 2-1). All listed stream segments in the Paint Creek watershed have been designated as having an Aquatic Life and a Human Health use (WVDEP, 1998a). The main stem of Paint Creek between Pax, WV, and Burnwell, WV, has also been identified as a trout stream (WVDNR, 2001). This segment must meet the Aquatic Life B2 criteria.

Table 2-1. Applicable West Virginia Water Quality Criteria.

Pollutant	Use Designation				
	Aquatic Life				Human Health
	B1, B4		B2		
	Acute ^a	Chronic ^b	Acute ^a	Chronic ^b	A ^c
Total Aluminum, mg/L	0.75	-	0.75	-	-
Total Iron, mg/L	-	1.50	-	0.50	1.50
Manganese, mg/L	-	-	-	-	1.00
pH	No values below 6.0 or above 9.0.				

Source: WVWQS, 1999; B1 = Warm water fishery stream, B4 = Wetlands, B2 = Trout waters, A = Water supply, public.

^a One-hour average concentration not to be exceeded more than once every three years on the average.

^b Four-day average concentration not to be exceeded more than once every three years on the average.

^c Not to exceed.

3.0 Source Assessment

This section examines and identifies the potential sources of acidity, aluminum, iron, and manganese in the Paint Creek watershed. Paint Creek watershed stream segments that are impaired due to depressed pH and elevated aluminum, iron and manganese concentrations are impaired by acidic mine drainage (AMD).

3.1 Acidic Mine Drainage

AMD forms when sulfide minerals are exposed to oxidizing conditions in coal and metal mining, highway construction, and other large-scale excavations. In coal mining regions, iron sulfides are predominately pyrite and marcasite (FeS_2). Upon exposure to H_2O and O_2 , sulfide minerals oxidize to form acidic, sulfate-rich drainage. Metal composition and concentration in AMD depend on the type and quantity of sulfide minerals present. The drainage quality emanating from underground mines or surface mine backfills is dependent on the amount of acid producing (sulfide) and alkaline (carbonate) minerals contained in the disturbed rock. In general, disturbing sulfide-rich and carbonate-poor rock produces acidic drainage and sulfide-poor and carbonate-rich rock produces alkaline drainage. Disturbed carbonate-rich rock can produce alkaline drainage even with significant sulfide concentrations.

The acidity in the AMD produced in coal mines is comprised of mineral acidity (Fe, Al, Mn) and H^+ acidity. Approximately 20,000 km (12,000 mi) of streams and rivers in the United States are degraded by AMD. About 90% of the AMD reaching streams originates in abandoned surface and deep mines. Since no company or individual is responsible for reclaiming abandoned mine lands (AML), little treatment of the drainage occurs and the contamination of surface and subsurface water resources continues unabated (Skousen, Sexstone and Ziemkiewicz, 2000).

The oxidation of iron disulfides and subsequent conversion to acidity occur through several reactions (Stumm & Morgan, 1970), which are detailed in Chapter 5 of Geidel and Caruccio (2000). If any of the processes represented by these equations are slowed or stopped, the generation of AMD is also slowed or ceases. Removal of air and/or water, two of the three principal reactants, from the system will stop pyrite oxidation. This occurs naturally when pyrite-bearing rocks are saturated. Because small amounts of pyrite are oxidized through weathering in undisturbed environments, only small amounts of acidity are generated, which are quickly diluted and neutralized by surrounding rocks. However, when large volumes of pyritic materials are exposed to oxidizing conditions, the pyrite reacts on a large scale and the reaction products (Fe, sulfate and acidity) are carried into surface and subsurface waters by runoff and infiltration. The generation of acidity is greatly increased by the oxidation of

ferrous iron into ferric iron. This process is rather slow under abiotic (no life) conditions, but iron-oxidizing bacteria, *Thiobacillus ferrooxidans*, greatly increase the kinetic rate of this reaction (Waksman, 1922). The presence of the biotic catalysts from these bacteria can increase the overall generation of acidity by a factor of one million (Leathen et al., 1953).

3.2 Point Sources

Permitted discharges are point sources at discrete locations in the Paint Creek watershed and can be classified into two major categories: mining point sources and non-mining point sources. Identified non-mining point sources in the watershed include small sewage treatment facilities, small commercial facilities registered under an industrial stormwater general WV/NPDES permit, and construction sites registered under a construction stormwater general WV/NPDES permit. These sources do not discharge significant amounts of the pollutants of concern of this study and are not considered further. Mining point sources include various coal mining operations, as well as sandstone quarries, and are the focus of this report.

The WVDEP Office of Mining and Reclamation (OMR) regulates deep and surface coal mines and quarries and issues National Pollutant Discharge Elimination System (NPDES) permits for the discharges from mining operations. NPDES permits contain effluent limitations and/or self-monitoring requirements for the pollutants of concern of this TMDL. The results of self-monitoring are regularly reported to the WVDEP in Discharge Monitoring Reports (DMRs).

Additionally, OMR issues permits for mineral extraction, preparation, refuse disposal and ancillary operations pursuant to West Virginia Code Chapter 22 Article 3 and Article 4, respectively. Such permits are known as “Article 3” and “Article 4” permits. Multiple Article 3 and Article 4 permits can be associated with an individual NPDES Permit.

In this TMDL Report, NPDES Permit and outlet numbers are used to identify individual point sources. The primary advantage of organization by NPDES Permits is that the individual outlets (Figure 3-1) are located in WVDEP’s databases, whereas Article 3 and 4 permits are located by the reported latitude and longitude of the geographical center of mining operations. Effluent limits and DMR data are also available by NPDES outlet. The NPDES permit outlet locations and associated information facilitated model calibration, development of the baseline condition (See Section 6.2) and allocation model runs (See Section 6.4).

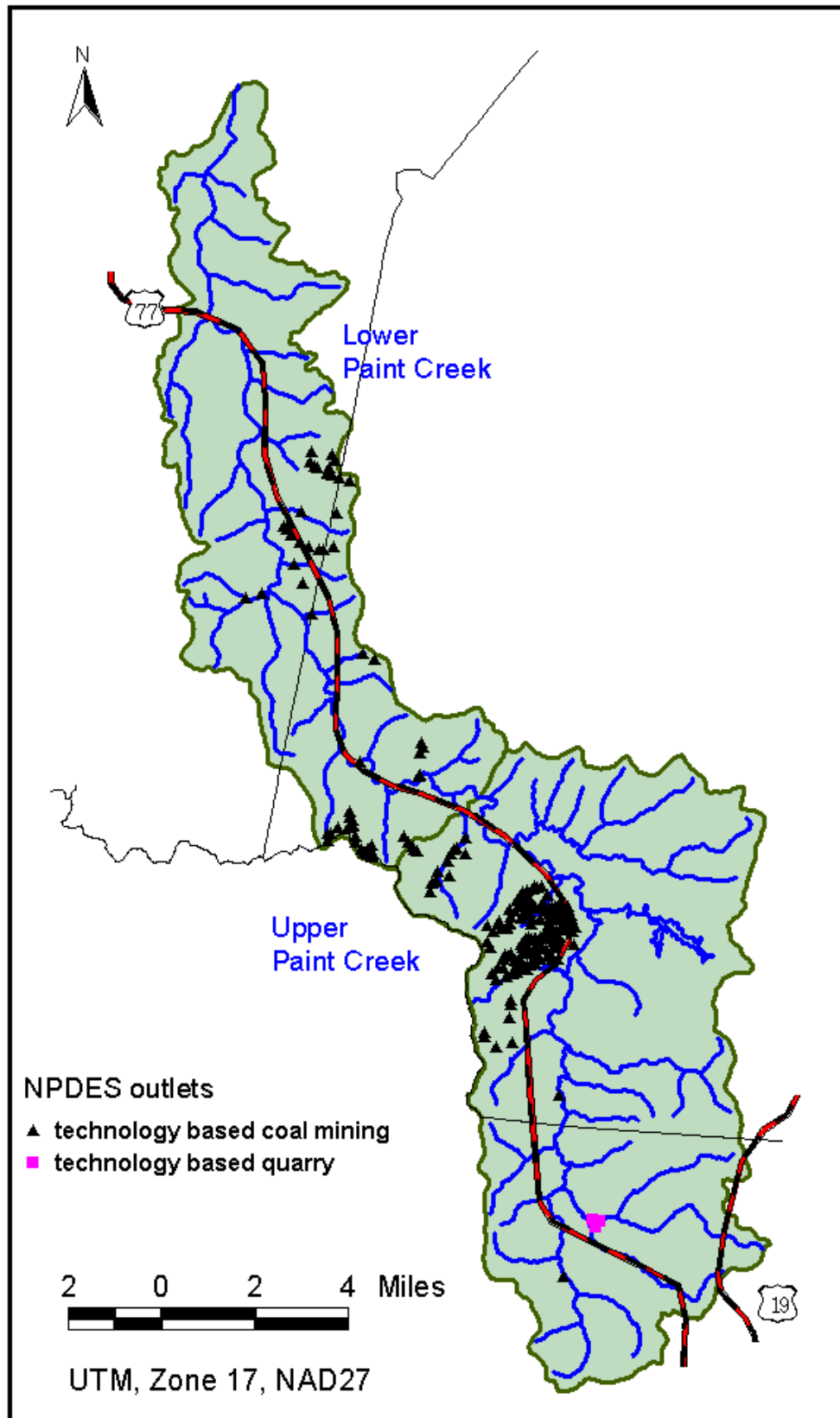


Figure 3-1. NPDES Permit types in the Paint Creek Watershed.

The WVDEP database information associated with Article 3 and Article 4 permits contains valuable information relative to the operational status of the mining activity. Facilities that have achieved “Phase 2 Release” status have performed reclamation to the point that treatment systems and drainage structures are removed. As such, those facilities were not considered in the baseline condition or in the allocation models. Table 3.1 describes the Article 3 / Article 4 status classifications. The association of Article 3 / Article 4 permit and operational status with NPDES permits was provided by WVDEP-OMR. Table A-1 in Appendix A presents the association between Article 3 / Article 4 permits and NPDES permits.

Table 3-1. Classifications of Article 3 and 4 Permit Status.

Status Class	Description	TMDL Application
New	Newly issued permit, may or may not have commenced discharge.	Assumed to be discharging in accordance with NPDES effluent limitations in base condition and allocation runs.
Renewed	Active mining facility.	
Inactive	Currently inactive operation but could become active at any time.	
Phase I Released	Active mining has ceased, site has been re-graded and reseeded; treatment facilities and outlet structures remain.	
Phase II Released	Active mining has ceased; treatment facilities and discharge structures usually have been removed. If landowner desires ponds not be removed, one year of data demonstrates influent quality complies with effluent limitations.	Not considered as an existing source in the TMDLs.
Completely Released	Active mining has ceased; treatment facilities and discharge structures have been removed.	
Revoked	Bond forfeited, NPDES permit may be expired or revoked, highest potential impact to water quality.	Base condition loading determined by model. Considered non-point source loading in the TMDL.

3.3 Non-point Sources

Non-point sources of acidity and metals also contribute to the environmental degradation of the Paint Creek watershed. The largest source of acidity and metals within the watershed consists of AMD from abandoned mine lands (AML). AML sites are those mines for which no company or individual is responsible for the quality or quantity of mine drainage. These mines were closed prior to the passage of the Surface Mining Control and Reclamation Act (SMCRA, 1977). Figure 3-4 is a map of the known non-point sources of AMD within the Paint Creek watershed. Bond forfeiture sites also represent mining non-point sources of acidity and metals. Because the little is known about the acidity and metals loading from non-point sources, the magnitude of these loads must be determined during the model calibration process.

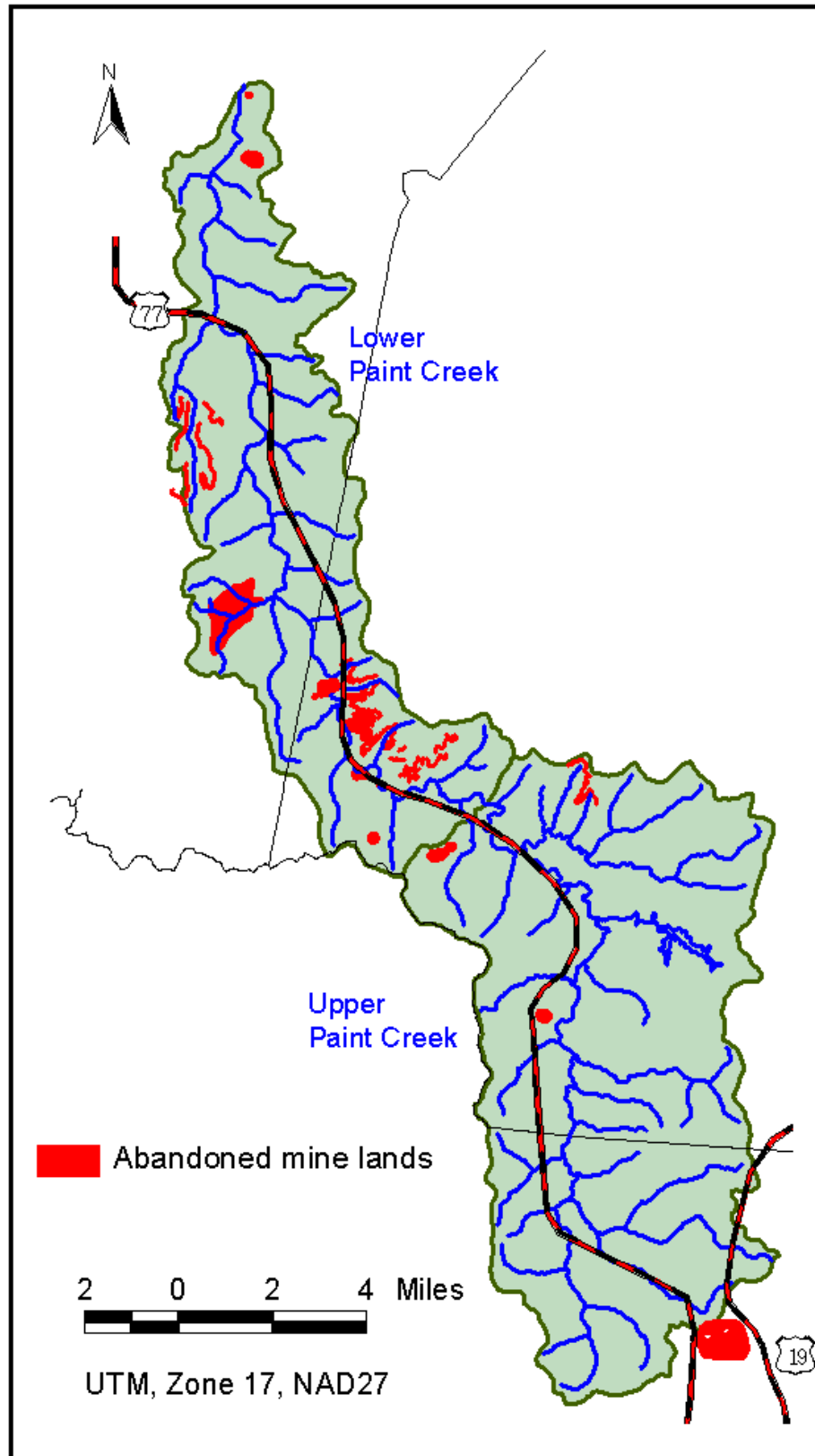


Figure 3-2. Known Mining NPS sites within the Paint Creek Watershed.

4.0 Technical Approach

Establishing the relationship between the in-stream water quality targets and source loadings is a critical component of TMDL development. It allows for evaluation of management options that will achieve the desired source load reductions. The link can be established through a range of techniques, from qualitative assumptions based on sound scientific principles to sophisticated modeling techniques. Ideally, the linkage will be supported by monitoring data that allow the TMDL developer to associate certain water quality responses to flow and loading conditions. The objective of this section is to present the approach taken to develop the linkage between sources and in-stream response for TMDL development in the Paint Creek watershed.

4.1 Model Framework Selection

Selection of the appropriate approach or modeling technique required consideration of the following:

- Expression of water quality criteria
- Dominant processes
- Scale of analysis

The relevant criteria for metals and pH were presented in Section 2. Numeric criteria, such as those applicable here, require evaluation of magnitude, frequency, and duration of appropriate water quality parameters. For metals, the West Virginia criteria are expressed as total metals. This dictates that the methodology predict the total metals concentration in the water column of the receiving water. The criteria for iron and aluminum are expressed as a concentration that cannot be exceeded at a rate greater than the specified exceedance frequency (e.g., not to exceed more than once every 3 years on average). Acute standards (e.g., aluminum, manganese and pH) typically require evaluation over short time periods and violations may occur under variable flow conditions. Chronic criteria (e.g., iron) require the evaluation of the response over a four-day averaging period. The approach or modeling technique must permit representation of in-stream concentrations under a variety of flow conditions, in order to evaluate critical periods for comparison to chronic and acute criteria.

The approach must also consider the dominant processes regarding pollutant loadings and in-stream fate. For the Paint Creek watershed, primary sources contributing to metals and pH impairments include an array of non-permitted (non-point) sources as well as permitted (point) sources. Unlike non-permitted discharges, permitted discharges are controlled by permit limits.

Key in-stream factors that must be considered include advection, dispersion, reaction and loading of simulated constituents. Significant chemical processes include: oxidation, precipitation and sedimentation. In addition to advection and dispersion, significant physical processes include stream reaeration and meteorological heating.

Scale of analysis must also be considered in the selection of the overall approach. The approach should have the capability to evaluate watersheds at multiple scales, particularly those of a few hundred acres in size. The listed waters in the Paint Creek watershed range from small streams to the main stem of the river. Selection of scale should be sensitive to locations of key features, such as abandoned mines and point source discharges. At the larger watershed scale, stream segments are lumped into subwatersheds for practical representation of the system, commensurate with the available data. Occasionally, there are site specific and localized acute problems, which may require more detailed segmentation or definition of detailed modeling grids.

Based on the considerations described above, analysis of the quality and quantity of hydrologic and water quality monitoring data, review of the literature and past pH and metals modeling experience, the source-response linkage in the Paint Creek watershed is represented with the Total Acidic Mine Drainage Loading (TAMDL) computer program. TAMDL was designed to simulate the stream transport, reaction and loading of those water quality constituents related to AMD. The Paint Creek TAMDL model and associated support software form a modeling system capable of representing loading from non-point and point sources found in the watershed and simulating in-stream processes.

4.2 Total Acidic Mine Drainage (TAMDL) Overview

The computer program TAMDL is designed to simulate those aspects of a watershed's stream water quality that are affected by acidic mine drainage. The current version of TAMDL simulates water temperature, net acidity, proton activity (pH), ferrous iron, ferric iron, manganese, aluminum and dissolved oxygen. Water quality conditions are simulated by numerically solving the one dimensional advection, dispersion, loading and reaction partial differential equation for each of these constituents. A detailed description of the theoretical basis of the TAMDL program is presented in Appendix B.

4.3 Model Configuration

The Paint Creek TAMDL model was configured by dividing the watershed into a series of hydrologically connected sub-watersheds and assembling the input data for each of the sub-watersheds. During a model simulation, TAMDL solves the governing partial differential equation for the stream segment in each sub-

watershed. The upstream boundary conditions for the lower sub-watershed models are defined by the results generated for the upper sub-watersheds. Simulated constituents include: aluminum, ferric iron, manganese, net acidity and pH. The key components of the TAMDL model are described in the following sections.

4.3.1 Watershed Subdivision

To represent watershed loadings and the resulting concentrations of acidity and metals, the Paint Creek watershed was divided into 62 sub-watersheds. The stream segments simulated in each of these sub-watersheds are presented in Figure 4-1 and divided by hydrologic boundaries. The division was based on digital elevation data, stream connectivity from EPA's Version 3 Reach File and the locations of WVDEP-SRG water quality sample collection stations. The TAMDL model was calibrated to each of the WVDEP-SRG stations. The division was performed with the assistance of the Watershed Characterization and Modeling System (WCMS) program developed by the National Resources Analysis Center (NRAC) at West Virginia University. The Paint Creek TAMDL model sub-watersheds are shown in Figure 4-2. The stream names corresponding to the TAMDL model sub-watersheds are listed in Table 4-1. The technical capabilities of WCMS are outlined in Fletcher and Strager (2000).

WCMS was developed to bring spatial data and water quality modeling to the desktop of WVDEP personnel and is a customized ArcView GIS interface that combines a wide variety of spatial data layers and water quality modeling components for meeting common WVDEP tasks. Running within ArcView 3x it provides desktop mapping and analysis capabilities for the entire state of West Virginia (Fletcher and Strager, 2000).

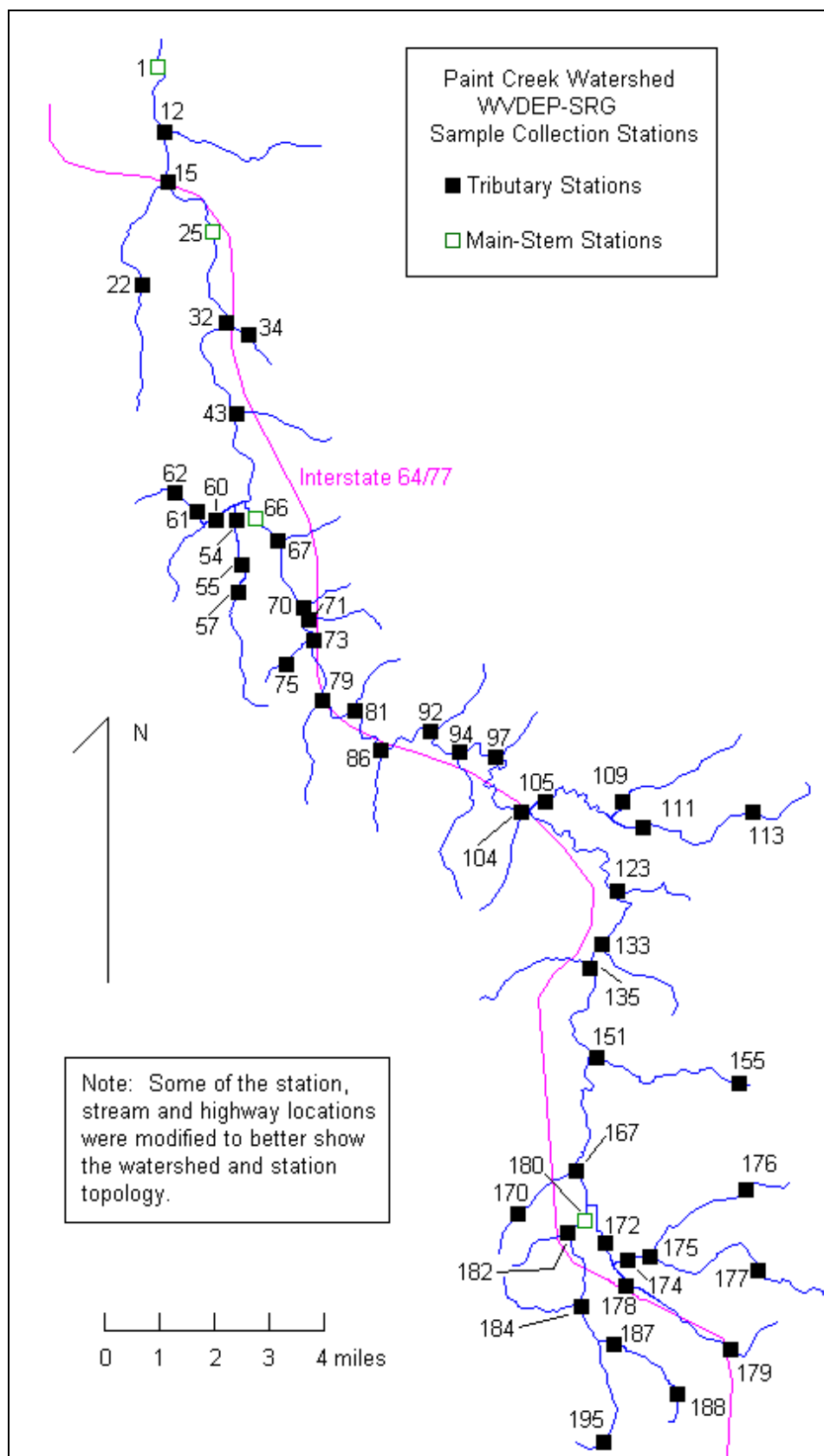


Figure 4-1. Domain of the Sub-models within the Paint Creek TAMD Model.



Figure 4-2. Paint Creek TAMDL Model Sub-watersheds.

Table 4-1. Paint Creek Stream Names and Sub-watershed Numbers.

Stream Name	Sub-watershed Number
Paint Creek below Banner Hollow	1
Banner Hollow	2
Paint Creek above Banner Hollow and below Fourmile Fork	3
Fourmile Fork	4
Paint Creek above Fourmile Fork and below Ash Branch	5
Ash Branch	6
Paint Creek above Ash Branch and below Toms Branch	7
Toms Branch	8
Paint Creek above Toms Branch and below Tenmile Branch	9
Long Branch	10
Tenmile Branch above Long Branch and below Unnamed Tributary	11
Unnamed Tributary of Tenmile Branch	12
Tenmile Branch above Unnamed Tributary	13
Paint Creek above Tenmile Branch and below Laurel Branch	14
Laurel Branch	15
Paint Creek above Laurel Branch and below Unnamed Branch	16
Unnamed Branch of Paint Creek	17
Paint Creek above Unnamed Branch and below Hickory Camp Branch	18
Hickory Camp Branch	19
Paint Creek above Hickory Camp Branch and below Cedar Creek	20
Cedar Creek	21
Paint Creek above Cedar Creek and below Fifteenmile Creek	22
Fifteenmile Creek	23
Paint Creek above Fifteenmile Creek and below Unnamed Tributary	24
Spring Branch	25
Paint Creek above Unnamed Tributary and below Skitter Creek	26
Skitter Creek	27
Paint Creek above Skitter Creek and below Rattlesnake Run	28
Rattlesnake Run	29
Paint Creek above Rattlesnake Run and below Milburn Creek	30
Milburn Creek	31
Paint Creek above Milburn Creek and below Lykins Creek	32
Lykins Creek	33
Paint Creek above Lykins Creek and below Bishop Fork	34
Bishop Fork	35
Paint Creek above Bishop Fork and below Mossy Creek	36
Mossy Creek below Lick Fork	37
Lick Fork	38
Mossy Creek above Lick Fork	39
Paint Creek above Mossy Creek and below Plum Orchard Creek	40
Plum Orchard Creek	41
Paint Creek above Plum Orchard Creek and below Horse Creek	42
Horse Creek	43
Paint Creek above Horse Creek and below Town Creek	44
Town Creek	45
Paint Creek above Town Creek and below Packs Branch	46
Packs Branch	47
Paint Creek above Packs Branch and below Dixons Branch	48
Dixons Branch	49
Paint Creek above Dixons Branch and below Sand Branch	50
Sand Branch below North Sand Branch	51
North Sand Branch below Maple Fork	52

Stream Name	Sub-watershed Number
Maple Fork	53
North Sand Branch above Maple Fork	54
South Sand Branch	55
Paint Creek above Sand Branch and below Laurel Branch	56
Laurel Branch	57
Paint Creek above Laurel Branch and below Davis Branch	58
Davis Branch	59
Paint Creek above Davis Branch and below Lefthand Fork	60
Lefthand Fork	61
Paint Creek above Lefthand Fork	62

4.3.2 Meteorological Data

Meteorological data was not a critical component of the Paint Creek watershed TAMDL model because some sources of AMD release more acid and metals during precipitation events and some do not. However, the oxidation and precipitation reactions simulated by the model have some dependence on temperature and dissolved oxygen concentration, so mean monthly temperature and wind speed data from the National Weather Service station at Charleston, WV was included in the model. While TAMDL does have the capability to simulate stream reaeration due to fluid motion, numerical experiments have indicated that the simulation of reaeration is impractical when the mean depth of the stream is less than one meter. Because the Paint Creek TAMDL model must simulate water quality conditions in a variety of flow conditions, it was decided that the model would assume that the stream was saturated with dissolved oxygen. Given the rugged topography of the watershed, this assumption is reasonable.

4.3.3 Point Source Representation

Point sources in the TAMDL model are represented as an acid and/or metal load being applied to a particular finite difference model node in a particular sub-watershed for a specified period. For matching model results to historical data, which is described in more detail in the Model Calibration section, it was necessary to represent the point sources using available historical data. If Discharge Monitoring Report (DMR) data are available, permitted mines are represented in the model using average flows and pollutant loads. The DMR data includes monthly averages and maximums for flow, pH, aluminum, iron, and manganese. The monthly average metals concentrations were multiplied by the discharge flows to estimate average loadings for these point sources.

4.3.4 Non-point Source Representation

Since the quality of drainage from both surface and deep mines depends upon the nature of the minerals in contact with the water (Skousen, Sexstone and Ziemkiewicz, 2000), observed acidity and metal concentrations will depend upon the flow path of the drainage as it reaches the receiving stream. The geometry of this drainage flow path is essentially a random variable independent of the quantity of mine drainage. This is illustrated by Figure 4-3, which is a scatter plot of reported iron concentrations versus reported discharge rates for permitted mine outlets in the Paint Creek watershed.

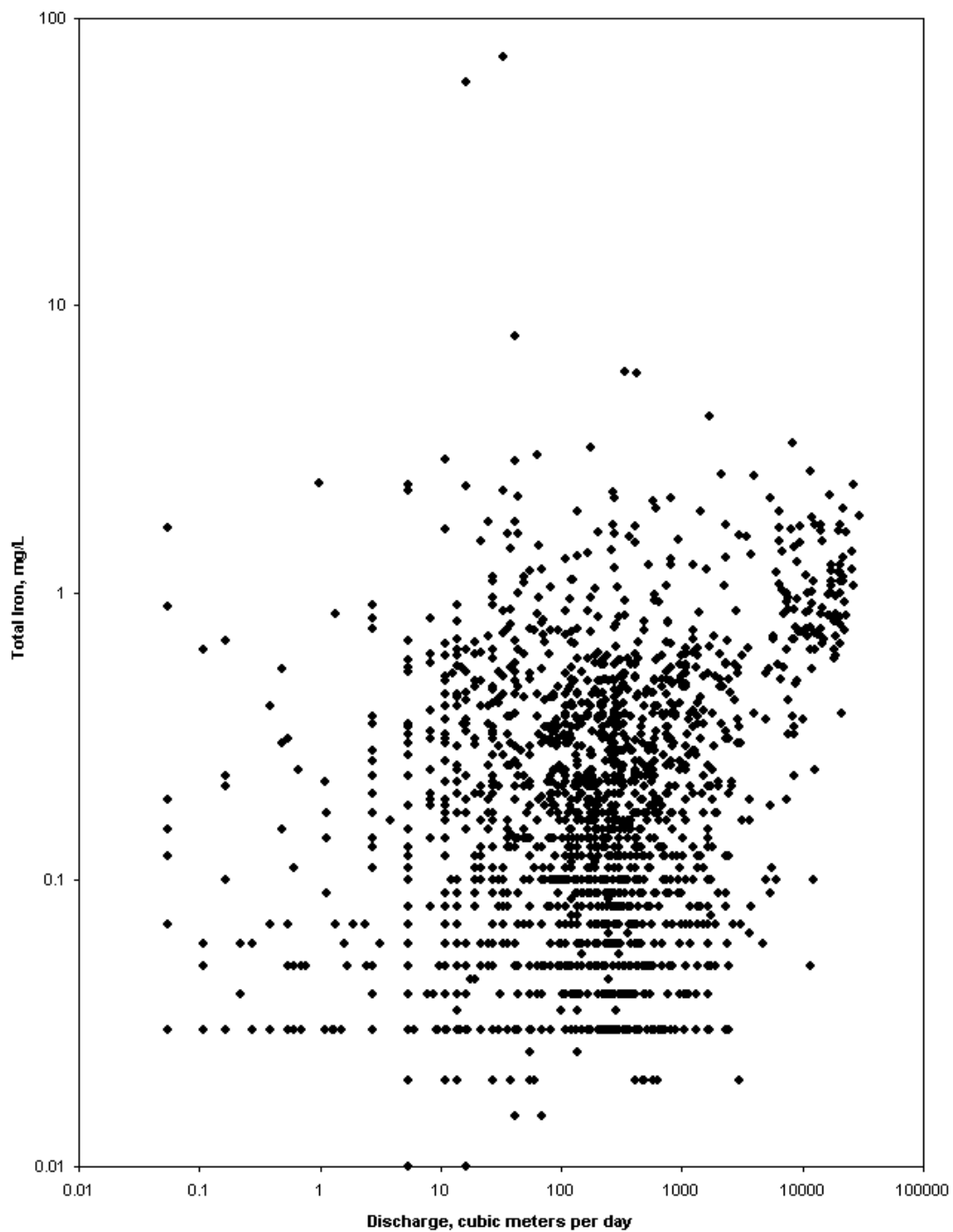


Figure 4-3. Iron Concentration vs. Discharge Rate from Paint Creek DMR data.

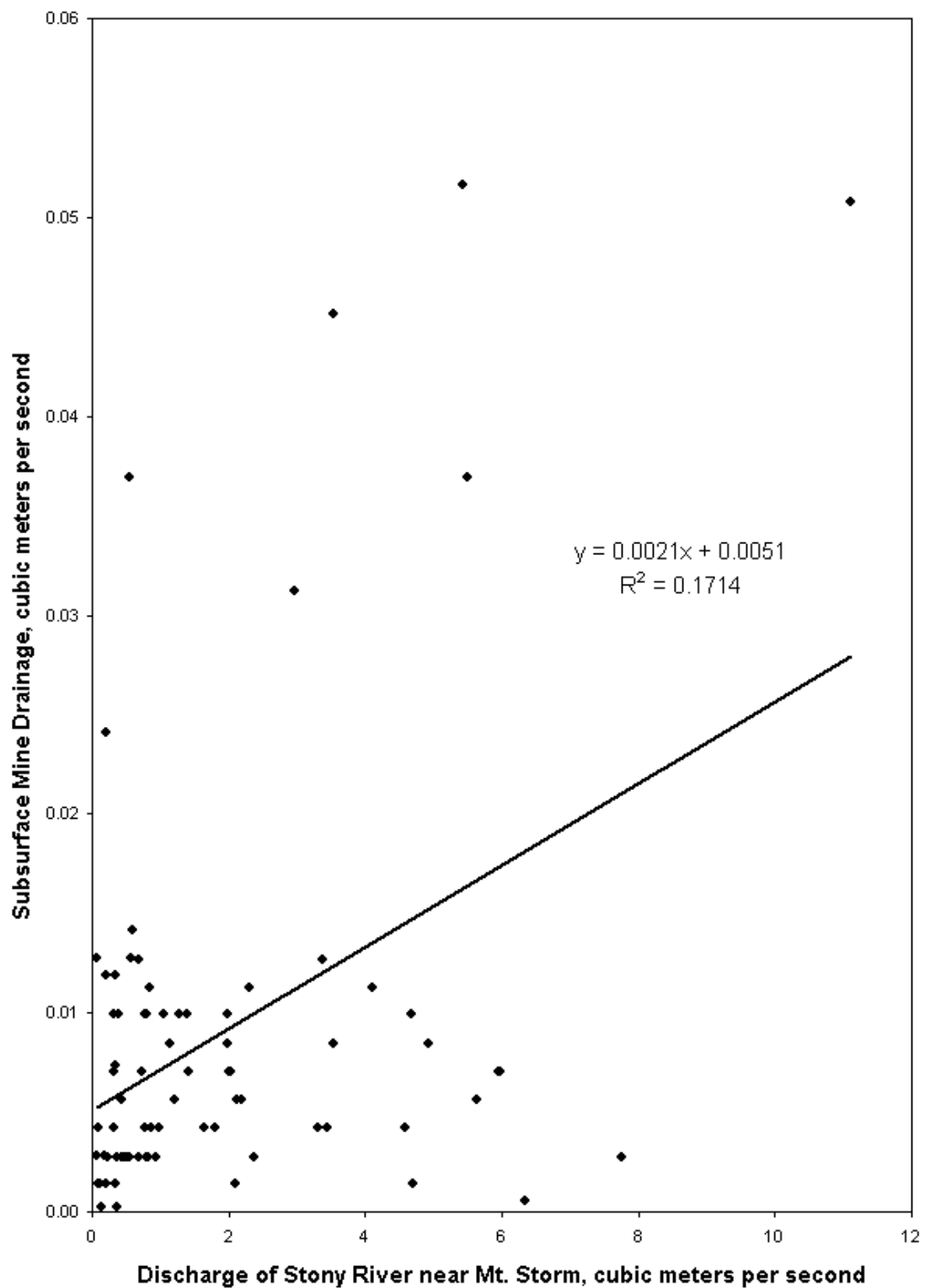


Figure 4-4. Deep Mine Drainage Discharge Rate versus Stony River Stream Discharge Rate near Mt. Storm, WV.

Because of Darcy's law, the lag time between precipitation events and increases in the mine drainage discharge rate is inversely proportional to the length of the flow path and the soil's hydraulic conductivity along the flow path. Therefore, the degree to which the mine drainage discharge rate is driven by individual precipitation events is a function of the geometry of the drainage flow path. This lack of correlation between mine effluent quality and discharge rate was also observed with the DMR concentrations for aluminum and manganese and pH.

Figure 4-4 illustrates the random nature of this drainage flow path for the discharge from a deep mine in Grant County, West Virginia. If the quantity of deep mine drainage was directly proportional to the precipitation rate, then it would be possible to correlate deep mine drainage and stream discharge rates. If the drainage flow path was consistent from one rainfall event to another, then it would have been possible to achieve a correlation between mine drainage and weekly or monthly averaged stream discharge rates. The extremely low correlation coefficient observed with the daily average stream data, $R^2 = 0.1714$, was actually greater than the correlation observed with the weekly and monthly averaged data, 0.1509 and 0.1405, respectively. In the absence of similar data for any of the deep mines in the Paint Creek watershed, this report will assume that the quantity of deep mine drainage is not directly proportional to either the short-term or long-term precipitation rate.

While the quantity of drainage from surface mines may, in some instances, be directly proportional to the precipitation rate, distinguishing between abandoned surface and deep mined areas on existing GIS coverages is difficult. Therefore, non-point sources are represented in the Paint Creek TAMDL model in the same manner as point sources. Unlike point sources, non-point sources of mine drainage in the Paint Creek watershed are not regularly sampled by either regulatory agencies or mine operators. With the exception of a limited amount of data for the Ten Mile Fork, Long Branch and Cedar Creek sub-watersheds, water quality data from AML portals was not available. Consequently, the determination of the magnitude of acid and metal loads from these sources was made during the calibration of the model.

4.3.5 Stream Representation

Modeling subwatersheds and calibrating hydrologic and water quality model components required simulating the transport of both water and pollutants through streams. Each subwatershed was represented with a single stream with uniform geometric and hydraulic characteristics. Stream segments were identified using EPA's RF3 stream coverage within the program WCMS. TAMDL simulates stream flow by assuming uniform flow conditions. Required stream data includes slope, Manning's roughness coefficient, and mean channel dimensions. Stream slopes and lengths were calculated from the DEM data and

the RF3 stream coverage available in WCMS. Manning's roughness coefficient and stream dimensions were estimated from observed data.

4.3.6 Hydrologic Representation

Hydrologic processes were simulated in the model by distributing the daily stream flow measured at a local gauging station according to sub-watershed drainage area. The flow hydrograph for each tributary sub-watershed was calculated by multiplying the daily flow observed at the station by the ratio of the sub-watershed drainage area and the gauging station drainage area. This method is more accurate when the gauging station and the tributary sub-watershed have similar drainage areas and runoff characteristics. The flow hydrographs for downstream sub-watersheds was calculated by summing the flow hydrographs of the adjacent upstream sub-watersheds. Because the model employed daily flow hydrographs and the small size of the sub-watersheds, hydrologic routing techniques were not employed.

Unfortunately, no USGS stations were located within the Paint Creek watershed; the closest USGS station was at site number 03200500 on the Coal River at Tornado, WV. The USGS Coal River station at Tornado, WV has a drainage area of 1260 km² (862 mi²) and a period of record extending back to July 1, 1908. The Coal River station was selected because of its relative proximity to the watershed (05050009 Coal) and its rather long period of record.

Since there was very little hydrologic data directly measured within the watershed to calibrate model hydrology, it was decided that the application of a more sophisticated hydrologic model would not improve the ability of the model to simulate water quality conditions.

4.3.7 Pollutant Representation

In addition to flow, the Paint Creek watershed TAMDLC model simulated four water quality constituents: pH, net acidity, iron, manganese, and aluminum. Stream pH was calculated by TAMDLC with an empirical constitutive relationship from the net acidity of the stream. The loading contributions of net acidity, iron, manganese, and aluminum from different non-point and point sources were represented in the TAMDLC model as sources applied to individual finite difference nodes. Discharge data from the DMRs for individual NPDES permits were used to calculate the point source contribution of these constituents. The contributions of these constituents from non-point sources were estimated as part of the model calibration process.

4.4 Model Calibration

After the model was configured, calibration was performed at multiple locations throughout the Paint Creek watershed. Calibration refers to the adjustment or fine-tuning of modeling parameters to reproduce observations. Because no USGS gauging stations are located within the watershed, model calibration focused on water quality.

Observed water quality data were obtained from EPA's STORET database as well as from the Stream Restoration Group (WVDEP-SRG). Data from both sources were obtained through WVDEP. Normally significant amounts of time-varying monitoring data are necessary to calibrate any water quality model. A total of 210 stream water quality data points was available for model calibration and are listed in Appendix C. Water quality data from samples collected from abandoned mine seeps were used to estimate some of the mining non-point source loads and are listed in Appendix C. Included in Appendix C are the results of a violation analysis performed on the observed stream water quality data.

Modeled versus observed in-stream concentrations were directly compared during model calibration. The water quality calibration consisted of executing the watershed model, comparing water quality time series output to available water quality observation data, and adjusting water quality parameters within a reasonable range. Results of the water quality calibration are presented in Appendix D.

5.0 Allocation Analysis

A TMDL is the total amount of a pollutant that can be assimilated by the receiving water while still achieving water quality standards. TMDLs can be expressed in terms of mass per time or by other appropriate measures. TMDLs are comprised of the sum of individual wasteload allocations (WLA) point sources, load allocations (LA) for non-point sources, and natural background levels. In addition, the TMDL must include a margin of safety (MOS), either implicitly or explicitly, that accounts for the uncertainty in the relationship between pollutant loads and the quality of the receiving stream. Conceptually, this definition is denoted by the following equation.

$$\text{TMDL} = \text{WLA} + \text{LA} + \text{MOS}$$

In order to develop aluminum, iron, manganese, and pH TMDLs for each of the streams in the Paint Creek watershed, the following approach was taken:

1. Define TMDL endpoints.
2. Simulate baseline conditions.

3. Assess source loading alternatives.
4. Determine the TMDL and source allocations.

5.1 TMDL Endpoints

TMDL endpoints represent the in-stream water quality targets used in quantifying TMDLs and their individual components. Different endpoints are necessary for each impairment type (i.e., aluminum, iron, manganese, and pH). The State of West Virginia's numeric water quality criteria for aluminum, iron, manganese, and pH and an implicit margin of safety were used to identify endpoints for TMDL development.

The TMDL endpoint for aluminum was selected as 0.7125 mg/L (based on the 0.75 mg/L criteria for aquatic life minus a 5% MOS). The iron endpoint was selected either as 0.475 mg/L (based on the 0.5 mg/L criteria for aquatic life-trout waters minus a 5% MOS) or 1.425 mg/L (based on the 1.5 mg/L criteria for aquatic life minus a 5% MOS). The manganese endpoint was selected as 0.95 mg/L (based on the 1.0 mg/L criteria for human health minus a 5% MOS). The water quality criterion for pH requires it to be between 6.0 and 9.0 standard units (SU), inclusive. Instead of using a percentage-based margin of safety for pH, this study used a MOS of 0.5 SU. Therefore, the TMDL endpoints for pH are 6.5 SU and 8.5 SU. The magnitudes of these margins of safety are based upon the expected reliability of the TMDL determination, in this case modeling, effort.

In the case of acid mine drainage, pH, is not a good indicator of stream acidity and can be a misleading characteristic. Water with a circum-neutral pH (~7) but containing elevated concentrations of dissolved ferrous ions can become acidic after oxidation of the ferrous iron. Therefore, a more practical approach to meeting the water standards of pH is to use the net acidity of the stream as a surrogate for pH. The net acidity of a solution is the solution's total acidity minus the total alkalinity and is related to the stream's pH in TAMDL with an empirical constitutive relationship. TAMDL expresses a stream net acidity in terms of the concentration of calcium carbonate required to neutralize the stream (i.e., mg of calcium carbonate equivalents per liter of water). The procedure by which TAMDL uses net acidity as a surrogate for pH is explained more completely in the documentation of the TAMDL computer program in Appendix B.

Components of the TMDLs for aluminum, iron, and manganese are presented in terms of mass per unit time in this report. For pH, this report presents the TMDL in terms of net acidity loading. When the net acidity load is positive, it is defined as the mass of calcium carbonate needed to neutralize the acidity per unit time. When it is negative, it is defined as the alkalinity equivalent to the mass of calcium carbonate per unit time multiplied by negative one.

5.2 Baseline Conditions

The calibrated model provided the basis for performing the allocation analysis. The first step in this analysis involved simulation of baseline conditions. Baseline conditions represent existing non-point source loading conditions and permitted point source discharge conditions. The baseline conditions allow for an evaluation of in-stream water quality under the “worst currently allowable” scenario.

The model was run for baseline conditions for the period October 1, 1992 through September 30, 1999. Predicted in-stream pH, instantaneous concentrations of aluminum and manganese, and four-day averaged iron concentration for the streams in the Paint Creek watershed were compared directly to the TMDL endpoints. This comparison allowed evaluation of the expected magnitude and frequency of exceedances under a range of hydrologic and environmental conditions, including dry periods, wet periods, and average periods. This simulation period also bracketed the observed water quality and hydrologic data set.

The baseline loads from permitted mine outlets were estimated by multiplying the concentrations presented in Table 5-1 by the estimated discharge flow rate. For iron and manganese, the concentrations present in Table 5-1 represent the wasteload allocations associated with existing permit limits. They were determined through a back-calculation procedure using the limitation development principles of EPA’s *Technical Support Document for Water Quality Based Toxics Control*. Although technology-based WV/NPDES Permits do not contain effluent aluminum concentration limits, more than 99.5% of observed effluent aluminum concentrations are less than value listed in Table 5-1. As such, the concentration is a reasonable representation of the aluminum wasteload allocation that is associated with existing permits. Figure 5-1 is a plot of the reported effluent aluminum concentration from Paint Creek DMRs during the simulation period of the baseline conditions model.

Table 5-1. Metals concentrations used in Representing Mine Discharge Loads.

Pollutant	Technology-based Permit Limits	Water Quality-based Permit Limits
Aluminum	4.3 mg/L*	0.75 mg/L
Iron	3.2 mg/L	1.5 mg/L, 0.5 mg/L (trout waters)
Manganese	2.0 mg/L	1.0 mg/L

*WVDEP technology-based mining permits require only the reporting of aluminum concentration.

The discharge flow rates from permitted mining outlets were estimated using one of two methods. If the outlet had observed DMR discharge data, then that data was used to calculate the average discharge rate for each calendar month of the simulation period. If the outlet had no DMR data for a particular month, then the annual average discharge was assigned to that month’s discharge rate.

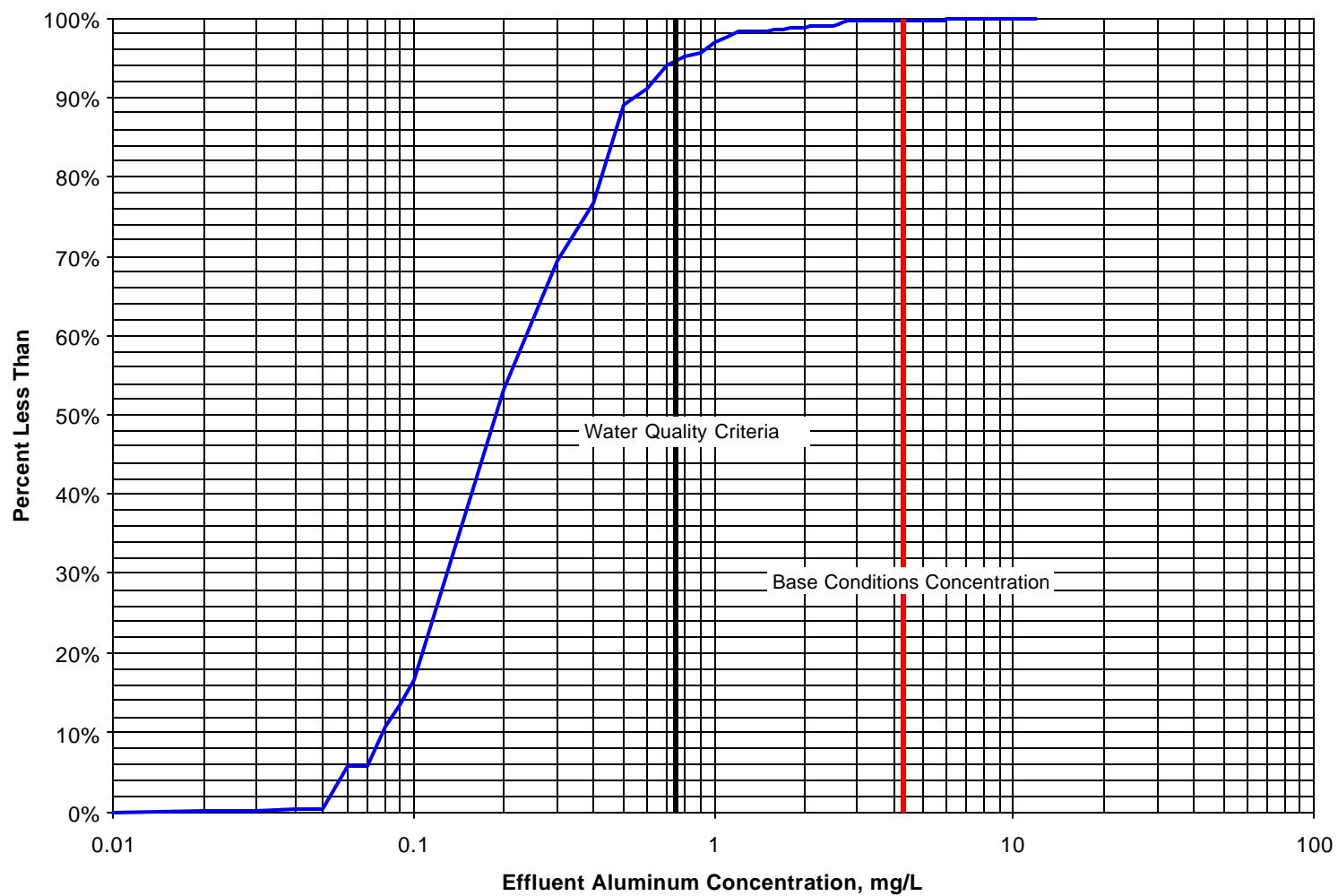


Figure 5-1. Cumulative Probability Distribution of Reported Effluent Aluminum Concentrations in Paint Creek Watershed.

If DMR discharge data were not available, then WCMS program was used to estimate the disturbed area of the mine outlet and the surface drainage rate for a stream near the mine outlet. Because the surface drainage rate (mean stream flow rate per drainage area) does vary by season, the average surface drainage rate for each calendar month was obtained from WCMS. Hydrologists familiar with the drainage rates from mines in West Virginia estimate that the disturbed area of a mining operation will discharge more water as an undisturbed area in a watershed. The monthly average mine discharge rate for the outlet was estimated by multiplying the product of the disturbed area of the outlet and the monthly average surface drainage rate for the watershed by a factor of 2.

While this latter method of estimating the monthly average discharge from a mine outlet is subject to errors introduced by the estimate of disturbed area and drainage rate, a more precise analysis would involve a detailed examination of the permit application for each mine in the watershed and was deemed to be impractical for this study. Estimates of mine outlet discharge are not normally required in mining permit applications.

5.3 Source Loading Alternatives

Simulation of baseline conditions provided the basis for evaluating each stream's response to variations in source contributions under virtually all conditions. This sensitivity analysis gave insight into the dominant sources and how potential decreases in loads would affect in-stream pH and metals concentrations. For example, loading contributions from abandoned mines, permitted facilities, and other non-point sources were individually adjusted and in-stream concentrations were observed.

Multiple scenarios were run for the impaired streams. Successful scenarios were those that achieved the TMDL endpoints under all conditions for pH, aluminum, iron, and manganese (through comparison of model results for the entire seven-year simulation period). A scenario would be judged unsuccessful, if exceedances in any of these parameters occurred more frequently than is permitted by the water quality standards. Model output was sampled in approximately one-day intervals for these assessments. In general, loads contributed by mines with revoked permits and abandoned mines were reduced first, because they generally had the greatest impact on in-stream water quality. If additional load reductions were required to meet the TMDL endpoints, then reductions were made in point source (permitted) contributions.

The general allocation philosophy used in this TMDL is further described as follows:

1. Pollutant reductions were not required of non-mining point or non-point sources. Non-mining point sources in this watershed do not discharge significant amounts of acidity, aluminum, iron and manganese. Unlike some watersheds in West Virginia, stream sediments in the Paint Creek watershed do not introduce significant amounts of aluminum to stream waters. While the research presented by Watts, Hinkle and Griffitts (1994) indicate that sediments in the main stem of Paint Creek contain up to 6.22% aluminum, no correlation between observed aluminum concentration and stream flow or total suspended solids concentration was found. In the absence of other sources, the pollutants contributed by non-mining non-point sources (forest, agriculture, urban, sediments) do not cause water quality criteria violation.
2. Pollutant reductions of mining non-point sources (revoked permits and AML sites) were required first, to the extent necessary to achieve in-stream compliance.
3. Pollutant reductions from mining point sources were required only if mining non-point sources are not present in the subwatershed, or if the reduction of existing mining non-point sources was inadequate to achieve in-stream compliance.

This methodology ensures water quality criteria compliance in all streams in the watershed, targets pollutant reductions from the primary causative sources of impairment, and minimizes the impact to existing point sources in the watershed.

For most of the Paint Creek sub-watersheds, it was possible to bring the stream into compliance with the water quality standards by making reasonable reductions in the pollutant loads from mining non-point sources or by making reductions in the effluent concentration limits from certain mining point sources. A reasonably achievable reduction of mining non-point sources is defined as being the reduction in loading required to bring metal concentrations down to a level that would be present if mining non-point sources were releasing water in accordance with technology-based limits.

A similar criteria for the reasonable reduction of the acid load from non-point sources was not adopted because WVDEP requires that the effluent from outlets with either technology-based or water quality limits be circum-neutral ($6.0 \leq \text{pH} \leq 9.0$). Because the lower limit of this range corresponds to a such a small level of net acidity, the uncertainty in calculating the reasonable acid load minimum exceed the magnitude of the acid load. Therefore this TMDL assumes that a complete reduction of the acid load from an AML site to be reasonable. While WVDEP does not restrict the effluent aluminum concentration from outlets with a technology-based limit, the reported aluminum effluent concentrations for mine outlets in the Paint Creek watershed, as shown in Figure 5-1, indicate that the water quality standard, 0.75 mg/L, is a reasonably achievable reduction in

concentration. The required reductions in load from mining non-point sources for each sub-watershed are given in Appendix F.

5.4 TMDLs and Source Allocations

A top-down methodology was followed to develop the TMDLs and allocate loads to sources. Headwaters and tributaries were analyzed first, because their impact frequently had a profound effect on downstream water quality. In impaired subwatersheds, loading contributions were reduced to the extent necessary to ensure compliance with in-stream criteria, and the loading associated with that condition was transferred to downstream subwatersheds. Conversely, where the model indicated that the baseline condition was compliant with water quality criteria, the loading associated with the baseline condition was transferred to downstream subwatersheds. The required headwater reductions often led to downstream water quality improvements, effectively decreasing necessary loading reductions from downstream sources.

In some situations, reductions in sources contributing to stream segments not included on the 303(d) list have been determined necessary to ensure universal compliance with water quality criteria in the watershed. The listed and non-listed stream segments requiring reduction in acid and metal loads is presented in Table C3 of Appendix C. Recent water quality data is not available for all streams in the watershed and the model is a technical tool available to determine if a particular permit is protective of water quality criteria. Other situations have been encountered where recent water quality data indicates that a particular stream segment is not impaired, yet the TMDL imposes point source wasteload allocations that represent a reduction of existing permit limitations. Certain permit holders are currently achieving discharge quality better than what is required by their permit may need to maintain such improved performance in order for the receiving water to consistently meet standards. The Paint Creek TMDLs for pH, aluminum, iron and manganese are listed in Tables 5-2, 5-3, 5-4 and 5-5, respectively. Sub-watersheds containing streams that have been placed on the 303(d) list are shown in bold type. Therefore, the rows in Table 5-2 which are shown in bold type correspond to stream segments which have been placed on the 303(d) list for pH, and the rows in Table 5-3 which are shown in bold type represent stream segments which are listed for aluminum. Tables 5-4 and 5-5 have bold rows for those stream segments listed for iron and manganese, respectively.

5.4.1 Wasteload Allocations

The wasteload allocations for aluminum, iron, and manganese are expressed in terms of a concentration within the ranges listed in Table 5-6. No wasteload allocation was necessary for acidity because the allowable technology-based

limits for pH are the same as the water quality limits. The minimum range in Table 5-6 reflects the in-stream water quality criteria, and the maximum was derived from existing technology-based permits using the procedures of the EPA's *Technical Support Document for Water Quality-based Toxics Control* (USEPA, 1991). The wasteload allocations for all NPDES outlets in the Paint Creek watershed are listed in Appendix E.

Table 5-2. pH TMDL for each of the Paint Creek Sub-Watersheds.

SWS	WLA, Mg/yr CaCO ₃ equivalents	LA, Mg/yr CaCO ₃ equivalents	Upstream Contribution, Mg/yr CaCO ₃ equivalents	TMDL, Mg/yr CaCO ₃ equivalents	Baseline NPS Load, Mg/yr CaCO ₃ equivalents	Relative NPS Load Reduction
1	0.0000	0.0000	16.1713	16.1713	0.0000	0.0000%
2	0.0000	-26.8459	0.0000	-26.8459	-26.8459	0.0000%
3	0.0000	0.0000	43.0172	43.0172	0.0000	0.0000%
4	0.0000	-16.9228	0.0000	-16.9228	-16.9135	0.0550%
5	0.0000	0.0000	59.9400	59.9400	0.0000	0.0000%
6	0.0000	-65.7450	0.0000	-65.7450	-65.7423	0.0041%
7	0.0000	0.0000	125.6850	125.6850	0.0000	0.0000%
8	0.0028	0.0000	0.0000	0.0028	0.0000	0.0000%
9	0.0329	0.0000	125.6493	125.6822	0.0000	0.0000%
10	0.0246	0.0547	0.0000	0.0793	52.0795	99.8949%
11	0.0019	0.0000	-40.3017	-40.2998	71.0703	100.0000%
12	0.0000	-19.5621	0.0000	-19.5621	-19.5621	0.0000%
13	0.0000	-20.7396	0.0000	-20.7396	-20.7396	0.0000%
14	0.0000	0.0000	165.8698	165.8698	0.0000	0.0000%
15	0.0834	591.6960	0.0000	591.7794	1829.8900	67.6649%
16	0.0000	0.0000	-425.9096	-425.9096	0.0000	0.0000%
17	0.0000	-0.7872	0.0000	-0.7872	-0.7872	0.0001%
18	0.0000	0.0000	-425.1224	-425.1224	0.0000	0.0000%
19	0.0000	-1.4399	0.0000	-1.4399	-1.4327	0.5018%
20	0.0000	0.0000	-423.6825	-423.6825	0.0000	0.0000%
21	0.0000	0.6322	0.0000	0.6322	6.8210	90.7313%
22	0.0000	0.0000	-424.3147	-424.3147	116.3010	100.0000%
23	0.0086	-10.1211	0.0000	-10.1125	-10.1211	0.0000%
24	0.0000	0.0000	-414.2022	-414.2022	0.0000	0.0000%
25	0.0000	0.1004	0.0000	0.1004	7.3589	98.6359%
26	0.0008	0.0000	-414.3034	-414.3026	0.0000	0.0000%
27	0.0577	-7.1882	0.0000	-7.1305	-7.1882	0.0004%
28	0.0196	0.0000	-407.1925	-407.1729	0.0000	0.0000%
29	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000%
30	0.0000	0.0000	-407.1925	-407.1925	0.0000	0.0000%
31	0.4473	-0.1830	0.0000	0.2643	-0.1830	0.0000%
32	0.0000	0.0000	-407.4567	-407.4567	0.0000	0.0000%
33	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000%

SWS	WLA, Mg/yr CaCO ₃ equivalents	LA, Mg/yr CaCO ₃ equivalents	Upstream Contribution, Mg/yr CaCO ₃ equivalents	TMDL, Mg/yr CaCO ₃ equivalents	Baseline NPS Load, Mg/yr CaCO ₃ equivalents	Relative NPS Load Reduction
34	0.0000	0.0000	-407.4567	-407.4567	0.0000	0.0000%
35	0.0183	-12.0387	0.0000	-12.0204	-12.0386	0.0008%
36	0.0000	0.0000	-395.4363	-395.4363	0.0000	0.0000%
37	0.0000	0.0000	-37.9555	-37.9555	0.0000	0.0000%
38	0.0000	-6.4823	0.0000	-6.4823	-6.4823	0.0000%
39	0.0000	-31.4732	0.0000	-31.4732	-31.4732	0.0000%
40	0.1625	0.0000	-357.6433	-357.4808	0.0000	0.0000%
41	0.0000	-22.7147	0.0000	-22.7147	-22.7147	0.0000%
42	0.0473	0.0000	-334.9759	-334.9286	0.0000	0.0000%
43	0.0000	-7.8421	0.0000	-7.8421	-7.8379	0.0535%
44	0.0000	0.0000	-327.1339	-327.1339	0.0000	0.0000%
45	0.2113	-5.0554	0.0000	-4.8442	-5.0554	0.0004%
46	0.0065	0.0000	-322.2962	-322.2897	0.0000	0.0000%
47	0.0000	-21.5686	0.0000	-21.5686	-21.5686	0.0000%
48	0.0010	0.0000	-300.7286	-300.7276	0.0000	0.0000%
49	0.0000	-12.0616	0.0000	-12.0616	-12.0616	0.0000%
50	0.0000	0.0000	-288.6670	-288.6670	0.0000	0.0000%
51	0.0000	0.0000	-121.5311	-121.5311	0.0000	0.0000%
52	0.0962	0.0000	-80.3733	-80.2771	0.0000	0.0000%
53	0.0000	-21.3956	0.0000	-21.3956	-21.3956	0.0000%
54	0.0000	-58.9777	0.0000	-58.9777	-58.9777	0.0000%
55	0.0000	-41.2540	0.0000	-41.2540	-41.2540	0.0000%
56	0.0000	0.0000	-167.1359	-167.1359	0.0000	0.0000%
57	0.0000	-3.2137	0.0000	-3.2137	-2.9969	7.2337%
58	0.0000	0.0000	-163.9221	-163.9221	0.0000	0.0000%
59	0.0000	-0.4560	0.0000	-0.4560	-0.4560	0.0000%
60	0.0000	0.0000	-163.4661	-163.4661	0.0000	0.0000%
61	0.0000	-14.9771	0.0000	-14.9771	-14.9771	0.0000%
62	0.0000	-148.4890	0.0000	-148.4890	-148.4890	0.0000%

Table 5-3. Aluminum TMDL for each of the Paint Creek Sub-Watersheds.

SWS	WLA, Mg/yr	LA, Mg/yr	Upstream Contribution, Mg/yr	TMDL, Mg/yr
1	0.0000	0.5657	19.9043	20.4700
2	0.0000	0.0457	0.0000	0.0457
3	0.0000	0.0012	19.8574	19.8586
4	0.0000	0.0106	0.0000	0.0106
5	0.0000	2.2317	17.6150	19.8467
6	0.0000	0.0063	0.0000	0.0063
7	0.0000	0.2631	17.3457	17.6088
8	0.0163	0.0164	0.0000	0.0327
9	1.0099	0.3911	15.9120	17.3130
10	0.4755	0.0079	0.0000	0.4834
11	0.0598	0.0117	0.0469	0.1184
12	0.0000	0.0165	0.0000	0.0165
13	0.0000	0.0304	0.0000	0.0304
14	0.0000	0.0012	15.3090	15.3102
15	0.4825	0.0133	0.0000	0.4957
16	0.0000	0.0027	14.8106	14.8132
17	0.0000	0.0029	0.0000	0.0029
18	0.0000	0.0001	14.8075	14.8076
19	0.0000	0.0044	0.0000	0.0044
20	0.0000	0.1293	14.6738	14.8031
21	0.0000	0.0135	0.0000	0.0135
22	0.0000	0.9552	13.7050	14.6603
23	0.1100	0.0039	0.0000	0.1140
24	0.0000	0.1362	13.4549	13.5911
25	0.0000	0.0112	0.0000	0.0112
26	0.0254	0.0023	13.4159	13.4436
27	0.3092	0.0101	0.0000	0.3193
28	0.6008	0.0042	12.4917	13.0966
29	0.0000	0.0029	0.0000	0.0029
30	0.0000	0.0017	12.4870	12.4887
31	2.8754	0.0091	0.0000	2.8845
32	0.0000	0.0012	9.6013	9.6025
33	0.0000	0.0021	0.0000	0.0021
34	0.0000	0.0041	9.5952	9.5992
35	0.2678	0.0041	0.0000	0.2719

SWS	WLA, Mg/yr	LA, Mg/yr	Upstream Contribution, Mg/yr	TMDL, Mg/yr
36	0.0000	0.0001	9.3232	9.3233
37	0.0000	0.0146	0.0237	0.0383
38	0.0000	0.0092	0.0000	0.0092
39	0.0000	0.0145	0.0000	0.0145
40	4.9906	0.0068	4.2875	9.2849
41	0.0000	0.1739	0.0000	0.1739
42	1.4529	0.0031	2.6575	4.1136
43	0.0000	0.0066	0.0000	0.0066
44	0.0000	0.0004	2.6505	2.6509
45	1.1922	0.0055	0.0000	1.1977
46	0.2002	0.2679	0.9847	1.4529
47	0.0000	0.0603	0.0000	0.0603
48	0.0299	0.0098	0.8846	0.9244
49	0.0000	0.0067	0.0000	0.0067
50	0.0000	0.0032	0.8747	0.8779
51	0.0000	0.0017	0.8466	0.8483
52	0.7899	0.0007	0.0210	0.8116
53	0.0000	0.0093	0.0000	0.0093
54	0.0000	0.0117	0.0000	0.0117
55	0.0000	0.0351	0.0000	0.0351
56	0.0000	0.0006	0.0258	0.0264
57	0.0000	0.0040	0.0000	0.0040
58	0.0000	0.0019	0.0199	0.0218
59	0.0000	0.0055	0.0000	0.0055
60	0.0000	0.0023	0.0121	0.0144
61	0.0000	0.0048	0.0000	0.0048
62	0.0000	0.0073	0.0000	0.0073

Table 5-4. Iron TMDL for each of the Paint Creek Sub-Watersheds.

SWS	WLA, Mg/yr	LA, Mg/yr	Upstream Contribution, Mg/yr	TMDL, Mg/yr
1	0.0000	0.4210	20.8020	21.2230
2	0.0000	0.0016	0.0000	0.0016
3	0.0000	0.0003	20.8001	20.8004
4	0.0000	0.0022	0.0000	0.0022
5	0.0000	1.4644	19.3336	20.7979
6	0.0000	0.0013	0.0000	0.0013
7	0.0000	0.0370	19.2953	19.3323
8	0.0637	0.0008	0.0000	0.0646
9	0.7515	0.0517	18.4274	19.2307
10	0.5615	0.3620	0.0000	0.9235
11	0.0445	0.0168	0.1514	0.2127
12	0.0000	0.0694	0.0000	0.0694
13	0.0000	0.0820	0.0000	0.0820
14	0.0000	0.0052	17.2860	17.2912
15	1.1912	0.0005	0.0000	1.1918
16	0.0000	0.0112	16.0830	16.0943
17	0.0000	0.0002	0.0000	0.0002
18	0.0000	0.0006	16.0822	16.0828
19	0.0000	0.0084	0.0000	0.0084
20	0.0000	0.0039	16.0699	16.0738
21	0.0000	0.0307	0.0000	0.0307
22	0.0000	0.3409	15.6983	16.0392
23	0.1956	0.0008	0.0000	0.1964
24	0.0000	0.0062	15.4957	15.5019
25	0.0000	0.0082	0.0000	0.0082
26	0.0189	0.0095	15.4591	15.4875
27	0.7544	0.0019	0.0000	0.7563
28	0.4471	0.0176	14.2380	14.7028
29	0.0000	0.0006	0.0000	0.0006
30	0.0000	0.0072	14.2303	14.2374
31	5.8786	0.0360	0.0000	5.9146
32	0.0000	0.0049	8.3108	8.3157
33	0.0000	0.0004	0.0000	0.0004
34	0.0000	0.0170	8.2933	8.3103
35	0.4180	0.0673	0.0000	0.4853

SWS	WLA, Mg/yr	LA, Mg/yr	Upstream Contribution, Mg/yr	TMDL, Mg/yr
36	0.0000	0.0004	7.8076	7.8080
37	0.0000	0.0307	0.0312	0.0619
38	0.0000	0.0162	0.0000	0.0162
39	0.0000	0.0150	0.0000	0.0150
40	1.7409	0.0288	5.9760	7.7457
41	0.0000	0.0097	0.0000	0.0097
42	1.0812	0.0013	4.8838	5.9663
43	0.0000	0.0070	0.0000	0.0070
44	0.0000	0.0016	4.8752	4.8767
45	2.7013	0.0023	0.0000	2.7035
46	0.1490	0.0023	2.0203	2.1717
47	0.0000	0.1227	0.0000	0.1227
48	0.0223	0.0020	1.8734	1.8977
49	0.0000	0.0066	0.0000	0.0066
50	0.0000	0.0094	1.8574	1.8668
51	0.0000	0.0003	1.7870	1.7873
52	1.6622	0.0001	0.0078	1.6701
53	0.0000	0.0019	0.0000	0.0019
54	0.0000	0.0059	0.0000	0.0059
55	0.0000	0.1169	0.0000	0.1169
56	0.0000	0.0010	0.0691	0.0701
57	0.0000	0.0080	0.0000	0.0080
58	0.0000	0.0033	0.0577	0.0610
59	0.0000	0.0548	0.0000	0.0548
60	0.0000	0.0005	0.0025	0.0029
61	0.0000	0.0010	0.0000	0.0010
62	0.0000	0.0015	0.0000	0.0015

Table 5-5. Manganese TMDL for each of the Paint Creek Sub-Watersheds.

SWS	WLA, Mg/yr	LA, Mg/yr	Upstream Contribution, Mg/yr	TMDL, Mg/yr
1	0.0000	2.631E-01	17.8511	18.1142
2	0.0000	9.223E-06	0.0000	0.0000
3	0.0000	1.191E-06	17.8511	17.8511
4	0.0000	2.471E-04	0.0000	0.0002
5	0.0000	1.312E+00	16.5390	17.8508
6	0.0000	1.863E-02	0.0000	0.0186
7	0.0000	4.109E-03	16.5163	16.5204
8	0.0398	2.748E-03	0.0000	0.0426
9	0.4697	3.806E+00	12.1980	16.4737
10	0.3509	1.901E-02	0.0000	0.3699
11	0.0278	1.059E-02	0.1244	0.1628
12	0.0000	4.534E-02	0.0000	0.0453
13	0.0000	7.901E-02	0.0000	0.0790
14	0.0000	1.186E-06	11.6653	11.6653
15	0.6373	6.466E-04	0.0000	0.6380
16	0.0000	2.555E-06	11.0273	11.0273
17	0.0000	3.782E-06	0.0000	0.0000
18	0.0000	1.358E-07	11.0273	11.0273
19	0.0000	4.441E-03	0.0000	0.0044
20	0.0000	6.027E-02	10.9626	11.0229
21	0.0000	2.229E-02	0.0000	0.0223
22	0.0000	4.443E-01	10.4960	10.9403
23	0.1222	1.030E-05	0.0000	0.1222
24	0.0000	6.355E-02	10.3102	10.3738
25	0.0000	2.009E-02	0.0000	0.0201
26	0.0118	2.168E-06	10.2783	10.2901
27	0.4328	2.644E-04	0.0000	0.4331
28	0.2794	4.012E-06	9.5658	9.8452
29	0.0000	1.907E-05	0.0000	0.0000
30	0.0000	1.629E-06	9.5658	9.5658
31	3.1949	9.147E-03	0.0000	3.2040
32	0.0000	1.118E-06	6.3617	6.3617
33	0.0000	4.854E-06	0.0000	0.0000
34	0.0000	3.876E-06	6.3617	6.3617
35	0.2586	2.683E-02	0.0000	0.2855

SWS	WLA, Mg/yr	LA, Mg/yr	Upstream Contribution, Mg/yr	TMDL, Mg/yr
36	0.0000	9.040E-08	6.0762	6.0762
37	0.0000	5.479E-04	0.0748	0.0753
38	0.0000	6.082E-02	0.0000	0.0608
39	0.0000	1.395E-02	0.0000	0.0140
40	2.3212	6.547E-06	3.6797	6.0009
41	0.0000	5.354E-02	0.0000	0.0535
42	0.6758	3.000E-06	2.9504	3.6262
43	0.0000	4.350E-03	0.0000	0.0043
44	0.0000	3.569E-07	2.9461	2.9461
45	1.5091	2.353E-05	0.0000	1.5091
46	0.0931	1.249E-01	1.2189	1.4369
47	0.0000	1.850E-03	0.0000	0.0018
48	0.0139	9.422E-06	1.2031	1.2171
49	0.0000	6.575E-03	0.0000	0.0066
50	0.0000	3.085E-06	1.1966	1.1966
51	0.0000	1.621E-06	1.1038	1.1038
52	1.0509	6.523E-07	0.0060	1.0568
53	0.0000	9.497E-05	0.0000	0.0001
54	0.0000	5.870E-03	0.0000	0.0059
55	0.0000	4.693E-02	0.0000	0.0469
56	0.0000	5.367E-07	0.0928	0.0928
57	0.0000	2.009E-02	0.0000	0.0201
58	0.0000	1.772E-06	0.0727	0.0727
59	0.0000	2.745E-02	0.0000	0.0275
60	0.0000	2.229E-06	0.0452	0.0452
61	0.0000	4.522E-02	0.0000	0.0452
62	0.0000	6.988E-06	0.0000	0.0000

Table 5-6. NPDES Effluent Concentration Ranges for Wasteload Allocations.

Water Quality Parameter	Wasteload Allocation Range
PH	$\geq 6.0, \leq 9.0$ S.U.
Aluminum	0.75 - 4.3 mg/L
Iron	0.5 ¹ or 1.5 ² - 3.2 mg/L
Manganese	1.0 - 2.0 mg/L

¹Water quality standard for total iron concentration in trout waters.

²Water quality standard for total iron concentration in warm waters.

The wasteload allocations must be converted to permit average monthly limits and maximum daily limits according to the technical support document, which considers the type of water quality criteria (acute, chronic, human health, maximum allowable, four-day average, etc.), effluent variability, and monitoring requirements. For an iron wasteload allocation of 3.2 mg/l, the average monthly value is 3.0 mg/l, the maximum daily limit is 5.2 mg/l, the assumed effluent variability is 0.6, and two samples per month are required. A manganese wasteload allocation equal to 2.0 mg/L translates into an average monthly limit of 2.0 mg/L and a maximum daily limit of 3.5 mg/L. Presently aluminum is not limited in permits but will be required in any new or reissued permits. An aluminum wasteload allocation equal to 4.3 mg/L translates into an average monthly limit of 2.5 mg/L and a maximum daily limit of 4.3 mg/L.

5.4.2 Load Allocations

Load allocations were made for the following dominant source categories:

1. Abandoned mine lands.
2. Revoked mining permits.

The load allocations for acidity, aluminum, iron and manganese for all of the stream segments in the Paint Creek watershed are listed in both Appendix F and Tables 5-2, 5-3, 5-4, and 5-5. The bold rows in these tables correspond to stream segments placed on the 303(d) list for the appropriate contaminant. These load allocations are presented as annual loads, in terms of metric tons (Mg) per year. They are presented on an annual basis (as an annual load), because they were developed to meet TMDL endpoints under a range of conditions observed throughout the year. The water quality data collected from some of the Paint Creek sub-watersheds indicate the presence of abandoned mine land that is not found on any available GIS coverage of abandoned mine lands (AML). While the precise locations of these AML sites are not known, this load allocation assumes that stream water quality condition is the best indicator of the presence of AML.

5.4.3 Seasonal Variations

A TMDL must consider seasonal variation in the derivation of the allocation. For the Paint Creek watershed pH and metals TMDLs, seasonal variation was considered in the formulation of the modeling analysis. By using continuous simulation (modeling over a period of several years), seasonal hydrologic and source loading variability was inherently considered. The simulated pH and metals concentrations were calculated on a daily time step by the model were compared to TMDL endpoints. An allocation, which meets these endpoints throughout the year, was developed.

5.4.4 Future Growth

This TMDL does not include specific future growth allocations to each subwatershed. Because of the general allocation philosophy used in this TMDL, such allocations would be made at the expense of active mining point sources in the watershed. However, the absence of specific future growth allocations does not prohibit new mining in the watershed. Future growth could occur in the watershed under the following scenarios:

1. A new facility could be permitted anywhere in the watershed, provided that effluent limitations are based upon the achievement of water quality standards end-of-pipe for the pollutants of concern in the TMDL.
2. Remining could occur without a specific allocation to the new permittee, if the requirements of existing State remining regulations are achieved. Remining activities are viewed as a partial non-point source load reduction from Abandoned Mine Lands.
3. Reclamation and release of existing permits could provide an opportunity for future growth if permit release is conditioned upon achieving discharge quality better than the wasteload allocation prescribed by the TMDL.

The TMDL may be refined in the future through remodeling. Such refinement may incorporate new information and/or to the redistribute pollutant loads. Trading may provide an additional opportunity for future growth, contingent upon the State's development of a statewide or watershed-based trading program.

5.4.5 Water Quality Trading

This TMDL neither prohibits nor authorizes trading in the Paint Creek watershed. Both the WVDEP and EPA generally endorse the concept of trading, and recognize that it may become an effective tool for TMDL implementation.

However, significant regulatory framework development is necessary before large-scale trading in West Virginia may be realized. EPA will cooperate with the WVDEP in their development of a statewide or watershed-based trading program. Further, EPA supports program development assisted by a consensus-based stakeholder process.

Prior to the development of a formal trading program, it is conceivable that the regulation of specific point source to point source trades may be feasible under the framework of the NPDES program. EPA commits to cooperate with the WVDEP to facilitate such trades if trading opportunities arise and are proven environmentally beneficial.

6.0 Reasonable Assurance

Two primary programs are in effect which provide reasonable assurance for maintenance and improvement of water quality in the watershed. WVDEP's efforts to reclaim abandoned mine lands, coupled with its duties and responsibilities for issuance of NPDES permits, will be the focal points in water quality improvement.

Additional opportunities for water quality improvement are both ongoing and anticipated. Historically, a great deal of research into mine drainage has been conducted by scientists at West Virginia University, the West Virginia Division of Natural Resources, the United States Office of Surface Mining, the National Mine Land Reclamation Center, the National Environmental Training Laboratory and many other agencies and individuals. Funding from EPA's 319 grant program has been used extensively to remedy mine drainage impacts. This myriad of activity is expected to continue and result in water quality improvement.

6.1 Reclamation

Two distinct units of WVDEP reclaim land and water resources impacted by abandoned mines. The Office of Abandoned Mine Lands and Reclamation remedies eligible sites under Title IV of the Surface Mining Control and Reclamation Act of 1977. The Office of Mining and Reclamation's Special Reclamation Program remedies sites where operating permits and bonds have been revoked. Funding of the Office of Abandoned Mine Lands and Reclamation is derived from a federal tax on coal producers. The Special Reclamation Program is funded by the Special Reclamation Fund, which has primary sources of income from civil penalties, forfeited bonds, and a three-cent per ton fee on all coal produced.

A description of the operating procedures and accomplishments of each program follows.

6.1.1 Office of Abandoned Mine Lands and Reclamation

Title IV of the Surface Mining Control and Reclamation Act (Public Law 95-87) is designed to help reclaim and restore coal mine areas abandoned prior to August 3, 1977, throughout the country. The AML Program supplements existing state programs and allows the State of West Virginia to correct many abandoned mine related problems that would otherwise not be addressed. The major purpose of the AML Program is to reclaim and restore abandoned mine areas to protect the health, safety, and general welfare of the public and the environment. The AML Program corrects abandoned mine-related problems in accordance with the prioritization process specified in Public Law 95-87, Section 403 (a), 1-3. The priorities of the AML Program are as follows:

1. Protection of public health, safety, general welfare, and property from extreme danger of adverse effects related to coal mining practices.
2. Protection of public health, safety, and general welfare from adverse effects related to coal mining practices.
3. Restoration of the environment, including the land and water resources, that were degraded by adverse effects related to coal mining practices. This involves the conservation and development of soil, water (not channelization), woodland, fish and wildlife, recreational resources, and agricultural productivity.

Priority 1 and 2 problem areas include unsafe refuse piles, treacherous highwalls, pollution of domestic water supplies from mine drainage, mine fires, subsidence and other abandoned mine-related problems. The AML Program is now also focused on Priority 3 problem areas and on treating and abating water quality problems associated with abandoned mine lands but is not required by law or any statutory authority to do so. By recognizing the need to protect, and in many cases, improve the quality of the state's water resources from the impacts of mine drainage pollution from abandoned coal mines, coordinated efforts are now being employed to deal with this non-point source pollution problem.

Although OAML&R has been actively involved in the successful remediation of mine drainage pollution, inadequate funding and the lack of cost-effective mine drainage pollution treatment and abatement technologies have limited water quality improvement efforts. In 1990, the Surface Mining Control and Reclamation Act was amended to include a provision allowing states and tribes to establish an Acid Mine Drainage Treatment and Abatement Program and Fund. States and tribes may set-aside up to 10% of their annual grant to begin to address abandoned polluted coal mine drainage problems. Money from the Acid Mine Drainage Treatment and Abatement Fund can be utilized to clean-up mine drainage pollution at sites where mining ceased prior to August 3, 1977, and where no continuing reclamation responsibility can be determined. In order to qualify and be eligible, qualified hydrologic units or watersheds must be identified and water quality must adversely impact biological resources. A plan

must be prepared and presented to the Natural Resources Conservation Service for review and the Office of Surface Mining for approval. Plans that include the most cost-effective treatment and abatement alternatives, the greatest downstream benefits to the ecosystem, and diverse cooperators and stakeholders, will be the highest priority for approval.

AML&R has created an Acid Mine Drainage Abatement Policy to guide efforts in treating and abating mine drainage pollution. The Policy acts to guide the expenditure of funds in order to achieve the maximum amount of mine drainage pollution treatment within the boundaries imposed by budgetary and statutory constraints. The goal is to utilize existing technologies and practical economic considerations to maximize the amount of treatment for dollars expended. The policy includes a holistic watershed characterization and remediation procedure known as the Holistic Watershed Approach Protocol, which was developed and implemented by the Stream Restoration Group of AML&R. The Protocol involves diverse stakeholders in the establishment of various sampling networks and subsequent water quality data generation that focus remediation efforts. The Protocol is first used to subdivide the watershed into focus areas. More specific data is then generated to allow identification of the most feasible pollution sources to address and the best available pollution abatement technology to apply. The Protocol also includes the establishment of post-construction sampling networks to assess the impacts of remediation efforts. The Protocol is iteratively implemented until all focus areas have been addressed and all feasible pollution abatement technologies have been applied. A detailed description of the Protocol was presented by Vukovich and Adolfson (2001), which was modified to prepare Appendix G of this report.

6.1.2 Special Reclamation Group

When notice of permit revocation is received from the Director, a liability estimate is completed within 60 days of the revocation. The liability estimate notes any special health and safety characteristics of the site and calculates the cost to complete reclamation according to the permit reclamation plan. At sites where acid mine drainage is present, the permit is flagged for water quality characterization and a priority index assigned.

The reclamation plan at all sites includes the application of the best professional judgment to address the site specific problems including acid mine drainage. Any change or modification to the permit reclamation plan is done by or under the supervision of a Registered Professional Engineer. All construction requires application of best management practices to insure quality work and protect the environment.

Prioritization of bond forfeiture sites is consistent with the criteria used in the Abandoned Mine Land and Reclamation (AML&R) program. The criteria, listed

below in order of priority, have been used successfully for many years on abandoned mine areas with similar characteristics to bond forfeiture sites.

1. The highest priority sites are those that entail protection of public health, safety, general welfare, and property from extreme danger. There are relatively few of these types of bond forfeiture sites; however, they are unquestionably first order priorities.
2. Second order priority sites are those where public health, safety, welfare, and property values are judged to be threatened. Examples include sites with a high potential for landslides or flooding or the presence of dangerous highwalls, derelict buildings or other structures.
3. Third order priorities comprise the bulk of bond forfeiture sites. Therefore, this ranking level is sub-divided into smaller groupings.
 - 3.1. The first sub-group is sites that are causing or have a high potential for causing off-site environmental damage to the land and water resources. Such off-site damage would most likely be from heavy erosion, or high loadings of acid mine drainage.
 - 3.2. The second sub-group would include sites that are of a lower priority, but are in close geographic proximity to first or second priority sites. It is more efficient and cost effective to "cluster" projects where possible.
 - 3.3. The third sub-group includes sites near high-use public recreation areas and major thoroughfares.
 - 3.4. The fourth sub-group includes sites that are nearly fully reclaimed by the operator and only require monitoring of vegetative growth or other parameters. Sites, which have a real potential for re-permitting by another operator or reclamation by a third party, will also be placed in this sub-group.

Reclamation construction contracts occur by submittal of a detailed Project Requisition to the State Purchasing Division. All state purchasing policies and procedures are applicable and the contract is awarded to the lowest qualified bidder. Special Reclamation personnel perform inspection and contract management activities through the life of the contract. When all reclamation work is satisfactorily completed, a one-year contract warranty period begins to insure adequate vegetative growth and drainage system operation. Upon completion of the contract warranty period and recommendation of the Regional Supervisor, the permit status is classified as "completed." A completed status removes the liability of the forfeited site and terminates WVDEP jurisdiction and responsibility as a Phase III bond release.

At the sites with significant and high priority AMD, treatment operations are conducted to the extent of available funding, pursuant to the authority granted in 22-3-11 (g) of the West Virginia Surface Coal Mining and Reclamation Act. That regulation limits the annual expenditure of funds for designing, constructing and maintaining water treatment systems to 25% of the annual amount of the fees collected.

6.2 Permitting

NPDES permits in the watershed will be issued, reissued or modified by the Office of Water Resources in close cooperation with the Office of Mining and Reclamation. Both offices have adjusted permitting schedules to accommodate the State's Watershed Management Framework, thus implementation of TMDL requirements at existing facilities will generally occur at the time of scheduled permit re-issuance. Permits for existing facilities in the Paint Creek watershed are scheduled to be reissued in 2001. WVDEP may provide short-term administrative extensions to expiring permits to ensure adequate time is available to properly implement the wasteload allocations of this TMDL.

7.0 Public Participation

EPA policy is that there must be full and meaningful public participation in the TMDL development process. Each state must, therefore, provide for public participation consistent with its own continuing planning process and public participation requirements. As a result, it is the intent of the WVDEP to solicit public input by providing opportunities for public comment and review of the draft TMDLs. The public meetings pertaining to the Paint Creek watershed occurred as follows:

May 10, 2000	Public meeting presenting an introduction to the TMDL process, together with the requirements of the consent decree.
February 27, 2001	Public meeting presented by WVDEP, EPA and WVU.
July 25 – September 7, 2001	45-day public comment period, published in the Charleston Gazette and Fayette Tribune
August 27, 2001	Public hearing held by WVDEP and EPA

8.0 References

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Appendix A Mining NPDES Permits in Paint Creek Watershed.

Table A1. Paint Creek NPDES Permits and Associated Article 3 & 4 Permits.

NPDES Permit	Associated Article 3 & 4 Permit # and Status					Revoked
	Not Started	Active	Phase 1	Phase 2	Phase 3	
WV1001370		O304887				
WV0096679			O017083			
WV1002066				S301089		
WV1012631					U017700	
WV0097071						U022583
WV1012487		S300295				
WV1012487		S303991				
WV1012487					S302389	
WV1012592		S304191				
WV1014951		O300595				
WV1014951		Q301696				
WV1014951		S302794				
WV1015214		S301496				
WV0057011		O602289				
WV0057011		H066700				
WV0092142		Q000182				
WV0092142		Q013380				
WV0092142		Q020078				
WV0092142		Q302286				
WV0092142						Q009873
WV1009311		S300795				
WV1009311		S304387				
WV1009311		S602089				
WV1002074		O301489				
WV1002074				S300391		
WV1002074				S301589		
WV1002074				S302186		
WV0028452		O301993				
WV0028452		O301198				
WV0028452					P069400	
WV0028452		S007480				
WV0028452		U042800				
WV0028452			U300395			
WV0028452		U300496				
WV0028452			U300591			
WV0028452					U300596	
WV0028452	U300597					
WV0028452					U302290	
WV0028452				U302591		
WV1002368		U301996				
WV1002368		U302990				
WV1002368					U304489	
WV1019317		O0301489				
WV1015257		O601186				

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV0057011	001	38 05 32.0 N	81 20 59.0 W	Reissued	Technology Based	Inactive	
WV0057011	002	38 05 31.0 N	81 20 57.0 W	Reissued	Technology Based	Inactive	
WV0057011	003	38 05 21.0 N	81 21 22.0 W	Reissued	Technology Based	Inactive	
WV0057011	004	38 05 40.0 N	81 21 03.0 W	Reissued	Technology Based	Inactive	
WV0057011	005	38 05 47.0 N	81 20 56.0 W	Reissued	Technology Based	Inactive	
WV0057011	006	38 05 40.0 N	81 21 15.0 W	Reissued	Technology Based	Inactive	
WV0057011	007	38 05 27.0 N	81 21 41.0 W	Reissued	Technology Based	Inactive	
WV0057011	008	38 05 26.0 N	81 21 37.0 W	Reissued	Technology Based	Inactive	
WV0057011	009	38 05 25.0 N	81 21 19.0 W	Reissued	Technology Based	Inactive	
WV0057011	010	38 05 14.0 N	81 21 06.0 W	Reissued	Technology Based	Inactive	
WV0057011	011	38 05 11.0 N	81 20 51.0 W	Reissued	Technology Based	Inactive	
WV0057011	012	38 05 31.0 N	81 21 47.0 W	Reissued	Technology Based	Inactive	
WV0057011	013	38 05 43.0 N	81 21 44.0 W	Reissued	Technology Based	Inactive	
WV0057011	014	38 05 21.0 N	81 21 22.0 W	Reissued	Technology Based	Inactive	
WV0057011	015	38 05 17.0 N	81 21 16.0 W	Reissued	Technology Based	Inactive	
WV0057011	016	38 05 19.0 N	81 21 22.0 W	Reissued	Technology Based	Inactive	
WV0057011	017	38 05 34.0 N	81 21 12.0 W	Reissued	Technology Based	Inactive	
WV0057011	018	38 04 35.0 N	81 21 09.0 W	Reissued	Technology Based	Active	
WV0057011	019	38 03 16.0 N	81 20 43.0 W	Reissued	Technology Based	Inactive	
WV0092142	001	37 51 26.0 N	81 15 07.0 W	Open	Technology Based	Active	Sandstone Quarry
WV0092142	002	37 51 25.0 N	81 14 55.0 W	Open	Technology Based	Active	Sandstone Quarry
WV0092142	003	37 51 17.0 N	81 15 01.0 W	Open	Technology Based	Inactive	Sandstone Quarry
WV0092142	004	37 51 24.0 N	81 15 01.0 W	Open	Technology Based	Inactive	Sandstone Quarry
WV0096679	001	37 59 57.0 N	81 20 34.0 W	Open	Technology Based	Inactive	
WV1001370	001	37 50 22.0 N	81 15 47.0 W	Open	Technology Based	Inactive	
WV1002066	001	37 55 58.0 N	81 17 10.0 W	Open	Technology Based	Active	
WV1002066	002	37 56 01.0 N	81 17 01.0 W	Open	Technology Based	Active	
WV1002066	003	37 56 01.0 N	81 17 29.0 W	Open	Technology Based	Active	
WV1002066	004	37 56 18.0 N	81 17 14.0 W	Open	Technology Based	Active	
WV1002066	005	37 56 22.0 N	81 17 11.0 W	Open	Technology Based	Active	
WV1002066	006	37 56 07.0 N	81 17 22.0 W	Open	Technology Based	Active	
WV1002368	001	38 00 07.0 N	81 19 11.0 W	Open	Technology Based	Active	
WV1002368	002	37 59 42.0 N	81 19 09.0 W	Open	Technology Based	Active	
WV1002368	003	37 59 40.0 N	81 19 11.0 W	Open	Technology Based	Active	

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV1002368	004	38 00 18.0 N	81 19 08.0 W	Open	Technology Based	Active	
WV1002368	005	38 00 12.0 N	81 19 06.0 W	Open	Technology Based	Active	
WV1009311	001	38 02 47.0 N	81 20 25.0 W	Open	Technology Based	Inactive	
WV1009311	002	38 03 15.0 N	81 20 30.0 W	Open	Technology Based	Inactive	
WV1009311	003	38 03 35.0 N	81 20 12.0 W	Open	Technology Based	Inactive	
WV1009311	004	38 03 16.0 N	81 21 56.0 W	Open	Technology Based	Active	
WV1009311	005	38 03 37.0 N	81 22 09.0 W	Open	Technology Based	Active	
WV1009311	006	38 02 41.0 N	81 21 45.0 W	Open	Technology Based	Active	
WV1009311	007	38 04 05.0 N	81 20 24.0 W	Open	Technology Based	Inactive	
WV1009311	008	38 04 20.0 N	81 22 24.0 W	Open	Technology Based	Inactive	
WV1009311	009	38 04 17.0 N	81 22 20.0 W	Open	Technology Based	Inactive	
WV1009311	010	38 04 10.0 N	81 22 13.0 W	Open	Technology Based	Inactive	
WV1009311	011	38 04 02.0 N	81 22 01.0 W	Open	Technology Based	Inactive	
WV1009311	012	38 03 56.0 N	81 21 48.0 W	Open	Technology Based	Inactive	
WV1009311	013	38 03 53.0 N	81 21 35.0 W	Open	Technology Based	Inactive	
WV1009311	014	38 03 54.0 N	81 21 27.0 W	Open	Technology Based	Inactive	
WV1009311	015	38 03 57.0 N	81 21 13.0 W	Open	Technology Based	Inactive	
WV1009311	016	38 04 08.0 N	81 20 37.0 W	Open	Technology Based	Inactive	
WV1009311	017	38 04 36.0 N	81 21 59.0 W	Open	Technology Based	Inactive	
WV1009311	018	38 04 23.0 N	81 22 15.0 W	Open	Technology Based	Inactive	
WV1009311	055	38 04 23.0 N	81 22 15.0 W	Open	Technology Based	Inactive	Estimated Location
WV1012487	003	37 57 40.0 N	81 16 20.0 W	Open	Technology Based	Active	
WV1012487	004	37 57 37.0 N	81 16 25.0 W	Open	Technology Based	Active	
WV1012487	005	37 57 30.0 N	81 16 11.0 W	Open	Technology Based	Active	
WV1012487	006	37 57 12.0 N	81 16 16.0 W	Open	Technology Based	Active	
WV1012487	007	37 57 17.0 N	81 15 44.0 W	Open	Technology Based	Active	
WV1012487	008	37 57 17.0 N	81 15 42.0 W	Open	Technology Based	Active	
WV1012487	009	37 57 13.0 N	81 15 43.0 W	Open	Technology Based	Active	
WV1012487	010	37 57 07.0 N	81 15 47.0 W	Open	Technology Based	Active	
WV1012487	011	37 57 06.0 N	81 15 38.0 W	Open	Technology Based	Active	
WV1012487	012	37 57 04.0 N	81 15 39.0 W	Open	Technology Based	Active	
WV1012487	013	37 56 54.0 N	81 15 31.0 W	Open	Technology Based	Active	
WV1012487	014	37 56 59.0 N	81 15 37.0 W	Open	Technology Based	Active	
WV1012487	015	37 56 53.0 N	81 15 36.0 W	Open	Technology Based	Active	

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV1012487	016	37 56 49.0 N	81 15 36.0 W	Open	Technology Based	Active	
WV1012487	017	37 56 50.0 N	81 15 40.0 W	Open	Technology Based	Active	
WV1012487	018	37 56 46.0 N	81 15 56.0 W	Open	Technology Based	Inactive	
WV1012487	019	37 56 42.0 N	81 16 03.0 W	Open	Technology Based	Active	
WV1012487	020	37 56 39.0 N	81 15 56.0 W	Open	Technology Based	Active	
WV1012487	021	37 56 34.0 N	81 15 34.0 W	Open	Technology Based	Active	
WV1012487	022	37 56 36.0 N	81 15 58.0 W	Open	Technology Based	Active	
WV1012487	023	37 56 30.0 N	81 15 58.0 W	Open	Technology Based	Active	
WV1012487	024	37 56 18.0 N	81 15 56.0 W	Open	Technology Based	Active	
WV1012487	025	37 56 21.0 N	81 16 11.0 W	Open	Technology Based	Active	
WV1012487	026	37 56 18.0 N	81 16 05.0 W	Open	Technology Based	Active	
WV1012487	027	37 56 11.0 N	81 16 40.0 W	Open	Technology Based	Active	
WV1012487	028	37 56 11.0 N	81 16 35.0 W	Open	Technology Based	Inactive	
WV1012487	029	37 56 12.0 N	81 16 27.0 W	Open	Technology Based	Active	
WV1012487	030	37 56 17.0 N	81 16 22.0 W	Open	Technology Based	Active	
WV1012487	031	37 56 06.0 N	81 16 41.0 W	Open	Technology Based	Active	
WV1012487	032	37 57 19.0 N	81 15 54.0 W	Open	Technology Based	Active	
WV1012487	033	37 57 15.0 N	81 15 52.0 W	Open	Technology Based	Active	
WV1012487	034	37 57 16.0 N	81 15 49.0 W	Open	Technology Based	Active	
WV1012487	035	37 57 12.0 N	81 16 15.0 W	Open	Technology Based	Active	
WV1012487	036	37 57 21.0 N	81 16 18.0 W	Open	Technology Based	Active	
WV1012487	037	37 57 15.0 N	81 16 18.0 W	Open	Technology Based	Active	
WV1012487	038	37 57 08.0 N	81 16 09.0 W	Open	Technology Based	Active	
WV1012487	039	37 57 14.0 N	81 16 10.0 W	Open	Technology Based	Active	
WV1012487	040	37 57 16.0 N	81 16 06.0 W	Open	Technology Based	Active	
WV1012487	041	37 57 16.0 N	81 16 02.0 W	Open	Technology Based	Active	
WV1012487	042	37 57 11.0 N	81 16 00.0 W	Open	Technology Based	Active	
WV1012487	043	37 55 54.0 N	81 17 23.0 W	Open	Technology Based	Active	
WV1012487	044	37 56 11.0 N	81 16 51.0 W	Open	Technology Based	Active	
WV1012487	045	37 56 15.0 N	81 16 56.0 W	Open	Technology Based	Active	
WV1012487	046	37 56 07.0 N	81 16 56.0 W	Open	Technology Based	Active	
WV1012487	047	37 56 03.0 N	81 16 55.0 W	Open	Technology Based	Active	
WV1012487	048	37 56 24.0 N	81 16 43.0 W	Open	Technology Based	Active	
WV1012487	049	37 56 26.0 N	81 16 42.0 W	Open	Technology Based	Active	

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV1012487	050	37 56 26.0 N	81 16 36.0 W	Open	Technology Based	Active	
WV1012487	051	37 56 31.0 N	81 16 26.0 W	Open	Technology Based	Active	
WV1012487	052	37 56 36.0 N	81 16 26.0 W	Open	Technology Based	Active	
WV1012487	053	37 56 35.0 N	81 16 22.0 W	Open	Technology Based	Active	
WV1012487	054	37 56 31.0 N	81 16 19.0 W	Open	Technology Based	Active	
WV1012487	055	37 56 28.0 N	81 16 14.0 W	Open	Technology Based	Active	
WV1012487	056	37 56 37.0 N	81 16 02.0 W	Open	Technology Based	Active	
WV1012487	057	37 56 40.0 N	81 16 04.0 W	Open	Technology Based	Active	
WV1012487	058	37 56 42.0 N	81 16 11.0 W	Open	Technology Based	Active	
WV1012487	059	37 56 47.0 N	81 16 03.0 W	Open	Technology Based	Active	
WV1012487	060	37 56 50.0 N	81 15 59.0 W	Open	Technology Based	Active	
WV1012487	061	37 56 52.0 N	81 15 53.0 W	Open	Technology Based	Active	
WV1012487	062	37 56 54.0 N	81 15 48.0 W	Open	Technology Based	Active	
WV1012487	063	37 57 03.0 N	81 15 51.0 W	Open	Technology Based	Active	
WV1012487	064	37 57 07.0 N	81 15 52.0 W	Open	Technology Based	Active	
WV1012487	065	37 57 09.0 N	81 15 52.0 W	Open	Technology Based	Active	
WV1012487	066	37 57 06.0 N	81 15 56.0 W	Open	Technology Based	Active	
WV1012487	067	37 57 03.0 N	81 16 00.0 W	Open	Technology Based	Active	
WV1012487	068	37 57 07.0 N	81 16 02.0 W	Open	Technology Based	Active	
WV1012487	069	37 57 06.0 N	81 16 03.0 W	Open	Technology Based	Active	
WV1012487	070	37 57 01.0 N	81 16 04.0 W	Open	Technology Based	Active	
WV1012487	071	37 57 03.0 N	81 16 08.0 W	Open	Technology Based	Active	
WV1012487	072	37 57 02.0 N	81 16 09.0 W	Open	Technology Based	Active	
WV1012487	073	37 56 56.0 N	81 16 10.0 W	Open	Technology Based	Active	
WV1012487	074	37 57 02.0 N	81 16 21.0 W	Open	Technology Based	Active	
WV1012487	075	37 57 02.0 N	81 16 31.0 W	Open	Technology Based	Active	
WV1012487	076	37 57 06.0 N	81 16 36.0 W	Open	Technology Based	Active	
WV1012487	077	37 57 10.0 N	81 16 37.0 W	Open	Technology Based	Active	
WV1012487	078	37 57 12.0 N	81 16 40.0 W	Open	Technology Based	Active	
WV1012487	079	37 57 14.0 N	81 16 34.0 W	Open	Technology Based	Active	
WV1012487	080	37 57 15.0 N	81 16 28.0 W	Open	Technology Based	Active	
WV1012487	081	37 57 16.0 N	81 16 31.0 W	Open	Technology Based	Active	
WV1012487	082	37 57 19.0 N	81 16 36.0 W	Open	Technology Based	Active	
WV1012487	083	37 57 19.0 N	81 16 42.0 W	Open	Technology Based	Active	

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV1012487	084	37 57 24.0 N	81 16 37.0 W	Open	Technology Based	Active	
WV1012487	085	37 57 34.0 N	81 16 41.0 W	Open	Technology Based	Active	
WV1012487	086	37 57 36.0 N	81 16 33.0 W	Open	Technology Based	Active	
WV1012487	087	37 57 39.0 N	81 16 30.0 W	Open	Technology Based	Active	
WV1012487	088	37 57 32.0 N	81 16 46.0 W	Open	Technology Based	Active	
WV1012487	089	37 57 28.0 N	81 16 49.0 W	Open	Technology Based	Active	
WV1012487	090	37 57 25.0 N	81 16 49.0 W	Open	Technology Based	Active	
WV1012487	091	37 56 34.0 N	81 16 49.0 W	Open	Technology Based	Active	
WV1012487	092	37 56 30.0 N	81 16 46.0 W	Open	Technology Based	Active	
WV1012487	093	37 56 30.0 N	81 16 43.0 W	Open	Technology Based	Active	
WV1012487	094	37 56 31.0 N	81 16 41.0 W	Open	Technology Based	Active	
WV1012487	095	37 56 35.0 N	81 16 43.0 W	Open	Technology Based	Active	
WV1012487	096	37 56 38.0 N	81 16 41.0 W	Open	Technology Based	Active	
WV1012487	097	37 56 41.0 N	81 16 42.0 W	Open	Technology Based	Active	
WV1012487	098	37 56 42.0 N	81 16 39.0 W	Open	Technology Based	Active	
WV1012487	099	37 56 40.0 N	81 16 34.0 W	Open	Technology Based	Active	
WV1012487	100	37 56 40.0 N	81 16 30.0 W	Open	Technology Based	Active	
WV1012487	101	37 56 41.0 N	81 16 27.0 W	Open	Technology Based	Inactive	
WV1012487	102	37 56 42.0 N	81 16 22.0 W	Open	Technology Based	Inactive	
WV1012487	103	37 56 40.0 N	81 16 17.0 W	Open	Technology Based	Active	
WV1012487	104	37 56 44.0 N	81 16 17.0 W	Open	Technology Based	Active	
WV1012487	105	37 56 47.0 N	81 16 13.0 W	Open	Technology Based	Active	
WV1012487	106	37 56 51.0 N	81 16 05.0 W	Open	Technology Based	Active	
WV1012487	107	37 56 54.0 N	81 56 02.0 W	Open	Technology Based	Inactive	
WV1012487	108	37 56 52.0 N	81 16 06.0 W	Open	Technology Based	Active	
WV1012487	109	37 56 56.0 N	81 16 35.0 W	Open	Technology Based	Active	
WV1012487	110	37 57 03.0 N	81 16 48.0 W	Open	Technology Based	Active	
WV1012487	111	37 56 54.0 N	81 15 39.0 W	Open	Technology Based	Active	
WV1012487	112	37 56 24.0 N	81 16 08.0 W	Open	Technology Based	Active	
WV1012487	113	37 56 24.0 N	81 16 26.0 W	Open	Technology Based	Active	
WV1012487	114	37 56 21.0 N	81 16 28.0 W	Open	Technology Based	Active	
WV1012487	115	37 56 42.0 N	81 15 54.0 W	Open	Technology Based	Active	
WV1012487	116	38 57 47.0 N	81 17 09.0 W	Open	Technology Based	Inactive	
WV1012487	117	38 57 29.0 N	81 17 16.0 W	Open	Technology Based	Inactive	

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV1012487	118	38 57 19.0 N	81 17 27.0 W	Open	Technology Based	Inactive	
WV1012487	119	38 57 11.0 N	81 17 32.0 W	Open	Technology Based	Inactive	
WV1012487	120	38 57 41.0 N	81 16 42.0 W	Open	Technology Based	Inactive	
WV1012487	121	38 57 59.0 N	81 16 42.0 W	Open	Technology Based	Inactive	
WV1012487	122	38 57 59.0 N	81 16 57.0 W	Open	Technology Based	Inactive	
WV1012487	123	38 57 37.0 N	81 17 09.0 W	Open	Technology Based	Inactive	
WV1012487	124	38 57 22.0 N	81 17 21.0 W	Open	Technology Based	Inactive	
WV1012487	125	38 56 55.0 N	81 17 38.0 W	Open	Technology Based	Inactive	
WV1012487	126	38 57 15.0 N	81 17 07.0 W	Open	Technology Based	Inactive	
WV1012487	127	38 57 33.0 N	81 17 04.0 W	Open	Technology Based	Inactive	
WV1012487	128	38 57 47.0 N	81 16 50.0 W	Open	Technology Based	Inactive	
WV1012487	129	38 57 20.0 N	81 16 47.0 W	Open	Technology Based	Inactive	
WV1012487	130	38 57 12.0 N	81 16 45.0 W	Open	Technology Based	Inactive	
WV1012487	131	38 56 34.0 N	81 17 26.0 W	Open	Technology Based	Inactive	
WV1012487	132	38 55 52.0 N	81 17 19.0 W	Open	Technology Based	Inactive	
WV1012487	133	38 56 05.0 N	81 16 38.0 W	Open	Technology Based	Inactive	
WV1012487	134	38 56 03.0 N	81 16 11.0 W	Open	Technology Based	Inactive	
WV1012487	136	38 55 52.0 N	81 17 19.0 W	Open	Technology Based	Inactive	
WV1012487	137	38 55 45.0 N	81 17 19.0 W	Open	Technology Based	Inactive	
WV1012487	138	38 55 49.0 N	81 17 03.0 W	Open	Technology Based	Inactive	
WV1012487	139	38 55 55.0 N	81 16 59.0 W	Open	Technology Based	Inactive	
WV1012487	140	38 55 55.0 N	81 16 55.0 W	Open	Technology Based	Inactive	
WV1012487	141	38 56 02.0 N	81 16 52.0 W	Open	Technology Based	Inactive	
WV1012487	142	38 56 00.0 N	81 16 35.0 W	Open	Technology Based	Inactive	
WV1012487	143	38 56 03.0 N	81 16 32.0 W	Open	Technology Based	Inactive	
WV1012487	144	38 55 58.0 N	81 16 14.0 W	Open	Technology Based	Inactive	
WV1012487	145	37 56 07.0 N	81 17 14.0 W	Open	Technology Based	Active	
WV1012487	146	37 56 20.0 N	81 17 07.0 W	Open	Technology Based	Active	
WV1012487	147	37 56 24.0 N	81 17 02.0 W	Open	Technology Based	Active	
WV1012487	148	37 56 19.0 N	81 16 50.0 W	Open	Technology Based	Active	
WV1012487	149	37 55 58.0 N	81 17 10.0 W	Open	Technology Based	Active	
WV1012487	150	37 56 09.0 N	81 16 48.0 W	Open	Technology Based	Active	
WV1012487	151	37 57 23.0 N	81 16 54.0 W	Open	Technology Based	Active	
WV1012487	152	37 57 17.0 N	81 16 53.0 W	Open	Technology Based	Active	

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV1012487	153	37 57 11.0 N	81 16 57.0 W	Open	Technology Based	Active	
WV1012487	154	37 57 13.0 N	81 17 06.0 W	Open	Technology Based	Active	
WV1012487	155	37 57 05.0 N	81 17 04.0 W	Open	Technology Based	Inactive	
WV1012487	156	37 56 58.0 N	81 17 06.0 W	Open	Technology Based	Active	
WV1012487	157	37 56 53.0 N	81 17 10.0 W	Open	Technology Based	Active	
WV1012487	158	37 56 52.0 N	81 17 15.0 W	Open	Technology Based	Active	
WV1012487	159	37 57 00.0 N	81 17 20.0 W	Open	Technology Based	Active	
WV1012487	160	37 56 55.0 N	81 17 37.0 W	Open	Technology Based	Active	
WV1012487	161	37 56 49.0 N	81 17 27.0 W	Open	Technology Based	Active	
WV1012487	162	37 56 57.0 N	81 17 19.0 W	Open	Technology Based	Active	
WV1012487	163	37 56 51.0 N	81 17 16.0 W	Open	Technology Based	Active	
WV1012487	164	37 56 52.0 N	81 17 10.0 W	Open	Technology Based	Active	
WV1012487	165	37 56 57.0 N	81 17 05.0 W	Open	Technology Based	Active	
WV1012487	166	37 57 04.0 N	81 17 03.0 W	Open	Technology Based	Active	
WV1012487	167	37 56 42.0 N	81 16 11.0 W	Open	Technology Based	Active	
WV1012487	168	37 56 31.0 N	81 17 32.0 W	Open	Technology Based	Inactive	
WV1012487	169	37 56 37.0 N	81 17 36.0 W	Open	Technology Based	Inactive	
WV1012592	001	38 22 32.0 N	81 20 05.0 W	Open	Technology Based	Inactive	
WV1012592	002	38 21 51.0 N	81 19 52.0 W	Open	Technology Based	Inactive	
WV1012592	003	38 21 44.0 N	81 20 39.0 W	Open	Technology Based	Inactive	
WV1012592	004	38 22 07.0 N	81 21 05.0 W	Open	Technology Based	Inactive	
WV1012592	005	38 22 21.0 N	81 21 14.0 W	Open	Technology Based	Inactive	
WV1012592	006	38 22 39.0 N	81 20 52.0 W	Open	Technology Based	Inactive	
WV1012592	007	38 21 42.0 N	81 21 15.0 W	Open	Technology Based	Inactive	
WV1012592	008	38 22 53.0 N	81 19 55.0 W	Open	Technology Based	Inactive	
WV1012592	009	38 02 33.0 N	81 19 52.0 W	Open	Technology Based	Inactive	
WV1012592	010	38 02 17.0 N	81 19 54.0 W	Open	Technology Based	Inactive	
WV1012592	011	38 01 51.0 N	81 20 15.0 W	Open	Technology Based	Inactive	
WV1012592	012	38 01 58.0 N	81 20 31.0 W	Open	Technology Based	Inactive	
WV1014951	001	37 58 04.0 N	81 21 23.0 W	Open	Technology Based	Inactive	
WV1014951	002	37 57 33.0 N	81 20 37.0 W	Open	Technology Based	Inactive	
WV1014951	003	37 58 29.0 N	81 21 20.0 W	Open	Technology Based	Active	
WV1014951	004	37 58 38.0 N	81 21 15.0 W	Open	Technology Based	Active	
WV1014951	005	37 58 44.0 N	81 21 05.0 W	Open	Technology Based	Active	

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV1014951	006	37 58 44.0 N	81 20 51.0 W	Open	Technology Based	Active	
WV1014951	007	37 58 54.0 N	81 20 50.0 W	Open	Technology Based	Active	
WV1014951	008	37 59 01.0 N	81 20 50.0 W	Open	Technology Based	Active	
WV1014951	009	37 58 53.0 N	81 20 44.0 W	Open	Technology Based	Active	
WV1014951	010	37 58 46.0 N	81 20 44.0 W	Open	Technology Based	Active	
WV1014951	011	37 58 35.0 N	81 20 41.0 W	Open	Technology Based	Active	
WV1014951	012	37 58 27.0 N	81 20 41.0 W	Open	Technology Based	Active	
WV1014951	013	37 58 20.0 N	81 20 33.0 W	Open	Technology Based	Active	
WV1014951	014	37 58 20.0 N	81 20 24.0 W	Open	Technology Based	Active	
WV1014951	015	37 57 35.0 N	81 27 00.0 W	Open	Technology Based	Inactive	
WV1014951	016	37 58 04.0 N	81 20 15.0 W	Open	Technology Based	Inactive	
WV1014951	017	37 57 55.0 N	81 20 24.0 W	Open	Technology Based	Inactive	
WV1014951	018	37 57 45.0 N	81 20 59.0 W	Open	Technology Based	Inactive	
WV1014951	019	37 57 36.0 N	81 21 09.0 W	Open	Technology Based	Inactive	
WV1014951	020	37 57 32.0 N	81 21 09.0 W	Open	Technology Based	Inactive	
WV1014951	021	37 57 33.0 N	81 21 15.0 W	Open	Technology Based	Inactive	
WV1014951	022	37 57 40.0 N	81 21 27.0 W	Open	Technology Based	Inactive	
WV1014951	023	37 57 37.0 N	81 21 35.0 W	Open	Technology Based	Inactive	
WV1014951	024	37 57 37.0 N	81 21 40.0 W	Open	Technology Based	Inactive	
WV1014951	025	37 57 41.0 N	81 21 35.0 W	Open	Technology Based	Inactive	
WV1014951	026	37 57 45.0 N	81 21 28.0 W	Open	Technology Based	Inactive	
WV1014951	027	37 57 47.0 N	81 21 17.0 W	Open	Technology Based	Inactive	
WV1014951	028	37 57 52.0 N	81 21 08.0 W	Open	Technology Based	Inactive	
WV1014951	029	37 58 00.0 N	81 21 03.0 W	Open	Technology Based	Inactive	
WV1014951	030	37 58 06.0 N	81 21 04.0 W	Open	Technology Based	Inactive	
WV1014951	031	37 58 20.0 N	81 19 17.0 W	Open	Technology Based	Active	
WV1014951	032	37 58 28.0 N	81 19 26.0 W	Open	Technology Based	Active	
WV1014951	033	37 58 18.0 N	81 19 14.0 W	Open	Technology Based	Active	
WV1014951	034	37 57 29.0 N	81 27 09.0 W	Open	Technology Based	Inactive	
WV1014951	035	37 58 11.0 N	81 20 18.0 W	Open	Technology Based	Active	
WV1014951	036	37 58 13.0 N	81 20 17.0 W	Open	Technology Based	Active	
WV1014951	037	37 58 14.0 N	81 20 15.0 W	Open	Technology Based	Active	
WV1014951	038	37 58 23.0 N	81 19 19.0 W	Open	Technology Based	Active	
WV1014951	039	37 58 14.0 N	81 20 24.0 W	Open	Technology Based	Active	

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV1014951	040	37 58 17.0 N	81 20 28.0 W	Open	Technology Based	Active	
WV1014951	041	37 58 48.0 N	81 20 44.0 W	Open	Technology Based	Active	
WV1014951	042	37 58 38.0 N	81 21 21.0 W	Open	Technology Based	Active	
WV1014951	043	37 58 35.0 N	81 19 32.0 W	Open	Technology Based	Active	
WV1014951	044	37 58 08.0 N	81 21 00.0 W	Open	Technology Based	Inactive	
WV1014951	045	37 58 25.0 N	81 20 17.0 W	Open	Technology Based	Active	
WV1014951	046	37 58 04.0 N	81 21 26.0 W	Open	Technology Based	Inactive	
WV0028452	001	37 58 22.0 N	81 18 17.0 W	Open	Technology Based	Active	
WV0028452	002	37 56 16.0 N	81 19 19.0 W	Open	Technology Based	Inactive	
WV0028452	003	37 58 07.0 N	81 18 31.0 W	Open	Technology Based	Active	
WV0028452	004	37 58 20.0 N	81 18 21.0 W	Open	Technology Based	Active	
WV0028452	005	37 58 15.0 N	81 18 27.0 W	Open	Technology Based	Active	
WV0028452	006	37 58 33.0 N	81 18 06.0 W	Open	Technology Based	Active	
WV0028452	007	37 58 15.0 N	81 18 05.0 W	Open	Technology Based	Inactive	
WV0028452	008	37 57 50.0 N	81 18 28.0 W	Open	Technology Based	Active	
WV0028452	009	37 57 32.0 N	81 18 55.0 W	Open	Technology Based	Active	
WV0028452	010	37 56 16.0 N	81 18 55.0 W	Open	Technology Based	Inactive	
WV0028452	012	37 57 42.0 N	81 18 53.0 W	Open	Technology Based	Active	
WV0028452	013	37 57 43.0 N	81 18 48.0 W	Open	Technology Based	Active	
WV0028452	014	37 57 55.0 N	81 18 44.0 W	Open	Technology Based	Active	
WV0028452	015	37 57 45.0 N	81 18 46.0 W	Open	Technology Based	Active	
WV1002074	001	37 54 34.0 N	81 18 11.0 W	Open	Technology Based	Inactive	
WV1002074	002	37 54 45.0 N	81 18 11.0 W	Open	Technology Based	Inactive	
WV1002074	003	37 54 44.0 N	81 18 15.0 W	Open	Technology Based	Inactive	
WV1002074	005	37 54 49.0 N	81 18 19.0 W	Open	Technology Based	Inactive	
WV1002074	006	37 54 55.0 N	81 18 24.0 W	Open	Technology Based	Inactive	
WV1002074	007	37 54 52.0 N	81 18 33.0 W	Open	Technology Based	Inactive	
WV1002074	008	37 54 49.0 N	81 18 41.0 W	Open	Technology Based	Inactive	
WV1002074	009	37 54 55.0 N	81 18 38.0 W	Open	Technology Based	Inactive	
WV1002074	014	37 54 55.0 N	81 18 24.0 W	Open	Technology Based	Inactive	
WV1002074	015	37 54 55.0 N	81 18 24.0 W	Open	Technology Based	Inactive	
WV1002074	016	37 54 55.0 N	81 18 24.0 W	Open	Technology Based	Inactive	
WV1002074	017	37 54 55.0 N	81 18 24.0 W	Open	Technology Based	Inactive	
WV1002074	018	37 54 55.0 N	81 18 24.0 W	Open	Technology Based	Inactive	

Table A2. Discharge Outlets for Paint Creek NPDES Permits.

NPDES Number	NPDES Outlet	Outlet Latitude	Outlet Longitude	Permit Status	Discharge Limits	Current Activity	Comments
WV1002074	019	37 54 55.0 N	81 18 24.0 W	Open	Technology Based	Inactive	
WV1002074	020	37 54 55.0 N	81 18 24.0 W	Open	Technology Based	Inactive	
WV1002074	021	37 54 53.0 N	81 18 49.0 W	Open	Technology Based	Inactive	
WV1002074	022	37 55 08.0 N	81 18 36.0 W	Open	Technology Based	Inactive	
WV1002074	023	37 55 02.0 N	81 18 31.0 W	Open	Technology Based	Inactive	
WV1002074	024	37 55 15.0 N	81 18 25.0 W	Open	Technology Based	Inactive	
WV1002074	025	37 55 20.0 N	81 18 16.0 W	Open	Technology Based	Inactive	
WV1002074	026	37 54 39.0 N	81 17 24.0 W	Open	Technology Based	Active	
WV1002074	027	37 54 51.0 N	81 17 39.0 W	Open	Technology Based	Active	
WV1002074	028	37 54 56.0 N	81 17 39.0 W	Open	Technology Based	Active	
WV1002074	029	37 54 44.0 N	81 17 00.0 W	Open	Technology Based	Active	
WV1002074	030	37 55 26.0 N	81 17 03.0 W	Open	Technology Based	Active	
WV1002074	031	37 55 32.0 N	81 17 03.0 W	Open	Technology Based	Active	
WV1002074	032	37 55 12.0 N	81 17 04.0 W	Open	Technology Based	Active	
WV1019317	001	37 55 12.0 N	81 17 04.0 W	Open	Technology Based	Inactive	Approximate Location
WV1015257	002	38 00 09.0 N	81 22 30.0 W	Open	Technology Based	Active	

Table A3. Monthly Reported Discharge Rate for Paint Creek NPDES Outlets.

There are various periods of record for these outlets.

NPDES Outlet	Jan (gpm)	Feb (gpm)	Mar (gpm)	Apr (gpm)	May (gpm)	Jun (gpm)	Jul (gpm)	Aug (gpm)	Sep (gpm)	Oct (gpm)	Nov (gpm)	Dec (gpm)
WV0028452_001	83.9	103.4	123.9	102.1	118.3	122.6	60.3	45.8	63.9	58.4	60.8	55.9
WV0028452_003	434.0	440.5	1156.0	906.8	640.0	599.5	192.8	193.1	153.8	136.3	178.4	199.8
WV0028452_004	42.7	45.5	62.4	42.3	54.4	67.6	21.4	12.3	6.4	12.9	15.3	18.3
WV0028452_005	81.2	81.2	230.0	8.0	5.5	81.2	81.2	81.2	81.2	81.2	81.2	81.2
WV0028452_006	4.0	3.7	6.4	3.5	1.5	10.5	2.3	3.3	2.8	3.0	2.3	2.5
WV1002074_026	8.0	0.5	2.7	4.6	11.0	2.8	0.1	4.1	4.1	4.1	4.1	0.1
WV1002074_028	8.0	1.0	6.0	5.0	5.0	5.0	5.0	5.0	5.0	5.0	5.0	5.0
WV1002074_029	0.1	0.0	0.7	0.1	0.2	0.1	0.0	0.2	0.2	0.2	0.2	0.0
WV1002074_030	0.5	2.7	2.5	2.7	2.7	5.0	2.7	2.7	2.7	2.7	2.7	2.7
WV1012487_153	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0
WV1012487_167	1.0	2.0	5.0	2.4	2.4	2.4	2.4	2.4	2.0	2.0	1.0	1.0
WV1014951_040	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0
WV1014951_037	2.5	2.5	2.5	3.0	2.0	2.5	2.5	2.5	2.5	2.5	2.5	2.5
WV1014951_039	31.5	61.0	2.0	31.5	31.5	31.5	31.5	31.5	31.5	31.5	31.5	31.5
WV1012487_115	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.1	0.0	0.0	0.0
WV1002368_002	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4
WV0057011_018	10.0	10.0	10.0	10.0	10.0	10.0	10.0	10.0	10.0	10.0	10.0	10.0
WV1012487_148	49.0	105.0	95.0	49.0	50.0	102.5	25.0	20.0	10.0	61.4	37.5	132.5
WV1012487_149	11.0	21.5	32.5	20.0	20.0	39.0	22.5	17.5	10.0	20.0	4.5	21.5
WV1012487_051	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
WV1012487_054	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5
WV1012487_039	20.0	20.0	20.0	20.0	20.0	20.0	20.0	20.0	20.0	20.0	20.0	20.0
WV1012487_050	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5
WV1012487_069	3.0	3.0	3.0	2.0	3.0	3.0	3.0	3.0	3.0	4.0	3.0	3.0
WV1012487_087	15.0	15.0	15.0	15.0	15.0	15.0	15.0	15.0	15.0	15.0	15.0	15.0
WV1012487_020	4.0	4.0	4.0	4.0	4.0	4.0	4.0	4.0	4.0	4.0	4.0	4.0
WV0092142_002	75.0	15.0	29.1	10.0	59.1	59.1	3.0	1.5	340.0	59.1	59.1	59.1
WV1012487_022	75.0	75.0	75.0	75.0	75.0	75.0	75.0	75.0	75.0	75.0	75.0	75.0
WV1012487_053	3.8	3.8	3.8	2.0	2.0	7.5	3.8	3.8	3.8	3.8	3.8	3.8
WV1012487_063	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
WV1012487_072	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0	2.0
WV0092142_001	3.0	2.5	2.3	2.0	1.0	3.0	2.5	1.5	5.0	2.5	1.0	2.5
WV1012487_067	2.0	2.0	2.0	1.0	2.0	2.0	2.0	2.0	2.0	3.0	2.0	2.0
WV1012487_071	6.7	6.7	6.7	2.0	6.0	12.0	6.7	6.7	6.7	6.7	6.7	6.7
WV1012487_049	30.0	30.0	30.0	30.0	30.0	30.0	30.0	30.0	30.0	30.0	30.0	30.0
WV1012487_070	3.0	3.0	3.0	3.0	4.0	2.0	3.0	3.0	3.0	3.0	3.0	3.0
WV1012487_014	16.7	16.7	16.7	16.7	16.7	16.7	32.5	7.5	10.0	16.7	16.7	16.7
WV1012487_059	7.7	7.7	7.7	5.0	8.0	10.0	7.7	7.7	7.7	7.7	7.7	7.7
WV1002368_004	12.0	17.0	20.8	13.0	15.8	16.0	6.5	15.7	15.7	15.7	15.5	21.5
WV1012487_058	2.5	7.5	15.0	7.5	18.0	7.0	7.0	7.0	7.0	1.0	1.0	3.5
WV1012487_083	6.5	9.0	8.8	7.0	2.0	1.0	6.0	5.0	4.0	5.8	7.3	5.8
WV1012487_150	46.3	66.3	89.2	48.8	40.0	37.5	35.0	31.3	45.0	80.0	50.5	53.3
WV1012487_078	3.5	4.3	8.0	6.8	6.5	7.5	5.0	4.8	1.0	2.0	2.5	3.5
WV1012487_076	7.0	3.3	4.8	2.3	1.5	2.3	9.3	10.0	5.0	2.0	1.0	2.0
WV1002066_001	27.2	25.2	18.2	17.8	16.2	20.5	7.7	5.9	9.8	10.5	10.1	14.9
WV1012487_035	44.2	74.2	53.3	42.8	200.0	60.0	12.7	0.0	8.0	15.9	7.7	22.5
WV1012487_005	31.5	97.5	96.7	43.0	40.8	57.5	4.0	12.0	21.5	21.0	25.2	22.0
WV1012487_005	31.5	97.5	96.7	43.0	40.8	57.5	4.0	12.0	21.5	21.0	25.2	22.0
WV1002066_002	27.3	35.3	30.6	30.0	33.8	43.5	27.2	14.6	15.3	10.9	24.5	31.1
WV1012487_024	26.4	49.4	79.4	39.8	91.7	30.9	13.3	15.0	8.8	20.0	18.0	23.4
WV1012487_030	43.5	45.8	78.8	47.5	97.5	42.4	16.7	30.0	11.8	16.3	6.8	19.0

Table A3. Monthly Reported Discharge Rate for Paint Creek NPDES Outlets.

There are various periods of record for these outlets.

NPDES Outlet	Jan (gpm)	Feb (gpm)	Mar (gpm)	Apr (gpm)	May (gpm)	Jun (gpm)	Jul (gpm)	Aug (gpm)	Sep (gpm)	Oct (gpm)	Nov (gpm)	Dec (gpm)
WV1012487_006	57.5	137.5	100.7	85.4	116.8	65.6	27.5	20.0	27.5	54.4	40.6	43.8
WV1002368_001	12.8	13.3	11.0	9.0	9.8	18.0	3.8	2.0	1.2	2.3	4.1	7.7
WV1012487_027	83.9	78.5	100.8	60.6	82.0	61.6	40.0	29.4	31.8	50.5	35.2	70.3
WV1009311_005	70.4	74.1	64.8	100.4	79.3	116.6	67.1	58.4	55.7	60.6	46.1	55.0
WV1009311_006	66.0	62.3	67.3	73.0	107.0	83.9	299.2	168.0	37.0	159.2	37.4	120.0

Table A4. Estimated Monthly Average Discharge Rate for Paint Creek NPDES Outlets without observed data.												
The monthly average discharge rate for these outlets were estimated by multiplying the surface drainage area for the outlet by the estimated local mine drainage rate. The local mine drainage rate (mine outlet drainage rate per unit drainage area) was assumed to be twice the local surface drainage rate, which was calculated from the monthly average flow rate and drainage area for a stream near the mine outlet.												
NPDES Outlet	Jan (gpm)	Feb (gpm)	Mar (gpm)	Apr (gpm)	May (gpm)	Jun (gpm)	Jul (gpm)	Aug (gpm)	Sep (gpm)	Oct (gpm)	Nov (gpm)	Dec (gpm)
WV0028452_007	3.4	3.7	4.4	3.3	2.4	1.5	1.1	1.0	0.7	1.2	1.7	1.5
WV0028452_008	0.8	0.8	1.0	0.7	0.5	0.3	0.2	0.2	0.2	0.3	0.4	0.3
WV0028452_009	0.8	0.8	1.0	0.7	0.5	0.3	0.2	0.2	0.2	0.3	0.4	0.3
WV0028452_012	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV0028452_013	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV0028452_014	3.8	4.1	4.9	3.7	2.7	1.6	1.2	1.1	0.8	1.3	1.9	1.7
WV0028452_015	3.8	4.1	4.9	3.7	2.7	1.6	1.2	1.1	0.8	1.3	1.9	1.7
WV0057011_003	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV0057011_006	3.2	3.4	4.1	3.1	2.3	1.4	1.0	0.9	0.6	1.1	1.6	1.4
WV0057011_007	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV0057011_008	10.3	11.2	13.4	10.0	7.4	4.4	3.2	2.9	2.1	3.6	5.2	4.6
WV0057011_009	22.2	24.1	28.8	21.5	15.8	9.6	6.9	6.3	4.5	7.7	11.2	9.9
WV0057011_010	6.0	6.5	7.7	5.8	4.2	2.6	1.9	1.7	1.2	2.1	3.0	2.6
WV0057011_011	6.4	6.9	8.2	6.1	4.5	2.7	2.0	1.8	1.3	2.2	3.2	2.8
WV0057011_012	26.2	28.4	33.9	25.3	18.7	11.3	8.2	7.5	5.3	9.0	13.2	11.6
WV0057011_013	13.1	14.2	17.0	12.7	9.3	5.6	4.1	3.7	2.6	4.5	6.6	11.6
WV0057011_014	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV0057011_015	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV0057011_016	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV0057011_017	4.8	5.2	6.2	4.6	3.4	2.1	1.5	1.4	1.0	1.6	2.4	2.1
WV0092142_003	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV0092142_004	4.1	4.5	5.4	4.0	2.9	1.8	1.3	1.2	0.8	1.4	2.1	1.8
WV0096679_001	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1001370_001	3.7	4.0	4.8	3.6	2.7	1.6	1.2	1.1	0.7	1.3	1.9	1.7
WV1002066_003	4.4	4.8	5.6	4.0	3.2	2.0	1.2	1.2	0.8	1.6	2.0	2.0
WV1002066_004	3.0	3.3	3.8	2.7	2.2	1.4	0.8	0.8	0.5	1.1	1.4	1.4
WV1002066_005	11.6	12.6	14.7	10.5	8.4	5.3	3.2	3.2	2.1	4.2	5.3	5.3
WV1002066_006	29.7	32.4	37.8	27.0	21.6	13.5	8.1	8.1	5.4	10.8	13.5	13.5
WV1002074_027	0.7	0.7	0.9	0.7	0.5	0.3	0.2	0.2	0.1	0.2	0.3	0.3
WV1002074_031	3.7	4.0	4.8	3.6	2.6	1.6	1.2	1.1	0.7	1.3	1.9	1.7
WV1002074_032	1.0	1.1	1.3	1.0	0.7	0.4	0.3	0.3	0.2	0.3	0.5	0.5
WV1002368_003	25.6	27.7	33.1	24.8	18.2	11.0	8.0	7.3	5.2	8.8	12.9	11.4
WV1002368_005	3.8	4.1	4.9	3.7	2.7	1.6	1.2	1.1	0.8	1.3	1.9	1.7
WV1009311_004	3.3	3.5	4.2	3.2	2.3	1.4	1.0	0.9	0.7	1.1	1.6	1.5
WV1009311_008	1.9	2.1	2.5	1.9	1.4	0.8	0.6	0.5	0.4	0.7	1.0	0.9
WV1009311_009	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1009311_010	1.9	2.1	2.5	1.9	1.4	0.8	0.6	0.5	0.4	0.7	1.0	0.9
WV1009311_011	7.3	7.9	9.5	7.1	5.2	3.2	2.3	2.1	1.5	2.5	3.7	3.2
WV1009311_012	3.1	3.3	4.0	3.0	2.2	1.3	1.0	0.9	0.6	1.1	1.5	1.4
WV1009311_013	0.8	0.8	1.0	0.7	0.5	0.3	0.2	0.2	0.2	0.3	0.4	0.3
WV1009311_014	0.8	0.8	1.0	0.7	0.5	0.3	0.2	0.2	0.2	0.3	0.4	0.3
WV1009311_015	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1009311_017	0.8	0.8	1.0	0.7	0.5	0.3	0.2	0.2	0.2	0.3	0.4	0.3
WV1009311_018	4.2	4.6	5.5	4.1	3.0	1.8	1.3	1.2	0.9	1.5	2.1	1.9
WV1009311_055	4.2	4.6	5.5	4.1	3.0	1.8	1.3	1.2	0.9	1.5	2.1	1.9

Table A4. Estimated Monthly Average Discharge Rate for Paint Creek NPDES Outlets without observed data.												
The monthly average discharge rate for these outlets were estimated by multiplying the surface drainage area for the outlet by the estimated local mine drainage rate. The local mine drainage rate (mine outlet drainage rate per unit drainage area) was assumed to be twice the local surface drainage rate, which was calculated from the monthly average flow rate and drainage area for a stream near the mine outlet.												
NPDES Outlet	Jan (gpm)	Feb (gpm)	Mar (gpm)	Apr (gpm)	May (gpm)	Jun (gpm)	Jul (gpm)	Aug (gpm)	Sep (gpm)	Oct (gpm)	Nov (gpm)	Dec (gpm)
WV1012487_003	2.1	2.3	2.7	2.0	1.5	0.9	0.7	0.6	0.4	0.7	1.1	1.0
WV1012487_004	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4
WV1012487_007	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_008	1.3	1.4	1.6	1.2	0.9	0.5	0.4	0.4	0.2	0.4	0.7	0.6
WV1012487_009	6.4	7.0	8.2	6.1	4.6	2.7	2.1	1.8	1.2	2.1	3.4	2.9
WV1012487_010	2.1	2.3	2.7	2.0	1.5	0.9	0.7	0.6	0.4	0.7	1.1	1.0
WV1012487_011	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_012	3.8	4.2	4.9	3.7	2.7	1.6	1.3	1.1	0.7	1.3	2.0	1.7
WV1012487_013	2.6	2.8	3.3	2.4	1.8	1.1	0.9	0.7	0.5	0.9	1.3	1.2
WV1012487_015	3.4	3.7	4.4	3.2	2.4	1.5	1.1	1.0	0.6	1.1	1.8	1.5
WV1012487_016	2.6	2.8	3.3	2.4	1.8	1.1	0.9	0.7	0.5	0.9	1.3	1.2
WV1012487_017	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_018	43.8	47.8	55.8	39.8	31.9	19.9	12.0	12.0	8.0	15.9	19.9	19.9
WV1012487_019	1.0	1.1	1.3	0.9	0.7	0.5	0.3	0.3	0.2	0.4	0.5	0.5
WV1012487_021	0.5	0.5	0.6	0.5	0.4	0.2	0.1	0.1	0.1	0.2	0.2	0.2
WV1012487_023	3.5	3.8	4.5	3.2	2.6	1.6	1.0	1.0	0.6	1.3	1.6	1.6
WV1012487_025	8.1	8.8	10.3	7.3	5.9	3.7	2.2	2.2	1.5	2.9	3.7	3.7
WV1012487_026	1.0	1.1	1.3	0.9	0.7	0.5	0.3	0.3	0.2	0.4	0.5	0.5
WV1012487_028	3.5	3.8	4.5	3.2	2.6	1.6	1.0	1.0	0.6	1.3	1.6	1.6
WV1012487_029	1.0	1.1	1.3	0.9	0.7	0.5	0.3	0.3	0.2	0.4	0.5	0.5
WV1012487_031	2.0	2.2	2.6	1.8	1.5	0.9	0.5	0.5	0.4	0.7	0.9	0.9
WV1012487_032	4.3	4.7	5.5	4.1	3.0	1.8	1.4	1.2	0.8	1.4	2.2	1.9
WV1012487_033	12.4	13.5	15.9	11.8	8.8	5.3	4.1	3.5	2.4	4.1	6.5	5.6
WV1012487_034	3.8	4.2	4.9	3.7	2.7	1.6	1.3	1.1	0.7	1.3	2.0	1.7
WV1012487_036	2.1	2.3	2.7	2.0	1.5	0.9	0.7	0.6	0.4	0.7	1.1	1.0
WV1012487_037	7.3	7.9	9.3	6.9	5.2	3.1	2.4	2.1	1.4	2.4	3.8	3.3
WV1012487_038	1.7	1.9	2.2	1.6	1.2	0.7	0.6	0.5	0.3	0.6	0.9	0.8
WV1012487_040	2.1	2.3	2.7	2.0	1.5	0.9	0.7	0.6	0.4	0.7	1.1	1.0
WV1012487_041	95.5	104.6	122.8	91.0	68.2	40.9	31.8	27.3	18.2	31.8	50.0	43.2
WV1012487_042	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_043	6.0	6.6	7.7	5.5	4.4	2.7	1.6	1.6	1.1	2.2	2.7	2.7
WV1012487_044	6.5	7.1	8.3	6.0	4.8	3.0	1.8	1.8	1.2	2.4	3.0	3.0
WV1012487_045	86.7	94.5	110.3	78.8	63.0	39.4	23.6	23.6	15.8	31.5	39.4	39.4
WV1012487_046	0.5	0.5	0.6	0.5	0.4	0.2	0.1	0.1	0.1	0.2	0.2	0.2
WV1012487_047	1.0	1.1	1.3	0.9	0.7	0.5	0.3	0.3	0.2	0.4	0.5	0.5
WV1012487_048	4.0	4.4	5.1	3.7	2.9	1.8	1.1	1.1	0.7	1.5	1.8	1.8
WV1012487_052	3.5	3.8	4.5	3.2	2.6	1.6	1.0	1.0	0.6	1.3	1.6	1.6
WV1012487_055	7.1	7.7	9.0	6.4	5.1	3.2	1.9	1.9	1.3	2.6	3.2	3.2
WV1012487_056	0.5	0.5	0.6	0.5	0.4	0.2	0.1	0.1	0.1	0.2	0.2	0.2
WV1012487_057	1.0	1.1	1.3	0.9	0.7	0.5	0.3	0.3	0.2	0.4	0.5	0.5
WV1012487_060	3.0	3.3	3.8	2.8	2.1	1.3	1.0	0.9	0.6	1.0	1.6	1.4
WV1012487_061	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_062	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_064	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_065	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4

Table A4. Estimated Monthly Average Discharge Rate for Paint Creek NPDES Outlets without observed data.												
The monthly average discharge rate for these outlets were estimated by multiplying the surface drainage area for the outlet by the estimated local mine drainage rate. The local mine drainage rate (mine outlet drainage rate per unit drainage area) was assumed to be twice the local surface drainage rate, which was calculated from the monthly average flow rate and drainage area for a stream near the mine outlet.												
NPDES Outlet	Jan (gpm)	Feb (gpm)	Mar (gpm)	Apr (gpm)	May (gpm)	Jun (gpm)	Jul (gpm)	Aug (gpm)	Sep (gpm)	Oct (gpm)	Nov (gpm)	Dec (gpm)
WV1012487_066	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4
WV1012487_068	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_073	9.8	10.7	12.6	9.3	7.0	4.2	3.3	2.8	1.9	3.3	5.1	4.4
WV1012487_074	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4
WV1012487_075	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_077	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4
WV1012487_079	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_080	3.4	3.7	4.4	3.2	2.4	1.5	1.1	1.0	0.6	1.1	1.8	1.5
WV1012487_081	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_082	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_084	9.0	9.8	11.5	8.5	6.4	3.8	3.0	2.6	1.7	3.0	4.7	4.1
WV1012487_085	2.1	2.3	2.7	2.0	1.5	0.9	0.7	0.6	0.4	0.7	1.1	1.0
WV1012487_086	2.1	2.3	2.7	2.0	1.5	0.9	0.7	0.6	0.4	0.7	1.1	1.0
WV1012487_088	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4
WV1012487_089	6.4	7.0	8.2	6.1	4.6	2.7	2.1	1.8	1.2	2.1	3.4	2.9
WV1012487_090	3.8	4.2	4.9	3.7	2.7	1.6	1.3	1.1	0.7	1.3	2.0	1.7
WV1012487_091	15.1	16.5	19.2	13.7	11.0	6.9	4.1	4.1	2.7	5.5	6.9	6.9
WV1012487_092	1.0	1.1	1.3	0.9	0.7	0.5	0.3	0.3	0.2	0.4	0.5	0.5
WV1012487_093	3.0	3.3	3.8	2.7	2.2	1.4	0.8	0.8	0.5	1.1	1.4	1.4
WV1012487_094	0.5	0.5	0.6	0.5	0.4	0.2	0.1	0.1	0.1	0.2	0.2	0.2
WV1012487_095	0.5	0.5	0.6	0.5	0.4	0.2	0.1	0.1	0.1	0.2	0.2	0.2
WV1012487_096	0.5	0.5	0.6	0.5	0.4	0.2	0.1	0.1	0.1	0.2	0.2	0.2
WV1012487_097	15.1	16.5	19.2	13.7	11.0	6.9	4.1	4.1	2.7	5.5	6.9	6.9
WV1012487_098	9.6	10.4	12.2	8.7	7.0	4.4	2.6	2.6	1.7	3.5	4.4	4.4
WV1012487_099	1.0	1.1	1.3	0.9	0.7	0.5	0.3	0.3	0.2	0.4	0.5	0.5
WV1012487_100	13.1	14.3	16.7	11.9	9.5	6.0	3.6	3.6	2.4	4.8	6.0	6.0
WV1012487_101	1.0	1.1	1.3	0.9	0.7	0.5	0.3	0.3	0.2	0.4	0.5	0.5
WV1012487_102	6.0	6.6	7.7	5.5	4.4	2.7	1.6	1.6	1.1	2.2	2.7	2.7
WV1012487_103	1.5	1.6	1.9	1.4	1.1	0.7	0.4	0.4	0.3	0.5	0.7	0.7
WV1012487_104	1.5	1.6	1.9	1.4	1.1	0.7	0.4	0.4	0.3	0.5	0.7	0.7
WV1012487_105	4.0	4.4	5.1	3.7	2.9	1.8	1.1	1.1	0.7	1.5	1.8	1.8
WV1012487_106	7.7	8.4	9.9	7.3	5.5	3.3	2.6	2.2	1.5	2.6	4.0	3.5
WV1012487_108	1.3	1.4	1.6	1.2	0.9	0.5	0.4	0.4	0.2	0.4	0.7	0.6
WV1012487_109	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4
WV1012487_110	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_111	1.3	1.4	1.6	1.2	0.9	0.5	0.4	0.4	0.2	0.4	0.7	0.6
WV1012487_112	1.5	1.6	1.9	1.4	1.1	0.7	0.4	0.4	0.3	0.5	0.7	0.7
WV1012487_113	2.0	2.2	2.6	1.8	1.5	0.9	0.5	0.5	0.4	0.7	0.9	0.9
WV1012487_114	0.5	0.5	0.6	0.5	0.4	0.2	0.1	0.1	0.1	0.2	0.2	0.2
WV1012487_145	2.5	2.7	3.2	2.3	1.8	1.1	0.7	0.7	0.5	0.9	1.1	1.1
WV1012487_146	2.5	2.7	3.2	2.3	1.8	1.1	0.7	0.7	0.5	0.9	1.1	1.1
WV1012487_147	203.5	222.0	259.1	185.0	148.0	92.5	55.5	55.5	37.0	74.0	92.5	92.5
WV1012487_151	2.1	2.3	2.7	2.0	1.5	0.9	0.7	0.6	0.4	0.7	1.1	1.0
WV1012487_152	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_154	1.3	1.4	1.6	1.2	0.9	0.5	0.4	0.4	0.2	0.4	0.7	0.6

Table A4. Estimated Monthly Average Discharge Rate for Paint Creek NPDES Outlets without observed data.												
The monthly average discharge rate for these outlets were estimated by multiplying the surface drainage area for the outlet by the estimated local mine drainage rate. The local mine drainage rate (mine outlet drainage rate per unit drainage area) was assumed to be twice the local surface drainage rate, which was calculated from the monthly average flow rate and drainage area for a stream near the mine outlet.												
NPDES Outlet	Jan (gpm)	Feb (gpm)	Mar (gpm)	Apr (gpm)	May (gpm)	Jun (gpm)	Jul (gpm)	Aug (gpm)	Sep (gpm)	Oct (gpm)	Nov (gpm)	Dec (gpm)
WV1012487_155	11.1	12.1	14.3	10.6	7.9	4.8	3.7	3.2	2.1	3.7	5.8	5.0
WV1012487_156	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4
WV1012487_157	7.7	8.4	9.9	7.3	5.5	3.3	2.6	2.2	1.5	2.6	4.0	3.5
WV1012487_158	7.7	8.4	9.9	7.3	5.5	3.3	2.6	2.2	1.5	2.6	4.0	3.5
WV1012487_159	2.1	2.3	2.7	2.0	1.5	0.9	0.7	0.6	0.4	0.7	1.1	1.0
WV1012487_160	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4
WV1012487_161	13.6	14.9	17.5	13.0	9.7	5.8	4.5	3.9	2.6	4.5	7.1	6.2
WV1012487_162	1.7	1.9	2.2	1.6	1.2	0.7	0.6	0.5	0.3	0.6	0.9	0.8
WV1012487_163	0.4	0.5	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1012487_164	7.3	7.9	9.3	6.9	5.2	3.1	2.4	2.1	1.4	2.4	3.8	3.3
WV1012487_165	2.6	2.8	3.3	2.4	1.8	1.1	0.9	0.7	0.5	0.9	1.3	1.2
WV1012487_166	0.9	0.9	1.1	0.8	0.6	0.4	0.3	0.2	0.2	0.3	0.4	0.4
WV1012487_168	9.1	9.9	11.5	8.2	6.6	4.1	2.5	2.5	1.6	3.3	4.1	4.1
WV1012487_169	6.5	7.1	8.3	6.0	4.8	3.0	1.8	1.8	1.2	2.4	3.0	3.0
WV1012592_011	28.4	30.5	36.7	27.0	20.1	12.5	9.0	8.3	5.5	9.7	13.9	12.5
WV1012592_012	40.0	42.9	51.7	38.0	28.3	17.6	12.7	11.7	7.8	13.7	19.5	17.6
WV1014951_003	9.9	10.8	12.9	9.6	7.1	4.3	3.1	2.8	2.0	3.4	5.0	4.4
WV1014951_004	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1014951_005	3.8	4.2	5.0	3.7	2.7	1.7	1.2	1.1	0.8	1.3	1.9	1.7
WV1014951_006	0.8	0.8	1.0	0.7	0.5	0.3	0.2	0.2	0.2	0.3	0.4	0.3
WV1014951_007	3.8	4.2	5.0	3.7	2.7	1.7	1.2	1.1	0.8	1.3	1.9	1.7
WV1014951_008	1.5	1.7	2.0	1.5	1.1	0.7	0.5	0.4	0.3	0.5	0.8	0.7
WV1014951_009	2.3	2.5	3.0	2.2	1.6	1.0	0.7	0.6	0.5	0.8	1.1	1.0
WV1014951_010	4.6	5.0	5.9	4.4	3.3	2.0	1.4	1.3	0.9	1.6	2.3	2.0
WV1014951_011	0.8	0.8	1.0	0.7	0.5	0.3	0.2	0.2	0.2	0.3	0.4	0.3
WV1014951_012	6.5	7.1	8.4	6.3	4.6	2.8	2.0	1.8	1.3	2.2	3.2	2.9
WV1014951_013	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1014951_014	2.3	2.5	3.0	2.2	1.6	1.0	0.7	0.6	0.5	0.8	1.1	1.0
WV1014951_031	3.4	3.7	4.4	3.3	2.4	1.5	1.1	1.0	0.7	1.2	1.7	1.5
WV1014951_032	3.6	3.9	4.6	3.4	2.5	1.5	1.1	1.0	0.7	1.2	1.8	1.6
WV1014951_033	3.4	3.7	4.4	3.3	2.4	1.5	1.1	1.0	0.7	1.2	1.7	1.5
WV1014951_035	6.9	7.5	8.9	6.6	4.9	2.9	2.1	1.9	1.4	2.4	3.4	3.1
WV1014951_036	0.4	0.4	0.5	0.4	0.3	0.2	0.1	0.1	0.1	0.1	0.2	0.2
WV1014951_038	3.4	3.7	4.4	3.3	2.4	1.5	1.1	1.0	0.7	1.2	1.7	1.5
WV1014951_041	1.9	2.1	2.5	1.8	1.4	0.8	0.6	0.5	0.4	0.7	1.0	0.8
WV1014951_042	3.4	3.7	4.5	3.3	2.5	1.5	1.1	1.0	0.7	1.2	1.7	1.5
WV1014951_043	98.8	107.4	127.4	95.2	70.2	42.2	30.8	27.9	20.0	33.6	49.4	44.0
WV1014951_045	4.2	4.6	5.5	4.1	3.0	1.8	1.3	1.2	0.9	1.5	2.1	1.9
WV1015257_002	68.0	74.2	88.1	64.9	47.9	29.4	21.6	18.6	13.9	23.2	34.0	30.2
WV1019317_001	1.0	1.1	1.3	1.0	0.7	0.4	0.3	0.3	0.2	0.3	0.5	0.5

Table A5. NPDES Net Acidity and Metal Concentrations for Base Conditions Model.

Water Quality Constituent	Concentration (mg/L)
Net Acidity*	0.14
Total Iron	3.20
Manganese	2.00
Total Aluminum	4.30

*Net Acidity = 0.14 corresponds for pH = 6.

APPENDIX B. Theoretical Basis of the *TAMD*L computer program

Introduction

The computer program *TAMD*L is designed to simulate those aspects of a watershed's stream water quality affected by acidic mine drainage. The current version of *TAMD*L simulates water temperature, net acidity, proton concentration (pH), ferrous iron, ferric iron, manganese, aluminum and dissolved oxygen. Water quality conditions are simulated by numerically solving the advection, dispersion, loading and reaction partial differential equation for each of these constituents. The effect of the advection and dispersion terms on the concentration and temperature of the stream at each model node are solved using the explicit MacCormack predictor – corrector finite difference method. The loading and reaction terms in the governing equations are solved using the fourth order Runge-Kutta method.

Because *TAMD*L uses net acidity to simulate the solution buffering, the governing partial differential equation is not used to calculate the proton concentration. Rather an empirical net acidity – pH model is employed to calculate the proton concentration each time step. The increase in proton concentration caused by the chemical reactions being simulated is translated into a corresponding increase in net acidity. The constituents simulated by the program are listed in Table B1.

Table B1. Water Quality Constituents Simulated by *TAMD*L.

Number	Water Quality Constituent	Symbol	Units
1	Stream Temperature	T	°C
2	Net Acidity Concentration	A	mg/l CaCO ₃
3	Proton Molar Concentration	[H ⁺]	M
4	Ferrous Iron Concentration	Fe ²⁺	mg/l
5	Ferric Iron Concentration	Fe ³⁺	mg/l
6	Manganese Concentration	Mn ²⁺	mg/l
7	Total Aluminum Concentration	Al ³⁺	mg/l
8	Dissolved Oxygen Concentration	O ₂	mg/l

Governing Equations

The following partial differential equation is solved by *TAMD*L for each water quality constituent, except for proton concentration.

$$\frac{\partial C_i}{\partial t} = m \frac{\partial^2 C_i}{\partial x^2} - V \frac{\partial C_i}{\partial x} + L_i + S_i \quad (1)$$

Where: C_i = Water quality constituent vector.
 x = Distance downstream.

V	=	Stream velocity.
L_i	=	Nodal source (sink) loading term.
S_i	=	Chemical and physical reaction source (sink) term(s).
i	=	Constituent index corresponding to rows of Table B1.
m	=	Hydrodynamic dispersion.

As was mentioned above, the governing equation for the proton concentration is the empirical algebraic net acidity – pH model.

$$pH = -\log_{10}(C_3) = c_1 \max(1, C_2^2)^{c_2 \text{ sign}(C_2)} \quad (2)$$

Where:	C_3	=	Proton Concentration, mole/L.
	C_2	=	Net Acidity, g/m ³ CaCO ₃ equivalent.
	$\text{sign}(C_2)$	=	-1, net alkaline conditions ($C_2 < 0$).
		=	0, neutral conditions ($C_2 = 0$).
		=	+1, net acidic conditions ($C_2 > 0$).
	c_1, c_2	=	Watershed specific empirical constants.

The watershed specific empirical constants in the net acidity – pH model are best determined by statistically fitting the model to the measured net acidity and pH data from samples collected in the watershed being simulated. If no data is available, Stiles and Fripp (2000) presented values for these constants obtained from data collected from several watersheds in West Virginia, Maryland and Pennsylvania.

The magnitude of the loading source (sink) term is specified by the user for each parameter and model node throughout the simulation. This loading is expressed by the user as a function of the simulation time. Normally acidic loading (i.e. the acidic drainage from a mine portal) is specified as a positive loading of net acidity. Alkaline treatment of stream water is specified as a negative loading of net acidity.

The loading of some important water quality parameters can be a function of the local runoff. The reduction of net acidity near a site containing alkaline materials would be a direct function of the local runoff. For this to be adequately modeled the loading must be calculated by either by a specially designed subroutine or an external program before the simulation.

The chemical and physical reactions modeled by *TAMDL* are listed below in Tables B2 and B3. The source (sinks) for each constituent are calculated by algebraically summing the consumption and production rates for each of the modeled chemical and physical reactions.

Table B2. Chemical Reactions Simulated by *TAMDL*.

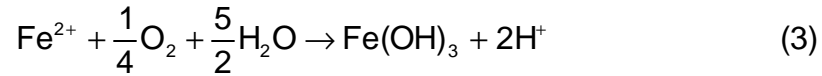
Chemical Reaction	Consumes	Produces
Ferrous Iron Oxidation	O ₂ , Fe ²⁺	Fe ³⁺ , H ⁺
Manganese Oxidation and Precipitation	O ₂ , Mn ²⁺	H ⁺
Aluminum Precipitation	Al ²⁺	H ⁺
Organic Material Decay	O ₂	–

Table B3. Physical Reactions Simulated by *TAMDL*.

Physical Reaction	Consumes	Produces
Aeration	–	O ₂
Meteorological Heating	–	T
Ferric Iron Sedimentation	Fe ³⁺	–

Ferrous Iron Oxidation

The computer program *TAMDL* uses the following chemical reaction to calculate the molar quantities of the constituents consumed and produced by ferrous iron oxidation. When the stream is effectively anoxic, $C_8 < 0.01 \text{ g/m}^3$, the oxidation of ferrous iron is neglected.



The following kinetic equation is used by the program for calculating the rate in which ferrous iron is being consumed both by abiotic and biotic oxidation. The rate in which dissolved oxygen is being consumed and the rate in which ferric iron and protons are being produced is determined from the stoichiometry presented in equation (3). This kinetic formulation was presented by Kirby, Thomas, Southam and Donald (1998).

$$R_4 = -3.125117192 \times 10^{-5} A_{\text{abiotic}} \exp\left(\frac{-E_{\text{abiotic}}}{R(C_1 + 273)}\right) C_4 C_8 C_3^{-2} - 3.125117192 \times 10^{-5} A_{\text{biotic}} \exp\left(\frac{-E_{\text{biotic}}}{R(C_1 + 273)}\right) C_4 C_{\text{ferro}} C_8 C_3 \quad (4)$$

Where:

C_4	=	Ferrous Iron Concentration, g/m ³ .
R_4	=	Ferrous Iron Oxidation Rate, g/m ³ /day.
C_8	=	Dissolved Oxygen Concentration, mg/L.
C_3	=	Proton Molar Concentration, M.
C_1	=	Stream Temperature, C.
A_{abiotic}	=	Empirical Chemical Rate Constant;
	=	$3.4560 \times 10^{10} \text{ mole/L/day}$.
E_{abiotic}	=	Empirical Chemical Rate Constant;

	=	96 kJ/mole.
R	=	Universal Gas Constant;
	=	8.314×10^{-3} kJ/mole/K.
C_{ferro}	=	<i>T. Ferrooxidans</i> dry mass concentration, g/m ³ .
A_{biotic}	=	Empirical Biological Rate Constant.
	=	8.8128×10^{13} mole/L/day.
E_{biotic}	=	Empirical Biological Rate Constant.
	=	58.77 kJ/mole.

Ferric Iron Sedimentation

TAMDL assumes that once the ferrous iron is oxidized to ferric iron, it immediately joins with dissolved oxygen to form ferric oxide. Ferric oxide leaves the computational domain by clinging to sediment particles in the stream and settling to the channel bottom. The re-suspension of the sediment particles with the attached ferric oxide is not simulated by *TAMDL*. The velocity that the sediment particles fall to the bottom of the channel is governed by Stokes Law. Stokes law requires that the particle Reynolds number (calculated with the particle diameter) be less than one (Roberson and Crowe, 1980). For the sediment particles most likely to carry ferric oxide, this assumption is quite reasonable. With the hydraulic depth, the settling velocity can be used to calculate the ferric iron sedimentation rate.

$$R_5 = -\frac{V_s C_5}{D} \quad (5)$$

Where:	R_5	=	Ferric iron sedimentation rate, g/m ³ /day.
	C_5	=	Ferric iron concentration, g/m ³ .
	V_s	=	Sediment settling velocity m/day.
	D	=	Hydraulic depth of the stream, m.
		=	Stream flow area / top width.

The following is the equation for the settling velocity that was derived from Stokes' Law.

$$V_s = \frac{4.0684 \times 10^9 (S_g - 1) D_s^2}{n} \quad (6)$$

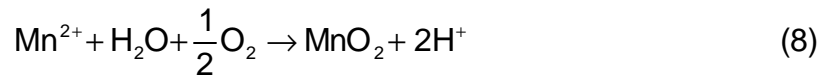
Where:	S_g	=	Specific weight of the sediment particles.
		=	Assumed by <i>TAMDL</i> to be 2.5.
	D_s	=	Diameter of the sediment particles, m.
	n	=	Kinematic viscosity of water, m ² /day.

Because the kinematic viscosity for water changes as the temperature changes, the following polynomial regression formula is used to calculate the kinematic viscosity from the stream temperature.

$$\begin{aligned}
n = & 1.25952 \times 10^{-10} (C_1 + 273)^6 - 2.226642 \times 10^{-7} (C_1 + 273)^5 \\
& + 0.0001639532 (C_1 + 273)^4 - 0.06436165 (C_1 + 273)^3 \\
& + 14.20698 (C_1 + 273)^2 - 1671.963 (C_1 + 273) + 81960.09
\end{aligned} \quad (7)$$

Manganese Oxidation and Precipitation

The formulation *TAMDL* employs to calculate the rate of manganese oxidation and precipitation was obtained from Stumm and Morgan (1981). Stoichiometry is defined by the program using the following chemical reaction. When the stream is effectively anoxic, $C_8 < 0.01 \text{ g/m}^3$, manganese oxidation and precipitation is neglected.



The following equation is an expression of the kinetic rate equation in terms of mass concentrations.

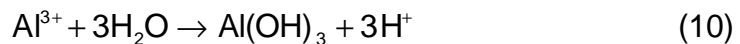
$$R_6 = \frac{-a_6 C_6 C_8 C_5}{C_3^2} \exp\left(\frac{-E_6}{R(C_1 + 273)}\right) \quad (9)$$

Where:

R_6	=	Manganese precipitation and oxidation rate, $\text{g/m}^3/\text{day}$.
C_6	=	Manganese concentration, g/m^3 .
C_8	=	Dissolved oxygen concentration, g/m^3 .
C_5	=	Ferric iron concentration, g/m^3 .
C_3	=	Proton concentration, mole/L.
a_6	=	User specified rate constant, $\text{L}^4/(\text{mg}^4\text{-day})$.
E_6	=	Empirical rate constant, 107.987 kJ/mole.
R	=	Universal gas constant, $8.314 \times 10^{-3} \text{ kJ/mole/K}$.

Aluminum Precipitation

The chemical reaction for aluminum precipitation is similar to the equation for manganese oxidation and precipitation except for the absence of oxidation because aluminum has only a single oxidation state. Likewise, the kinetic formulation for aluminum precipitation is similar to the kinetic formulation for abiotic ferrous oxidation.



$$R_7 = \frac{-a_7 C_7 A_7}{C_3^3} \exp\left(\frac{-E_7}{R(C_1 + 273)}\right) \quad (11)$$

Where: C_7 = Aluminum concentration, g/m³.
 R_7 = Aluminum precipitation rate, g/m³/day.
 a_7 = User specified empirical rate constant, dimensionless.
 E_7 = Empirical rate constant, 58.2 kJ/mole.
 A_7 = Empirical rate constant, 3160 mole³/L³/day.

If the user specifies a negative value for the empirical rate constant, a_7 , then *TAMD*L does not calculate the aluminum precipitation rate and restricts the aluminum concentration to the solubility limit for aluminum. The maximum aluminum concentration possible under equilibrium conditions, C_7^{equil} , is given by the following equation.

$$C_7^{equil} = \exp(35.071 + 6.9078 \log_{10}(C_3)) \quad (12)$$

Stream Reaeration and Organic Material Decay

The simulated dissolved oxygen concentration is affected by stream reaeration and the decay of organic material. Stream reaeration is assumed to be a function of the mean stream velocity, V , and the difference between the dissolved oxygen saturated concentration, C_{sat} , and the concentration, C_8 , O'Connor and Dobbins (1958). The affect of organic material decay on the dissolved oxygen concentration is simulated by *TAMD*L using a zeroth order sediment oxygen demand formulation adapted from the lake model *CE-QUAL-W2* (Cole and Buchak, 1995). The sediment oxygen demand per model node, k_{SOD} , is a constant specified by the user for all of the nodes in the entire computational domain. When the stream is nearly anoxic, $C_8 < 1.0$ g/m³, sediment oxygen demand is neglected. While the growth (decay) of algae is a common source (sink) of dissolved oxygen for streams, *TAMD*L assumes that algal growth in streams affected by AMD is negligible. Although dissolved oxygen can be a limiting factor in ferrous oxidation and manganese oxidation and precipitation, the precise simulation of dissolved oxygen is not crucial to the modeling of most streams affected by AMD. The following set of equations is the complete formulation the program uses to calculate sources (sinks) of dissolved oxygen.

$$R_8 = \frac{k_{SOD}}{R} + 1.023^{C_1 - 25} k_8 (C_{SAT} - C_8) \quad (13)$$

$$k_8 = 6.0804 \times 10^{-5} \frac{V}{D^{5/3}}, \text{ when } V < 42340 \text{ m/day}$$

$$k_8 = 0.013376 \frac{V^{1/2}}{D^{3/2}}, \text{ when } V \geq 42340 \text{ m/day} \quad (14)$$

Where:	R_8	=	Dissolved oxygen rate of production, g/day.
	C_8	=	Dissolved oxygen concentration, g/m ³ .
	k_{SOD}	=	Sediment oxygen demand per model node, g/m ³ /day.
	R	=	Hydraulic radius, m.
		=	Stream flow area / wetted channel perimeter.
	C_{SAT}	=	Saturated dissolved oxygen concentration, g/m ³ .
	k_8	=	O'Conner-Dobbins aeration rate constant, 1/day.
	D	=	Hydraulic depth of the stream, m.
		=	Stream flow area / top width.

The dissolved oxygen saturated concentration is calculated with the following regression formula derived from the saturated concentration values obtained from Tchobanoglous (1991).

$$C_{SAT} = -0.0005158(C_1 + 273)^3 + 0.04958(C_1 + 273)^2 - 15.94(C_1 + 273) + 1721 \quad (15)$$

Meteorological Heating

In order to account for the affect of weather on the stream temperature, *TAMDL* simulates the heating and cooling of the stream due to its contact with the atmosphere. Although the stream temperature has some affect on the rate of oxidation and precipitation of ferrous iron, manganese and aluminum, the precise simulation of stream temperature is not crucial to the modeling of streams affected by AMD. The simplified formulation used by the program assumes that the amount of heat transferred the atmosphere is proportional to the difference in temperature and the wind speed, and inversely proportional to the mean depth of flow across the channel.

$$R_1 = \frac{K_1 W (T_{AIR} - C_1)}{D} \quad (16)$$

Where:	R_1	=	Thermal energy gained (lost), calories/day.
	C_1	=	Stream temperature, C.
	W	=	Wind speed, m/day.
	T_{air}	=	Air Temperature, C.
	D	=	Hydraulic depth, m.
	K_1	=	Convective Heat Transfer Coefficient.
	»		1.0×10^{-5} calories/K

Stoichmetric Considerations

The aforementioned reactions must be modeled simultaneously because one water quality constituent can be produced by one reaction and consumed by

another reaction. Furthermore, a single reaction may consume and produce several constituents. Therefore, the elements of the source term in equation (1) are defined by the following stoichmetric relationships. If the user has specified that the aluminum precipitation be controlled by the aluminum equilibrium relationship, equation (12), the acidity contribution of aluminum precipitation is included in a separate portion of the model. The empirical net acidity – pH relationship is used to express increases in proton concentration in terms of increases in acidity.

$$S_1 = R_1 \quad (17)$$

$$S_2 = \frac{\text{sign}(C_2) \max(1, |C_2|) \exp(\ln(10) c_1 \max(1, C_2^2)^{c_2 \text{sign}(C_2)})}{2 \ln(10) c_1 c_2 \max(1, C_2^2)^{c_2 \text{sign}(C_2)}} S_2' \quad (18)$$

$$S_2' = \frac{2}{55847.0} R_4 + \frac{2}{54938.0} R_6 + \frac{3}{26981.54} R_7 \quad (19)$$

$$S_3 = 0 \quad (20)$$

$$S_4 = R_4 \quad (21)$$

$$S_5 = R_5 - R_4 \quad (22)$$

$$S_6 = R_6 \quad (23)$$

$$S_7 = R_7 \quad (24)$$

$$S_8 = R_8 + \left(\frac{1}{4} \right) \left(\frac{31.9988}{55.847} \right) R_4 + \left(\frac{1}{2} \right) \left(\frac{31.9988}{54.9380} \right) R_6 \quad (25)$$

Boundary and Initial Conditions

Upstream of the computational domain for each simulation, the user specifies the boundary temperature and concentrations. The specified upstream boundary temperature and concentrations may vary with simulation time. Normally, the upstream boundary condition is calculated from the results of the model for the upstream sub-watershed. If there is no upstream sub-watershed, the upstream boundary condition must be implied from the results of water quality sampling.

At the downstream end of each computational domain, *TAMD*L assumes that the spatial gradient of the temperature and concentration is zero. Downstream boundary condition required because of the dispersion (second derivative) term in governing equations. If there is no flow through the computational domain,

TAMDL automatically applies the downstream boundary condition to the upstream boundary. The concentrations specified for the upstream boundary are ignored.

TAMDL requires that the initial temperature and concentration be specified for each node. Initial conditions are not very important when one desires a steady state solution. When one is simulating a transient problem, the precise selection of initial conditions may have an important effect on the results calculated in the early portion of the simulation. Realistic initial conditions can be generated by simulating water quality conditions for a period prior to the desired simulation period.

Nodal Source Loading

The nodal source loading term, L_i , in equation (1), *TAMDL*'s governing partial differential equation is specified by the user for one or more of the simulation's model nodes. The fields, which must be specified in source loading section of the input data file, are listed in Table B4.

Table B4. Source Loading Data Fields.

Field	Description	Model Units
1	Model node to apply load.	—
2	Thermal load.	J/day
3	Net acidity load.	g/day CaCO ₃
4	Alkalinity production rate constant.	day ⁻¹
5	Ferrous iron load.	g/day
6	Ferric iron load.	g/day
7	Manganese load.	g/day
8	Total aluminum load.	g/day
9	Dissolved oxygen load.	g/day
10	Start Time for this load.	day
11	Ending time for this load.	day

Because the empirical net acidity – pH relationship, equation (2), is used by *TAMDL* is used to calculate the pH from the simulated stream net acidity, proton molar loading is not needed by the program. The alkalinity production constant is used to determine the alkalinity being generated at any given simulation time from either naturally occurring or placed limestone in the stream channel. Because the alkalinity being produced by the limestone is directly proportional to the net acidity of the stream, the resulting net acidity being applied to a particular model node is calculated with equation (26).

$$L_3 = L'_3 - L'_4(C_3 - C_3^{sat}) \quad (26)$$

Where: L_3 = Total net acidity application rate.

L'_3	=	Constant net acidity application rate.
L'_4	=	Alkalinity production rate constant.
C_3	=	Net acidity of the stream.
C_3^{sat}	=	Net acidity of solution saturated with limestone, mg/L CaCO_3 .

Experiments have shown that it is impossible under atmospheric conditions to raise the pH of a solution above 8.3 with the application of limestone. Therefore, the net acidity of a solution saturated with limestone is calculated with the empirical net acidity – pH relationship. Under certain circumstances, the application of equation (26) by *TAMDL* can result in numerical oscillation in the calculated net acidity concentration. Therefore, some caution should be observed in interpreting the results of a simulation of a stream with limestone in the channel at one or more model nodes.

Model Hydrology

The advection (first spatial derivative) term in equation (1), the governing partial differential equation, requires that the mean stream flow velocity be a known quantity at all locations in the model domain. Since the efficient numerical algorithm employed by *TAMDL* in solving the governing equation requires that both the velocity and the dispersion coefficient remain uniform throughout the model domain, the program requires that the watershed be divided into several sub-watersheds. Where two tributaries meet, the program applied continuity in order to calculate the upstream boundary conditions for the lower sub-watershed.

A TMDL study with unlimited funding would undoubtedly use a dedicated hydrologic software package to simulate the basin's water budget and calculate the mean stream flow velocity for each sub-watershed. Modifying *TAMDL* to read the results of a dedicated hydrologic package would not be difficult; however, few watersheds possess the quantity of hydrologic data required to warrant such a sophisticated analysis.

Because a sophisticated hydrologic analysis is rarely warranted, the current version of *TAMDL* assumes normal flow throughout the domain, unless the user, for testing purposes, has specified a constant depth of flow. Stream channels are also assumed to be prismatic trapezoids with a rectangular flood plain. Equation (27), Manning's equation is used by the program to iteratively calculate the mean velocity from the bottom slope and the hydraulic elements calculated with equations (28) and (29).

$$V = \frac{f}{n} \left(\frac{A}{P} \right)^{2/3} \sqrt{S_o} \quad (27)$$

$$A = (b + z \min(h, h_b)) \min(h, h_b) + b_f \max(0, h - h_b) \quad (28)$$

$$P = b + 2 \min(h, h_b) \sqrt{1 + z^2} + u(h - h_b)(b_f - b - 2(z + 1)h_b + 2h) \quad (29)$$

Where:	h	=	Depth of stream flow.
	h_b	=	Depth of the bank of the main trapezoidal channel.
	b	=	Bottom width of main trapezoidal channel.
	b_f	=	Width of rectangular flood plain.
	z	=	Inverse slope of the main trapezoidal channel sides.
	A	=	Total cross sectional area of the stream flow.
	P	=	Total wetted perimeter of the stream.
	n	=	Manning's roughness coefficient.
	S_o	=	Bottom slope of the main trapezoidal channel bottom.
	f	=	Manning unit conversion constant.
		=	1.00, if SI units are employed.
		=	1.49, if US customary units are employed.

In order for one value of Manning's roughness coefficient to be valid across a large range of discharge rates, the stream must be in the fully rough region of the turbulent regime. This condition is checked by the hydraulics subroutine in *TAMDL* with the criterion presented by Henderson (1966). When this criterion is not being satisfied, a warning message is printed in the log output file. Equation (30) is the non-dimensional form of the fully rough criterion.

$$\frac{g^{3.5} n^6}{f^6 n} \sqrt{RS} > 2.82 \times 10^{-4} \quad (30)$$

Where:	g	=	Acceleration due to gravity.
	n	=	Kinematic viscosity.

Because *TAMDL* uses SI units exclusively for hydraulic calculations, equation (30) can be further simplified to yield equation (31). Equation (31) requires that the temperature of the stream is approximately 20 °C.

$$n^6 \sqrt{RS} > 9.54 \times 10^{-14} \quad (31)$$

Selection of Numerical Algorithm

Because the *TAMDL* will normally be executed on an Intel based personal computer, efficient use of computational resources is very important. Therefore, the appropriate numerical algorithm should be both explicit and at least second order accurate in both time and space. One well tested algorithm that satisfies this requirement is the explicit MacCormack Predictor – Corrector method (Anderson, Tannehill and Pletcher, 1984). Because this finite difference

algorithm is normally applied to the solution of the advection – dispersion equation, the loading and chemical reaction terms in the governing equation must be solved analytically or with a numerical technique for first order ordinary differential equations.

Because the equations describing the kinetic rates of the aforementioned reactions are both complex and non-linear, it was decided that both the loading and reaction terms should be solved numerically. First order ordinary differential equations are commonly solved with one of the Runge-Kutta methods (Boyce and DiPrima, 1977). In order to make the simplify the *TAMDL* source code, it was decided that intermediate time steps to solve the chemical reaction and loading terms would not be employed. Therefore, to achieve the desirable accuracy, it was decided to use the fourth order Runge-Kutta method to solve the contributions of these terms to the temporal change in concentration.

Explicit MacCormack Predictor – Corrector Method

Because this method is explicit, *TAMDL* is not required to solve a matrix during each simulated time step. Instead, a predictor and corrector step are calculated every time step. The formulas for the predictor step are shown in equations (32) and (33).

$$F_{i,j}^n = V C_{i,j}^n \quad (32)$$

$$P_{i,j}^n = C_{i,j}^n - \frac{\Delta t}{\Delta x} (F_{i+1,j}^n - F_{i,j}^n) + \frac{m \Delta t}{(\Delta x)^2} (C_{i+1,j}^n - 2C_{i,j}^n + C_{i-1,j}^n) \quad (33)$$

Where:

$C_{i,j}^n$	=	Concentration of the j water quality parameter at the i model node at the n time step.
$P_{i,j}^n$	=	Predicted value of concentration for next time step.
$F_{i,j}^n$	=	Advection of j water quality parameter at the i model node at the n time step.
m	=	Hydrodynamic dispersion of constituents.
Δt	=	Size of the current time step.
Δx	=	Distance between model nodes.

The predicted concentration for the $n+1$ time step is used in equation (34) and (35) to calculate the corrected concentration for the $n+1$ time step. The net effect of the chemical reaction and loading terms on the concentration will be added to the corrected concentration.

$$G_{i,j}^n = V P_{i,j}^n \quad (34)$$

$$\hat{C}_{i,j}^{n+1} = \frac{1}{2} \left[C_{i,j}^n + P_{i,j}^n - \frac{\Delta t}{\Delta x} (G_{i,j}^n - G_{i-1,j}^n) + \frac{m\Delta t}{(\Delta x)^2} (P_{i+1,j}^n - 2P_{i,j}^n + P_{i-1,j}^n) \right] \quad (35)$$

Where: $G_{i,j}^n$ = Advection of j predicted water quality parameter at i model node at the n time step.
 $\hat{C}_{i,j}^{n+1}$ = Corrected concentration for the next time step.

Upstream and downstream boundary conditions are applied by changing the predictor and corrector formulas for the first and last model nodes to reflect the appropriate conditions.

Fourth Order Runge-Kutta Method

The fourth order Runge-Kutta method was selected because of both its familiarity and low truncation error. The small magnitude of truncation error, $O(\Delta t^5)$, is important because the time step is best controlled by the Explicit MacCormack method's stability criteria and *TAMDL* does not use intermediate time steps in evaluating the effects of the reaction and loading terms. The program adds the contribution of the loading and chemical reaction terms to the corrected concentration for the next time step using the equation (36). The weighted estimated average of the loading and kinetic rates during the time step are calculated by equations (37), (38), (39) and (40). Summation notation is not employed in these equations, and the parenthesis is used to indicate functional relationship, not multiplication.

$$C_{i,j}^{n+1} = \hat{C}_{i,j}^{n+1} + \frac{\Delta t}{6} (H_{i,j} + 2I_{i,j} + 2J_{i,j} + K_{i,j}) \quad (36)$$

$$H_{i,j} = S_{i,j}(\hat{C}_{i,j}^{n+1}) \quad (37)$$

$$I_{i,j} = S_{i,j} \left(\hat{C}_{i,j}^{n+1} + \frac{\Delta t}{2} H_{i,j} \right) \quad (38)$$

$$J_{i,j} = S_{i,j} \left(\hat{C}_{i,j}^{n+1} + \frac{\Delta t}{2} I_{i,j} \right) \quad (39)$$

$$K_{i,j} = S_{i,j} (\hat{C}_{i,j}^{n+1} + \Delta t J_{i,j}) \quad (40)$$

Where: $S_{ij}(C_{ij})$ = Net loading and kinetic reaction term for water quality constituent j at model node i given the concentration of the constituents specified by the array C_{ij} .

Time Step Calculation

Because the MacCormack Predictor – Corrector method is explicit, the size of the time step controls the stability of the advective – dispersion calculations. To a lesser extent, this is also true for the simulation of the net loading and kinetic reaction terms. Unfortunately, a simple stability criterion for the MacCormack algorithm does not exist when the method is applied to the advection – dispersion equation. However, *TAMDL* uses the minimum of three criteria suggested by Anderson, Tannehill and Pletcher (1984). Because the mean stream velocity, V , may change each time step, the inequality (41) is evaluated at the beginning of each time step.

$$\Delta t \leq \min \left(\frac{(\Delta x)^2}{2m}, \frac{\Delta x}{|V|}, \frac{(\Delta x)^2}{2m + |V|\Delta x} \right) \quad (41)$$

In order to give the user full control over the simulation, the user can specify a maximum time step for the program to use. However, the quality of the simulation can be degraded if a time step much smaller than that specified by inequality (41) is employed. Like the MacCormack method, no simple stability criterion for the Runge-Kutta method exists. However, *TAMDL* uses the following heuristic formula to ensure that it does not calculate negative concentrations of ferrous iron, ferric iron, manganese, total aluminum or dissolved oxygen.

$$\Delta t \leq \min \left(-\frac{C_{i,j}^n}{\max(S_{i,j}(C_{i,j}^n), 0)} \right) \forall i, j \text{ where } j = 4, 5, 6, 7 \text{ or } 8 \quad (42)$$

Because this maximum allowable time step is a function of the concentrations of some of the water quality constituents, inequality (42) is also evaluated each time step. Inequality (42) is not evaluated for temperature or net acidity because negative values of these quantities do not automatically invalidate the results of the simulation.

Conclusion

The computer program *TAMDL* is a powerful tool in simulating the one dimensional water quality of streams affected by acidic mine drainage and by the remediation of acidic mine drainage. This computer program simulates the transport and reaction of temperature, net acidity, pH, ferrous iron, ferric iron, manganese, aluminum and dissolved oxygen. With the exception of pH, the one dimensional advection – dispersion – loading – reaction partial differential equation is solved for these constituents. The pH is simulated with an empirical net acidity – pH relationship.

The advection terms in the governing equation involve the mean flow velocity, which is iteratively calculated with Manning's equation. Because the selected

numerical algorithm requires that the velocity be constant throughout the model domain, the watershed is divided into several sub-watersheds of nearly constant hydrologic properties.

The contribution of the advection – dispersion terms in the governing equation are solved with the explicit MacCormack Predictor – Corrector finite difference method. Whereas the contribution of the loading – reaction terms are calculated with the fourth order Runge – Kutta method. Because the size of the time step is important for the stability of both of these algorithms, the time step size is calculated at the beginning of each time step.

Formal List of *TAMDL* Modeling Assumptions

1. *TAMDL* assumes that the transport, reaction and loading is governed by the one dimensional partial differential equation given in equation (1).
2. The program assumes that the magnitude of the hydrodynamic dispersion for all of the simulated water quality constituents, except for proton concentration (pH), is the same.
3. The program assumes that the water quality of streams affected by acidic mine drainage are controlled by the temperature, net acidity, pH and the concentration of ferrous iron, ferric iron, manganese, total aluminum and dissolved oxygen.
4. The program assumes that the stream pH is adequately simulated by the empirical net acidity – pH relationship presented in equation (2).
5. The program assumes that all of the ferrous iron that is oxidizes immediately forms ferric hydroxide as shown in chemical reaction (3).
6. The program assumes that the oxidation of ferrous iron is irreversible.
7. The program assumes that the sedimentation of ferric iron is controlled by the settling of stream sediment.
8. The program assumes that the settling of stream sediment is controlled by Stokes law.
9. The program assumes that the size distribution of stream sediment is adequately represented by the mean sediment size.
10. The program assumes that the specific density of the stream sediment is equal to 2.5.

11. The program assumes that settled ferric iron sediment cannot be re-suspended by the stream.
12. The program assumes that the oxidation and precipitation of manganese is irreversible.
13. The program assumes that the precipitation of aluminum is irreversible.
14. The program assumes that the stream discharge may be considered uniform within a particular sub-watershed.
15. The program assumes that the hydraulics of the stream can be characterized by a mean values for the manning roughness coefficient, bottom slope, bottom width, side slope, channel height and flood plain width, within a particular sub-watershed.
16. The program assumes that the hydraulics of the stream are nearly uniform, i.e. can be approximated with a uniform flow equation, and can be adequately modeled with Manning's equation.
17. The program assumes that the magnitude of time step appropriate for simulating the in-stream chemical and physical processes is comparable to the magnitude of time step appropriate for simulating the transport of the modeled water quality constituents.

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Table C5. List of stream segments requiring reduction in AMD sources according to base conditions model.

Sub-WS	Listed for pH?	Reduction required in acidity?	Listed for iron?	Reduction required in iron?	Listed for Mn?	Reduction required in Mn?	Listed for Al?	Reduction required in Al?	Sub-WS	Not listed but requiring reduction in acidity?	Not listed but requiring reduction in iron?	Not listed but requiring reduction in Mn?
1	No	No	No	No	No	No	Yes	Yes	1	No	No	No
2	No	No	No	No	No	No	No	Yes	2	No	No	No
3	No	No	No	No	No	No	Yes	Yes	3	No	No	No
4	No	Yes	No	No	No	No	No	Yes	4	Yes	No	No
5	No	No	No	No	No	No	Yes	Yes	5	No	No	No
6	No	Yes	No	No	No	No	No	Yes	6	Yes	No	No
7	No	No	No	No	No	No	Yes	Yes	7	No	No	No
8	No	Yes	No	No	No	No	No	Yes	8	Yes	No	No
9	Yes	No	No	No	No	No	Yes	Yes	9	No	No	No
10	Yes	Yes	Yes	No	Yes	Yes	Yes	Yes	10	No	No	No
11	Yes	Yes	Yes	Yes	Yes	Yes	Yes	Yes	11	No	No	No
12	No	Yes	No	Yes	No	Yes	No	Yes	12	Yes	Yes	Yes
13	Yes	Yes	Yes	Yes	Yes	Yes	Yes	Yes	13	No	No	No
14	Yes	No	No	Yes	No	No	Yes	Yes	14	No	Yes	No
15	No	Yes	No	Yes	No	Yes	No	Yes	15	Yes	Yes	Yes
16	Yes	No	No	Yes	No	No	Yes	Yes	16	No	Yes	No
17	No	Yes	No	No	No	No	No	Yes	17	Yes	No	No
18	Yes	No	No	Yes	No	No	Yes	Yes	18	No	Yes	No
19	Yes	Yes	Yes	No	Yes	Yes	Yes	Yes	19	No	No	No
20	Yes	No	No	Yes	No	No	Yes	Yes	20	No	Yes	No
21	Yes	Yes	No	Yes	No	Yes	No	Yes	21	No	Yes	Yes
22	Yes	No	No	Yes	No	No	Yes	Yes	22	No	Yes	No
23	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	23	Yes	No	No
24	No	No	No	Yes	No	No	No	Yes	24	No	Yes	No
25	Yes	Yes	No	No	No	Yes	No	Yes	25	No	No	Yes
26	No	No	No	Yes	No	No	No	Yes	26	No	Yes	No
27	No	No	Yes	Yes	Yes	Yes	Yes	Yes	27	No	No	No
28	No	No	No	Yes	No	No	No	Yes	28	No	Yes	No
29	No	No	No	No	No	No	No	No	29	No	No	No
30	No	No	No	Yes	No	No	No	Yes	30	No	Yes	No
31	No	No	No	Yes	No	Yes	No	Yes	31	No	Yes	Yes

Table C5. List of stream segments requiring reduction in AMD sources according to base conditions model.

Sub-WS	Listed for pH?	Reduction required in acidity?	Listed for iron?	Reduction required in iron?	Listed for Mn?	Reduction required in Mn?	Listed for Al?	Reduction required in Al?	Sub-WS	Not listed but requiring reduction in acidity?	Not listed but requiring reduction in iron?	Not listed but requiring reduction in Mn?
32	No	No	No	Yes	No	No	No	Yes	32	No	Yes	No
33	Yes	Yes	Yes	No	Yes	No	Yes	No	33	No	No	No
34	No	No	No	Yes	No	No	No	Yes	34	No	Yes	No
35	No	Yes	No	No	No	No	No	Yes	35	Yes	No	No
36	No	No	No	Yes	No	No	No	Yes	36	No	Yes	No
37	No	No	No	Yes	No	Yes	No	Yes	37	No	Yes	Yes
38	No	No	No	No	No	No	No	Yes	38	No	No	No
39	No	No	No	Yes	No	Yes	No	Yes	39	No	Yes	Yes
40	No	No	No	Yes	No	No	No	Yes	40	No	Yes	No
41	No	No	No	Yes	No	No	No	No	41	No	Yes	No
42	No	No	No	Yes	No	No	No	Yes	42	No	Yes	No
43	No	Yes	No	Yes	No	No	No	Yes	43	Yes	Yes	No
44	No	No	No	Yes	No	No	No	Yes	44	No	Yes	No
45	No	No	No	Yes	No	Yes	No	Yes	45	No	Yes	Yes
46	No	No	No	Yes	No	No	No	Yes	46	No	Yes	No
47	No	No	Yes	Yes	Yes	No	Yes	Yes	47	No	No	No
48	No	No	No	Yes	No	Yes	No	Yes	48	No	Yes	Yes
49	No	No	No	Yes	No	Yes	No	Yes	49	No	Yes	Yes
50	No	No	No	Yes	No	Yes	No	Yes	50	No	Yes	Yes
51	No	No	No	Yes	No	Yes	No	Yes	51	No	Yes	Yes
52	No	No	No	Yes	No	Yes	No	Yes	52	No	Yes	Yes
53	No	Yes	No	Yes	No	Yes	No	Yes	53	Yes	Yes	Yes
54	No	No	No	Yes	No	Yes	No	Yes	54	No	Yes	Yes
55	No	No	No	Yes	No	Yes	No	Yes	55	No	Yes	Yes
56	No	No	No	Yes	No	No	No	Yes	56	No	Yes	No
57	No	No	No	Yes	No	Yes	No	Yes	57	No	Yes	Yes
58	No	No	No	Yes	No	No	No	Yes	58	No	Yes	No
59	No	No	No	Yes	No	No	No	Yes	59	No	Yes	No
60	No	No	No	Yes	No	No	No	Yes	60	No	Yes	No
61	No	No	No	Yes	No	No	No	Yes	61	No	Yes	No
62	No	No	No	Yes	No	No	No	Yes	62	No	Yes	No

Table C5. List of stream segments requiring reduction in AMD sources according to base conditions model.

Sub-WS	Listed for pH?	Reduction required in acidity?	Listed for iron?	Reduction required in iron?	Listed for Mn?	Reduction required in Mn?	Listed for Al?	Reduction required in Al?	Sub-WS	Not listed but requiring reduction in acidity?	Not listed but requiring reduction in iron?	Not listed but requiring reduction in Mn?
overall	21%	27%	13%	73%	13%	37%	29%	95%	overall	16%	65%	27%

Not listed but requiring reduction in AI?
No
Yes
No
Yes
No
Yes
No
Yes
No
No
No
Yes
No
No
Yes
No
Yes
No
No
No
Yes
No
No
Yes
Yes
Yes
No
Yes
No
Yes
Yes

Not listed but requiring reduction in AI?
68%

Appendix D: Results of the Paint Creek TAMDL Model

Because many of the Paint Creek sub-watersheds had little calibration data, this appendix will concentrate on those relatively few sub-watersheds from which most of the water quality samples were collected. Figures D1, D2, D3 and D4 are time series plots for WVDEP-SRG sample collection station PC054 where most of the Long Branch samples were collected. With a few exceptions, the observed readings appear to be matched by the model. The Long Branch and Ten Mile Fork sub-watersheds were, by far, the most difficult Paint Creek sub-watersheds to calibrate because of the action of the limestone sand treatment facilities installed by WVDEP-AML.

Figures D5, D6, D7 and D8 are time series plots for WVDEP-SRG sample collection station PC060, which is located near the center of the lower Ten Mile Fork sub-watershed. The sudden drop in simulated pH shown by Figure D5 is the result of a very low average pH reading reported by one of the mines in the sub-watershed. The increase in pH during water year 1998 is the result of the installation of the limestone sand treatment facilities upstream of the sample collection station. Additional calibration data would have permitted a better delineation of the decrease in acidity.

Figures D9, D10, D11 and D12 are time series plots for WVDEP-SRG sample collection station PC073, which is located near the mouth of Cedar Creek. Along with Ten Mile Fork and Long Branch, Cedar Creek is responsible for a major part of the watershed's discharge from abandoned mines. The observed readings show a significant increase in acidity and metals loading during water year 1999. Without additional calibration and AML seep data for this sub-watershed it is impossible to determine precisely the cause of the decline in water quality.

Table D1 and D2 contain the basic model parameters for each of the model's sub-watersheds. Excluded from the list are the nodal sources required for model calibration. The bottom channel width and Manning's roughness coefficient for each sub-watershed were calculated using the Manning equation from the flow area and wetted perimeter observed by WVDEP-SRG. WVDEP-SRG personal measured stream flow along with collecting water quality samples under high, medium and low flow conditions.

While some of the assigned values for the Manning's roughness coefficient may be greater than the design values given in Table 5-6 of (Chow, V.T. *Open Channel Hydraulics* 1959, McGraw-Hill Book Company, Inc., New York, NY), the values given by Chow (1959) were intended to be used with the Manning equation for reasonably straight channels with uniform or gradually varied flow. The Paint Creek tributaries are generally not straight and have a wide variation in both channel slope and cross sectional dimensions. Because TAMDL is simulating the hydraulics of non-prismatic channels under gradually or rapidly varied flow

conditions with the Manning uniform flow equation, the presence of some unusually high values for Manning's n is not unexpected. This result could have been avoided, if there was enough detailed hydraulic and geometric stream data to warrant the use of a more sophisticated hydraulic model.

Experimental TAMDL simulations were also executed to ensure that these values yielded reasonable results with the stream bottom slopes obtained from the DEM in WCMS. Only the hydraulic parameters for Cedar Creek, sub-watershed #21, required adjustment to obtain reasonable results. These adjustments were necessary given that the original hydraulic parameters yielded depths less than 5 cm during low flow conditions. Extremely shallow depths can cause numerical instabilities during TAMDL simulations because the rates of some of the modeled physical and chemical processes are inversely proportional to the hydraulic depth.

Non-point source loading for calibration and baseline models.

The upstream boundary conditions for the downstream sub-watersheds are defined by the model results calculated for those sub-watersheds immediately upstream. Sub-watersheds that have no other sub-watersheds above them, tributary sub-watersheds, have upstream boundary conditions defined by an upstream concentration. The total load entering a tributary sub-watershed is the sum of the upstream concentration multiplied by the stream's discharge and the loads entering the sub-watershed's model nodes.

In the Paint Creek TAMDL models, non-point sources can be divided into three classes: a source with a constant loading rate throughout the simulation, a source with a loading rate that varies during the simulation, or a source with a constant loading rate that is started during the simulation period. Seeps that drain abandoned underground mines were assumed to have a constant load throughout the simulation. Minor streams that entered the sub-watershed's stream segment were assigned a loading rate that varied during the simulation. In-stream AMD treatment facilities that were installed during the simulation period were assigned a constant alkaline (negative net acidic) loading rate that started when the facility was installed. The Paint Creek TAMDL model uses a datum date and time of 2400 hours, 31 December 1989.

Point sources entering the calibration model.

These loads were calculated from the reported monthly average concentrations and flow rates reported by the mine operators on the DMR submitted to WVDEP. All of the known DMR data was used to estimate the monthly mean flow rate for all of the currently discharging mine outlets. Because it is quite common for mine outlets to be permitted well in advance of the time the outlet begins to release mine drainage, only reported DMR data were included in the calibration model's

point sources. This approach is vulnerable to errors caused by incomplete DMR records, but it is the most accurate representation of reality.

Point sources for the models.

The point sources for the baseline conditions model were calculated using the reported or estimated monthly mean flow rate for each outlet. Upstream concentrations for the tributary sub-watersheds in the allocation model were reduced during the load allocation. The non-point model node sources for the allocation model. Like the concentrations in Table D7, the loads in Table D8 were reduced during the load allocation.

The wasteload allocation was conducted by lowering the effluent concentrations of the permitted mine outlets until the model results indicated that the TMDL endpoints were being maintained during the simulation. The reduced point source loading rates that resulted from the lowering of the effluent concentrations are listed in Table D9.

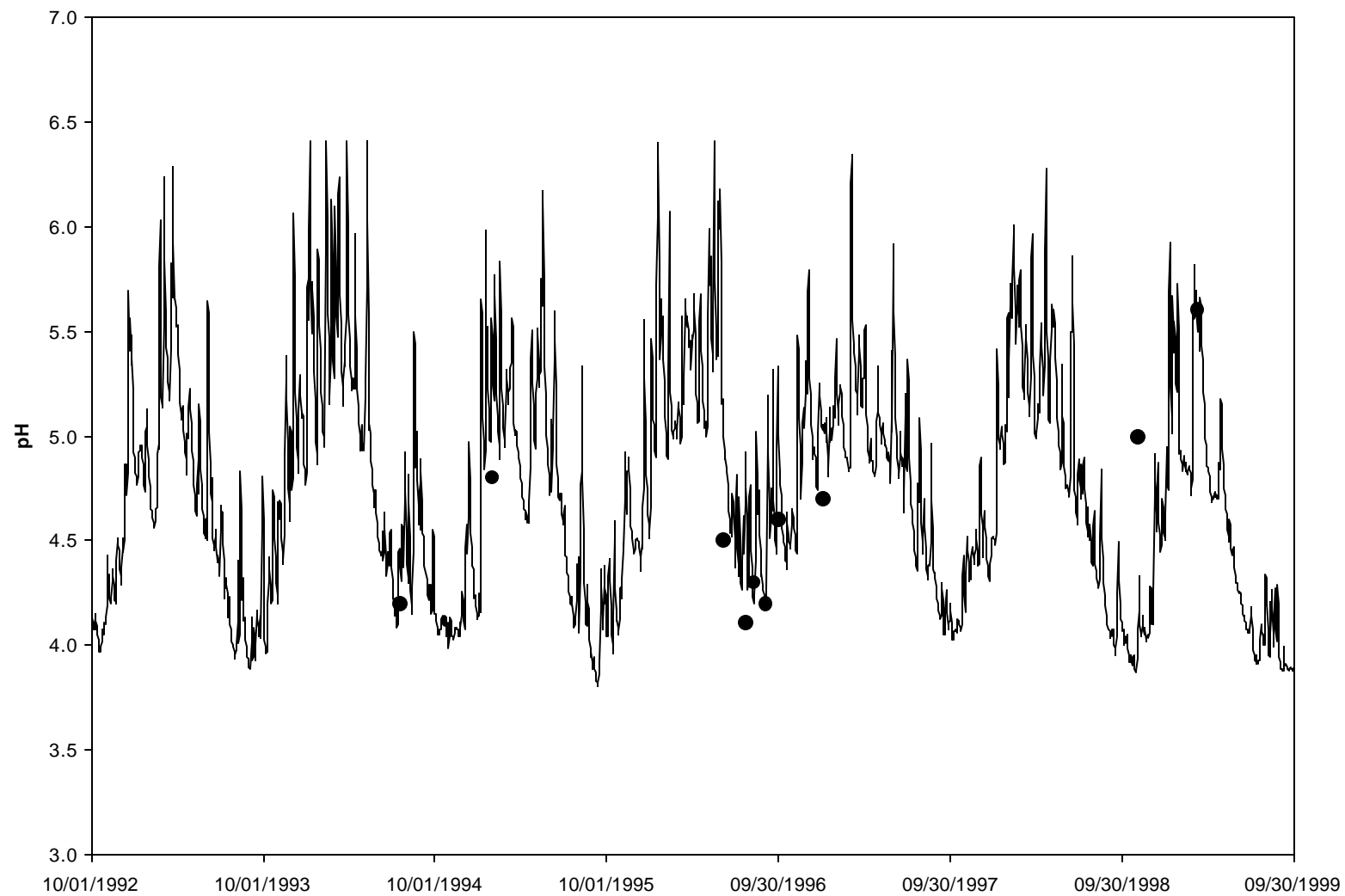


Figure D1. Time Series Plot of Observed and Calculated pH for WVDEP-SRG Station PC054.

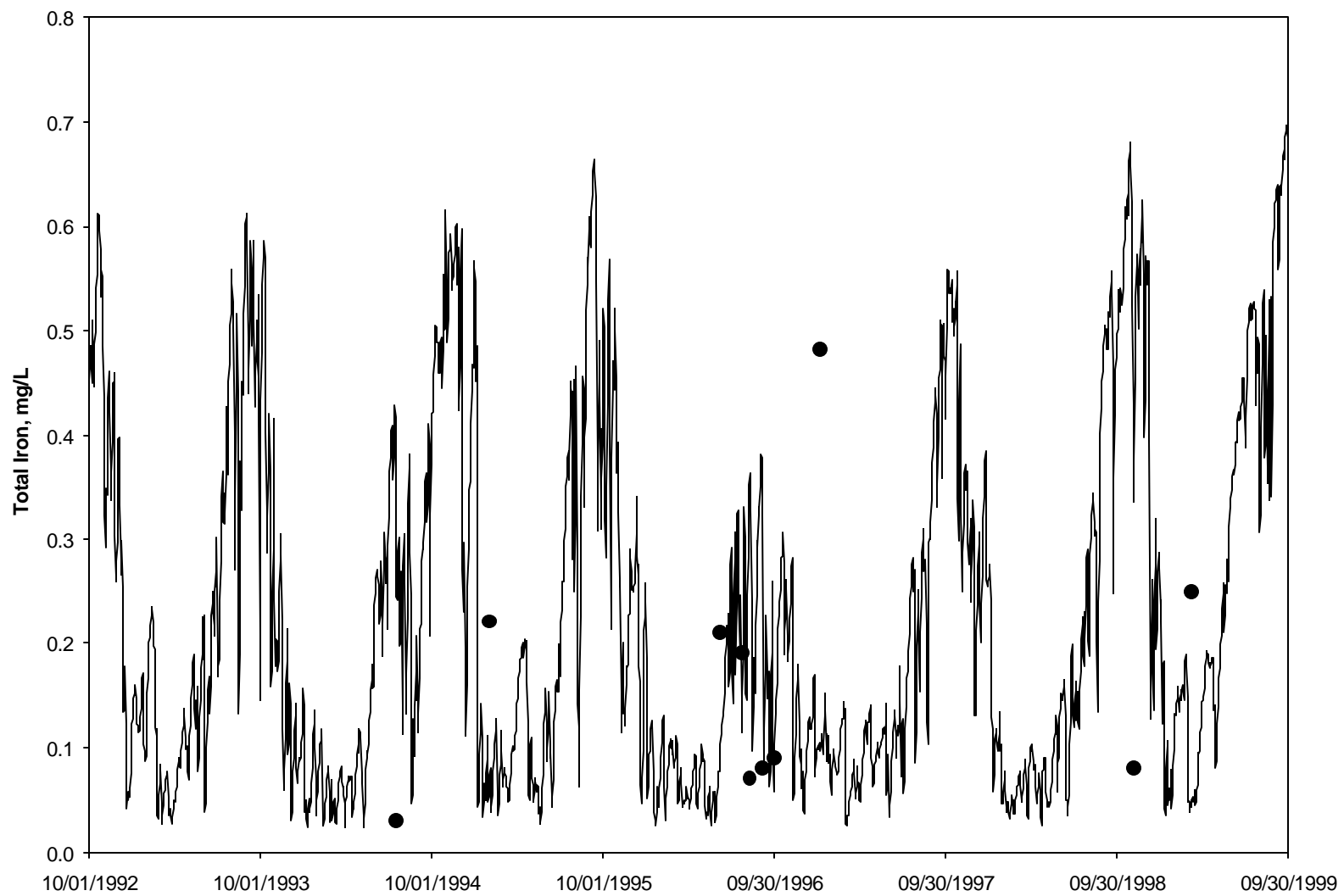


Figure D2. Time Series Plot of Observed and Calculated Total Iron for WVDEP-SRG Station PC054.

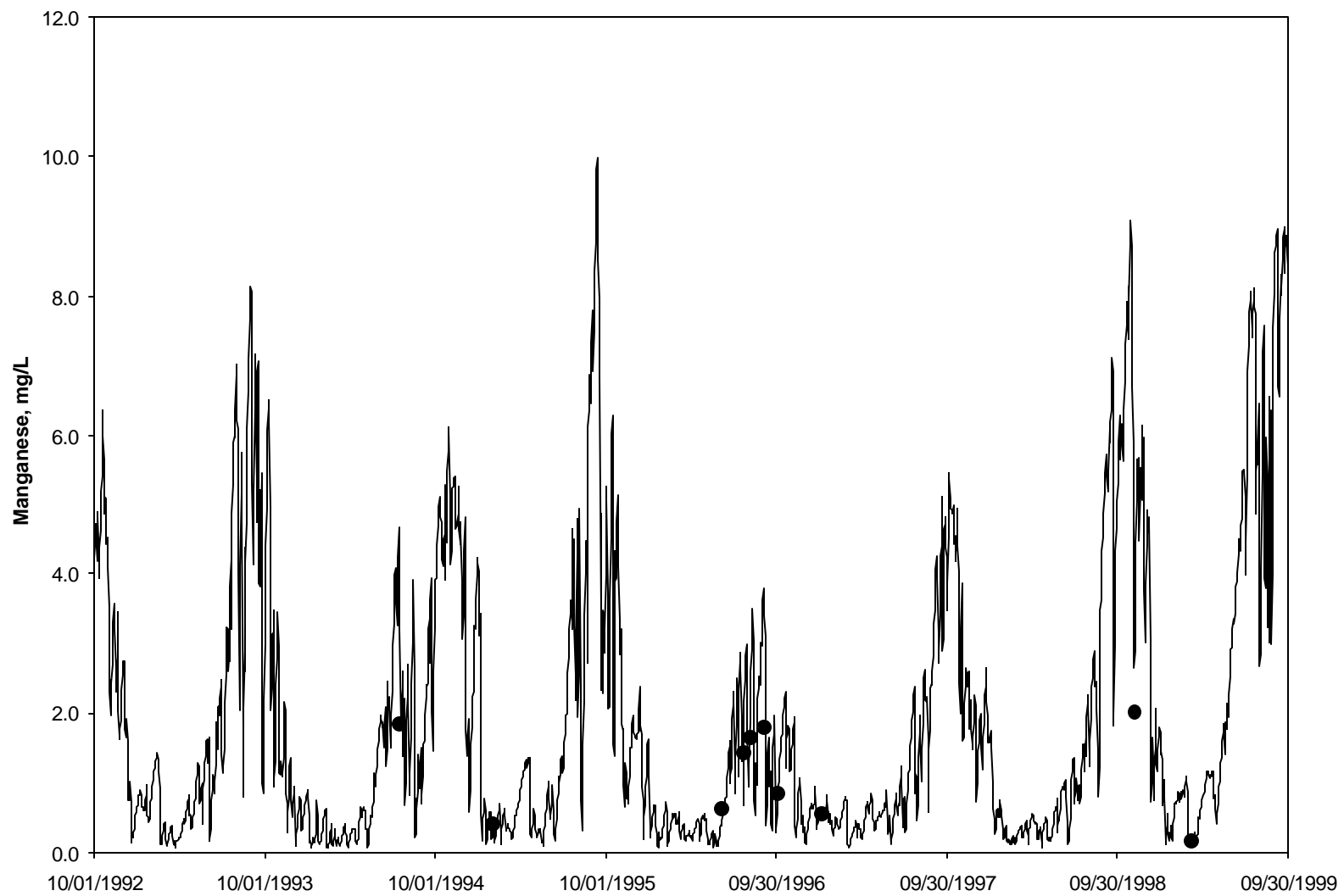


Figure D3. Time Series Plot of Observed and Calculated Manganese for WVDEP-SRG Station PC054.

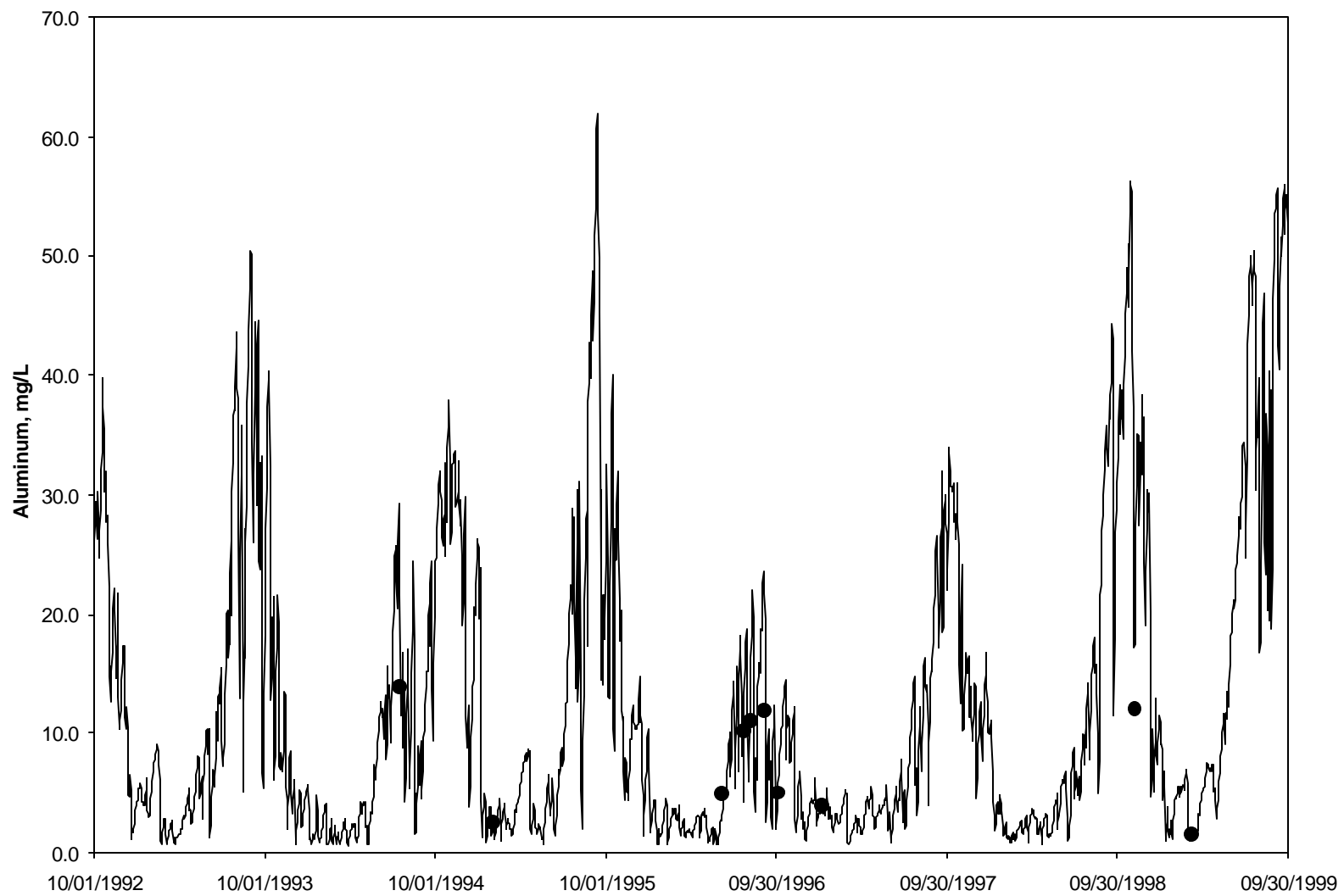


Figure D4. Time Series Plot of Observed and Calculated Aluminum for WVDEP-SRG Station PC054.

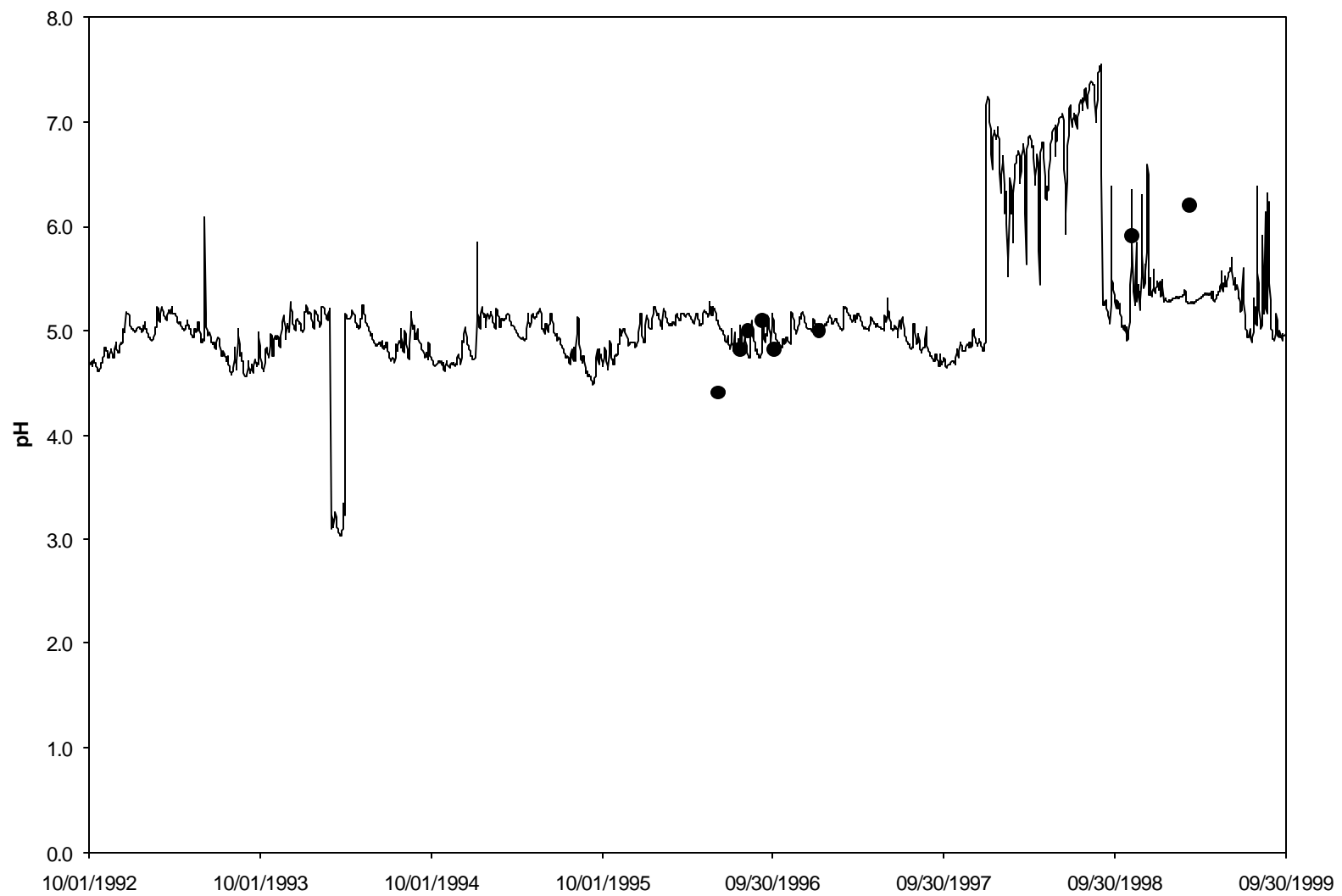


Figure D5. Time Series Plot of Observed and Calculated pH for WVDEP-SRG Station PC060.

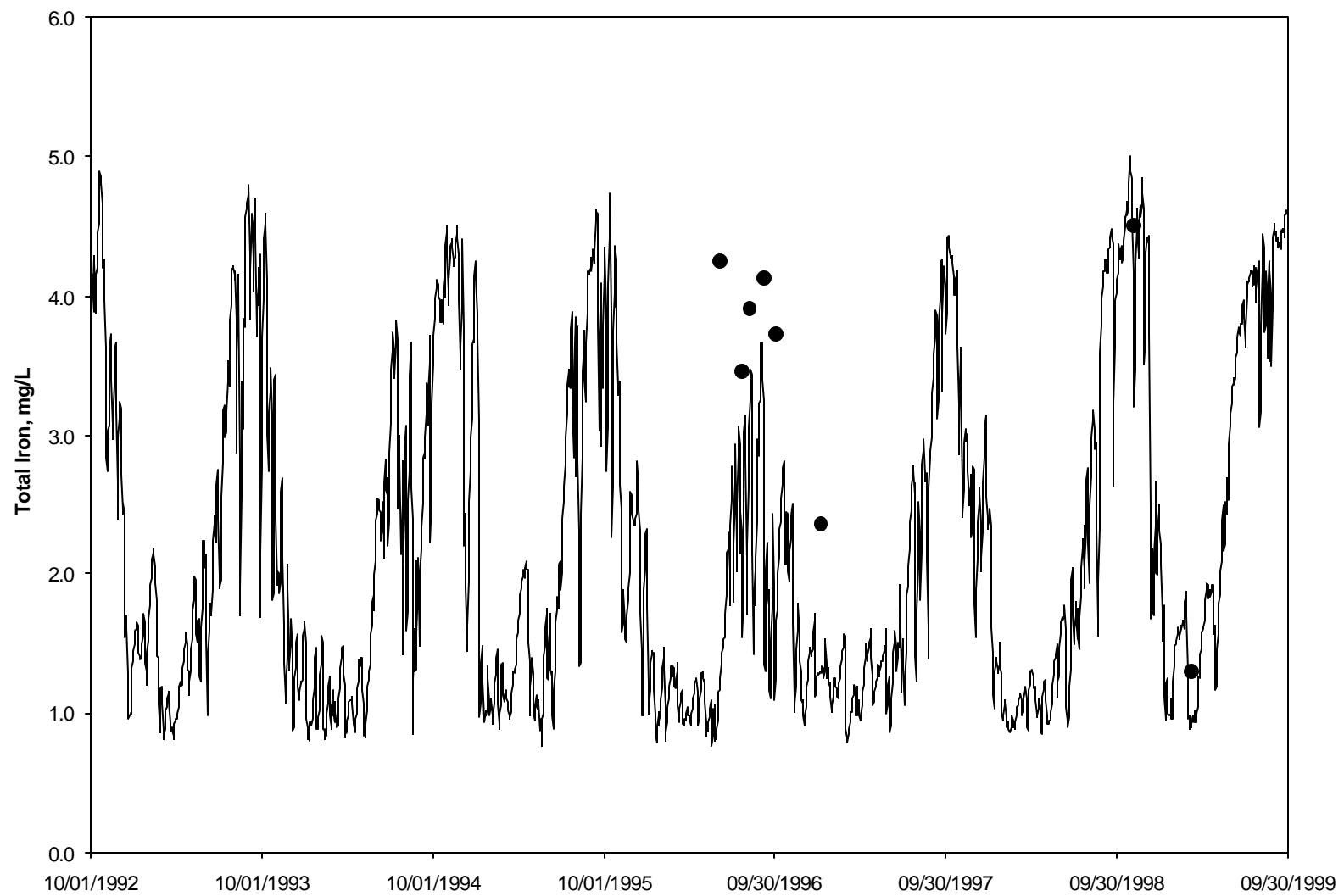


Figure D6. Time Series Plot of Observed and Calculated Total Iron for WVDEP-SRG Station PC060.

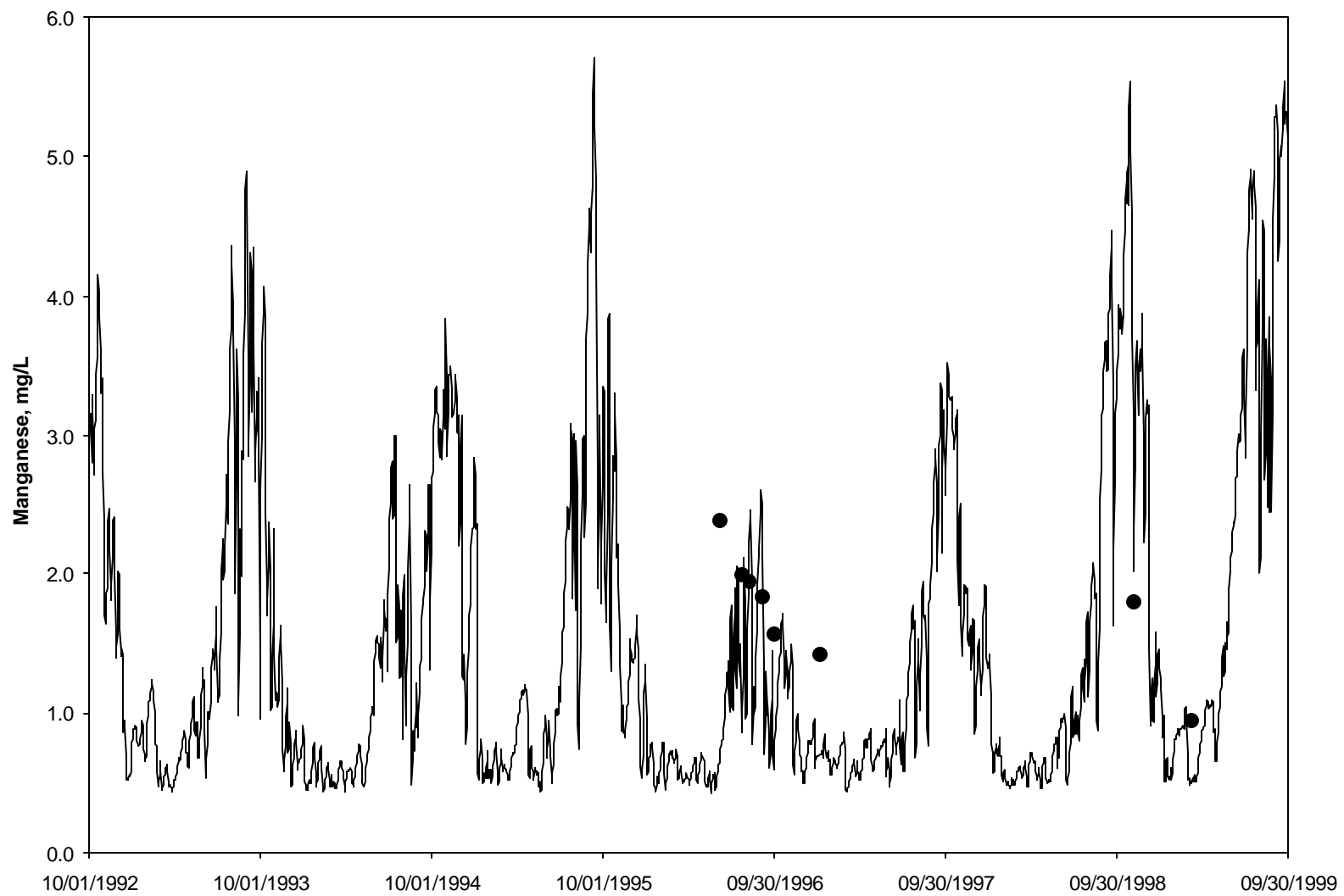


Figure D7. Time Series Plot of Observed and Calculated Manganese for WVDEP-SRG Station PC060.

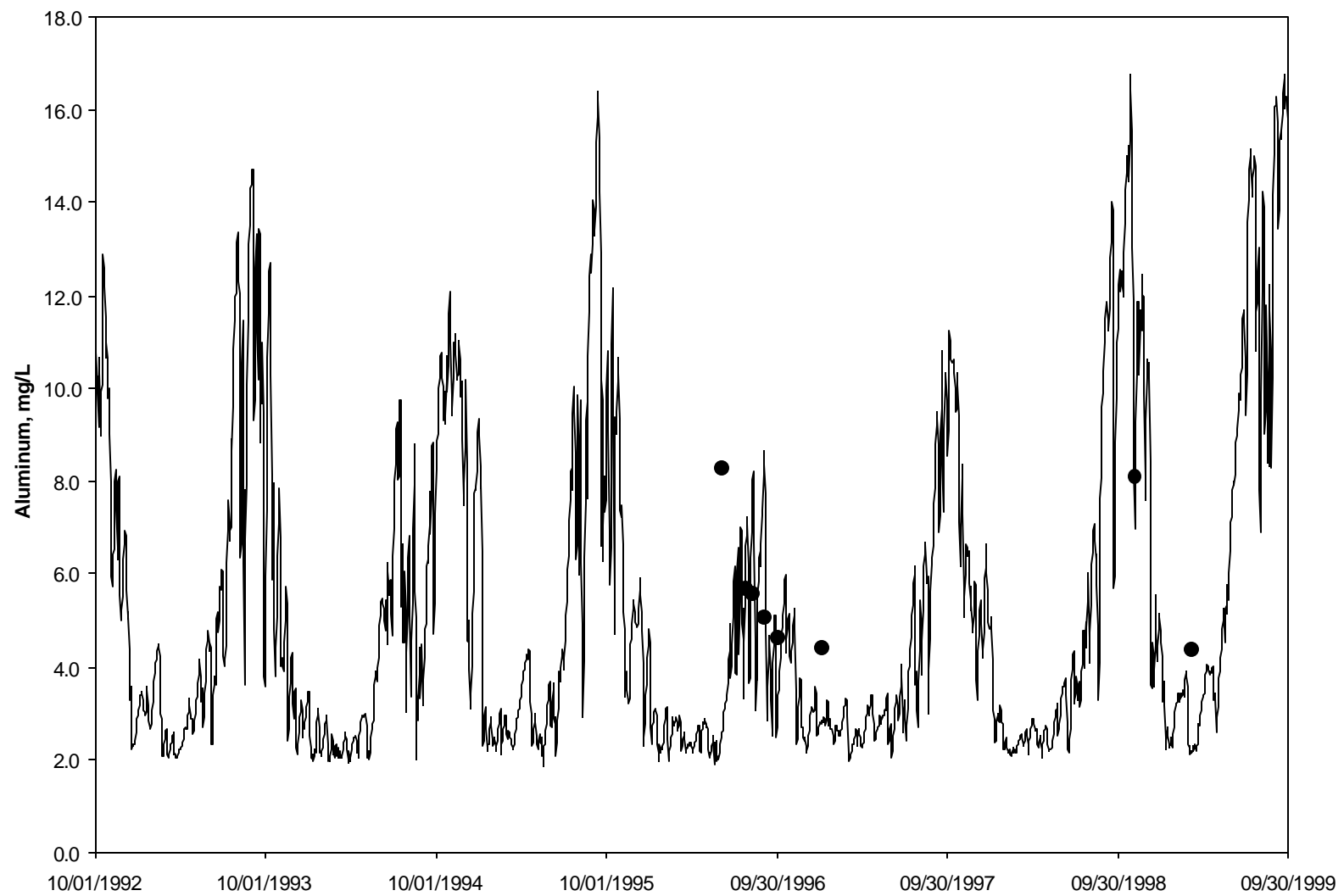


Figure D8. Time Series Plot of Observed and Calculated Total Aluminum for WVDEP-SRG Station PC060.

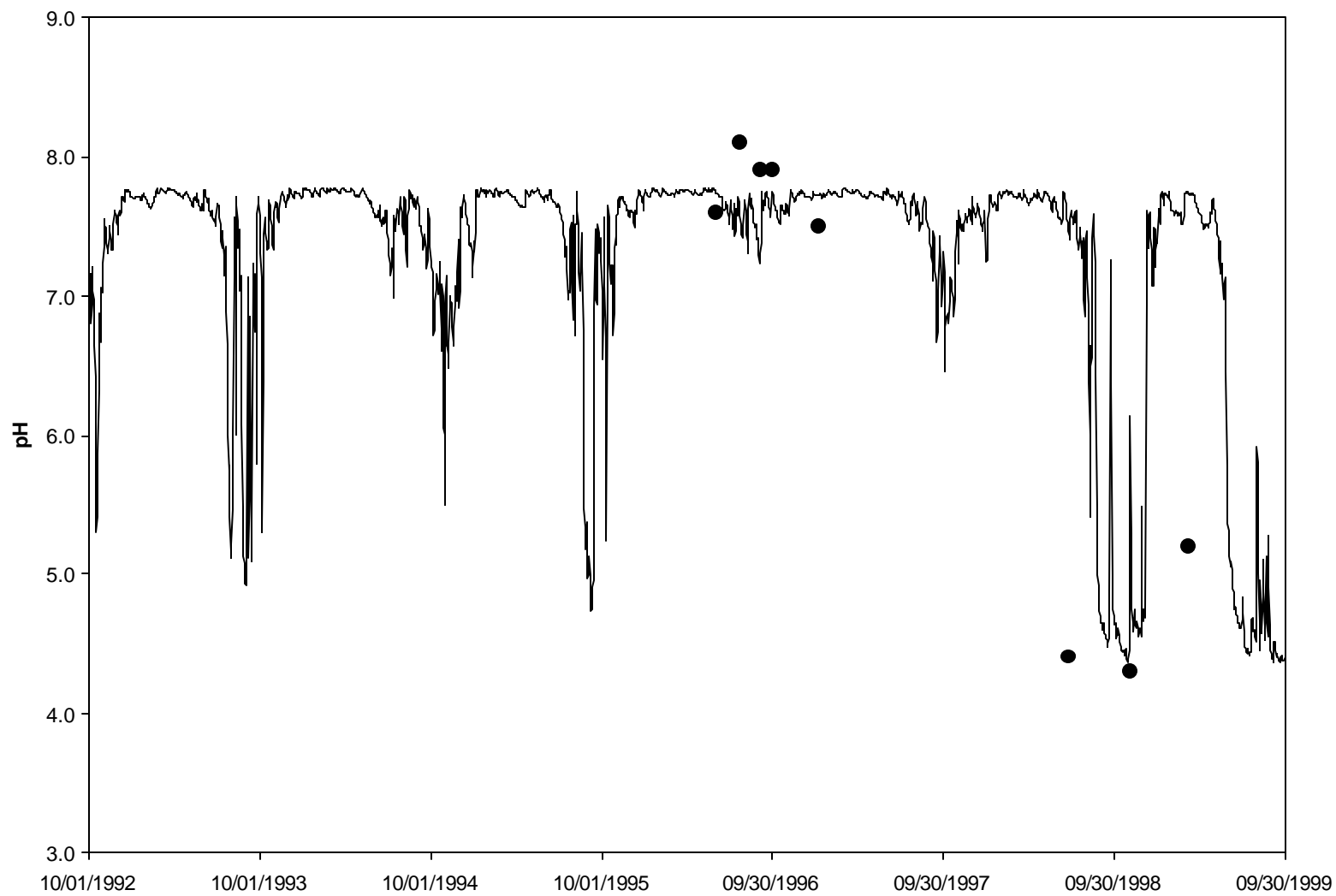


Figure D9. Time Series Plot of Observed and Calculated pH for WVDEP-SRG Station PC073.

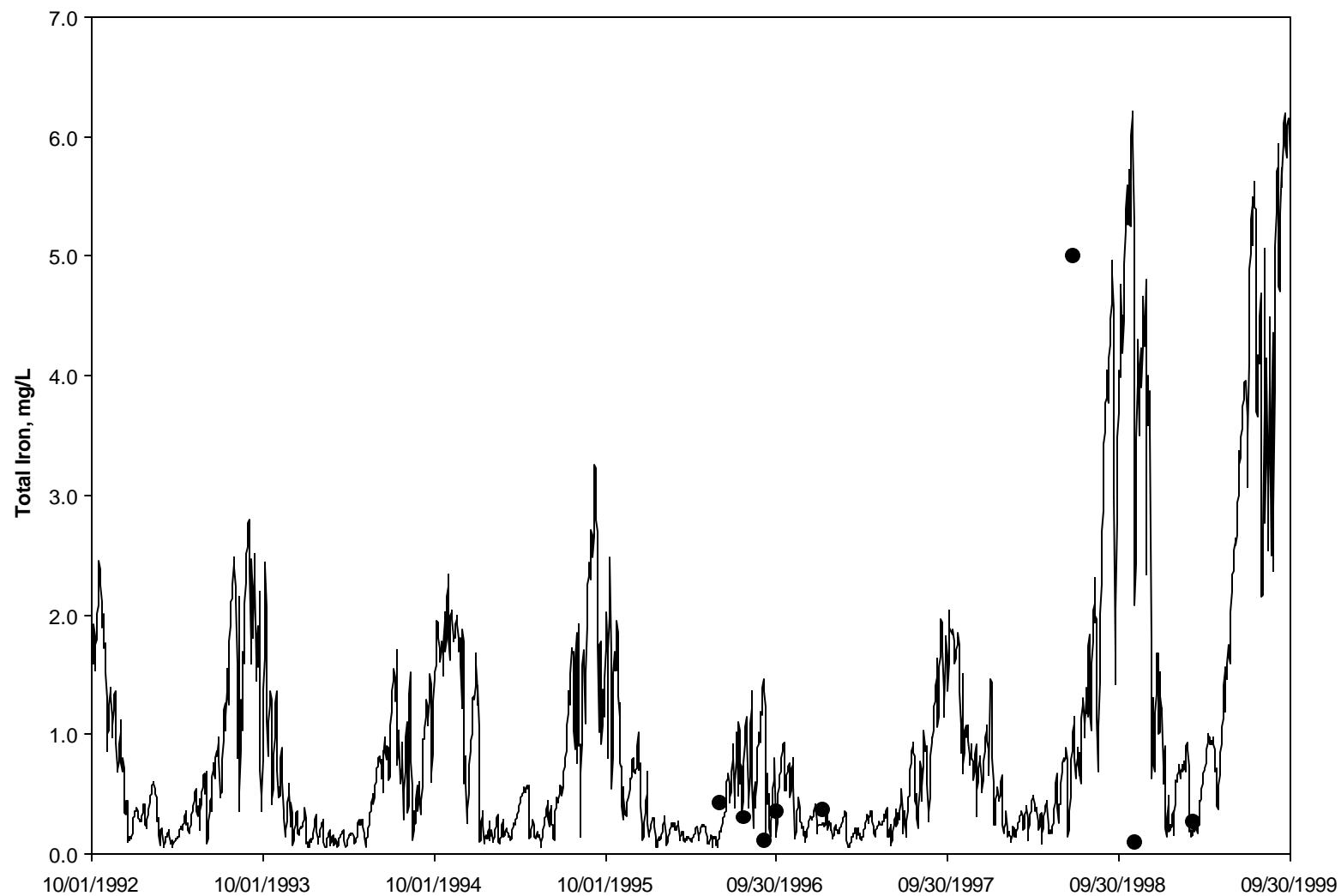


Figure D10. Time Series Plot of Observed and Calculated Total Iron for WVDEP-SRG Station PC073.

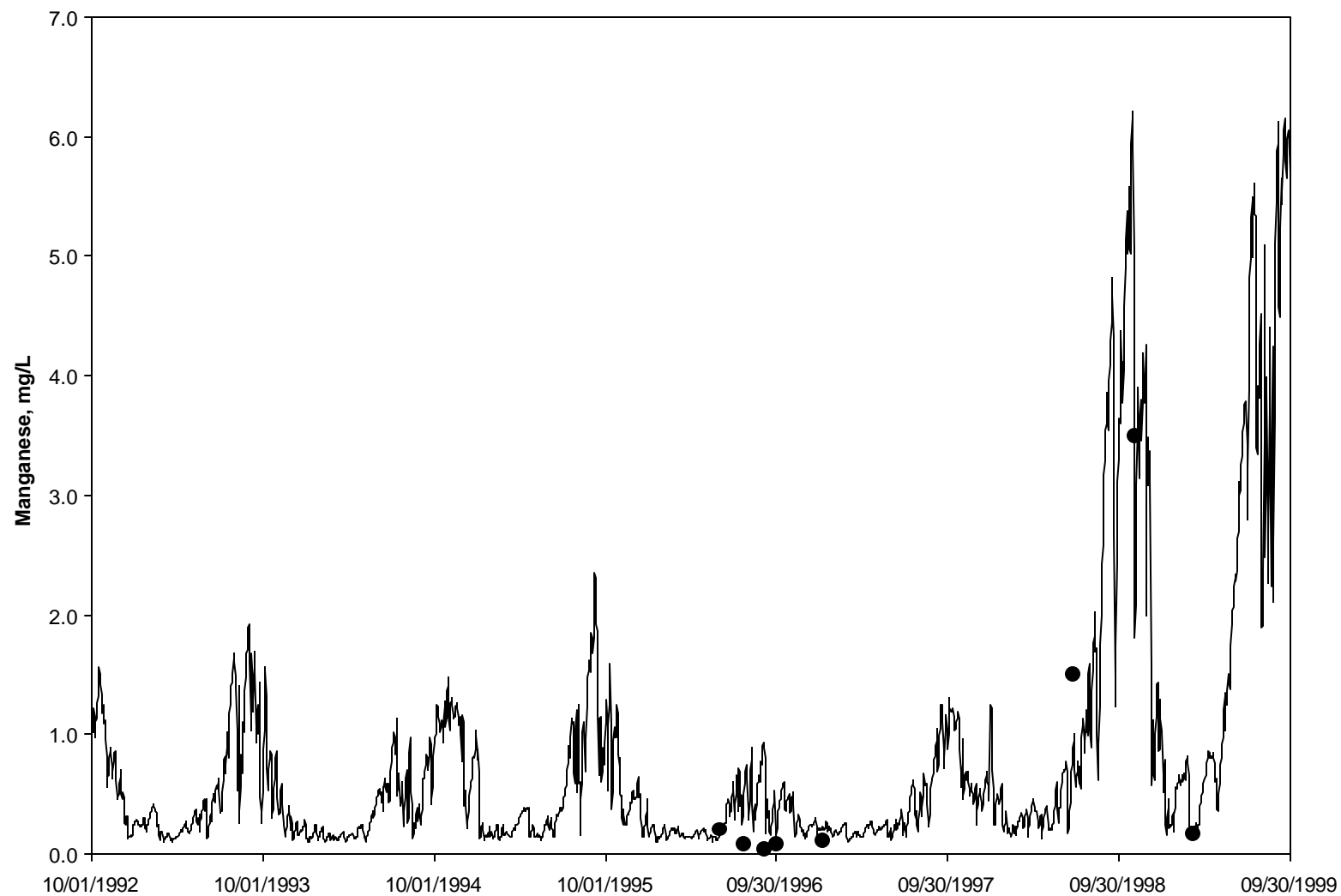


Figure D11. Time Series Plot of Observed and Calculated Manganese for WVDEP-SRG Station PC073.

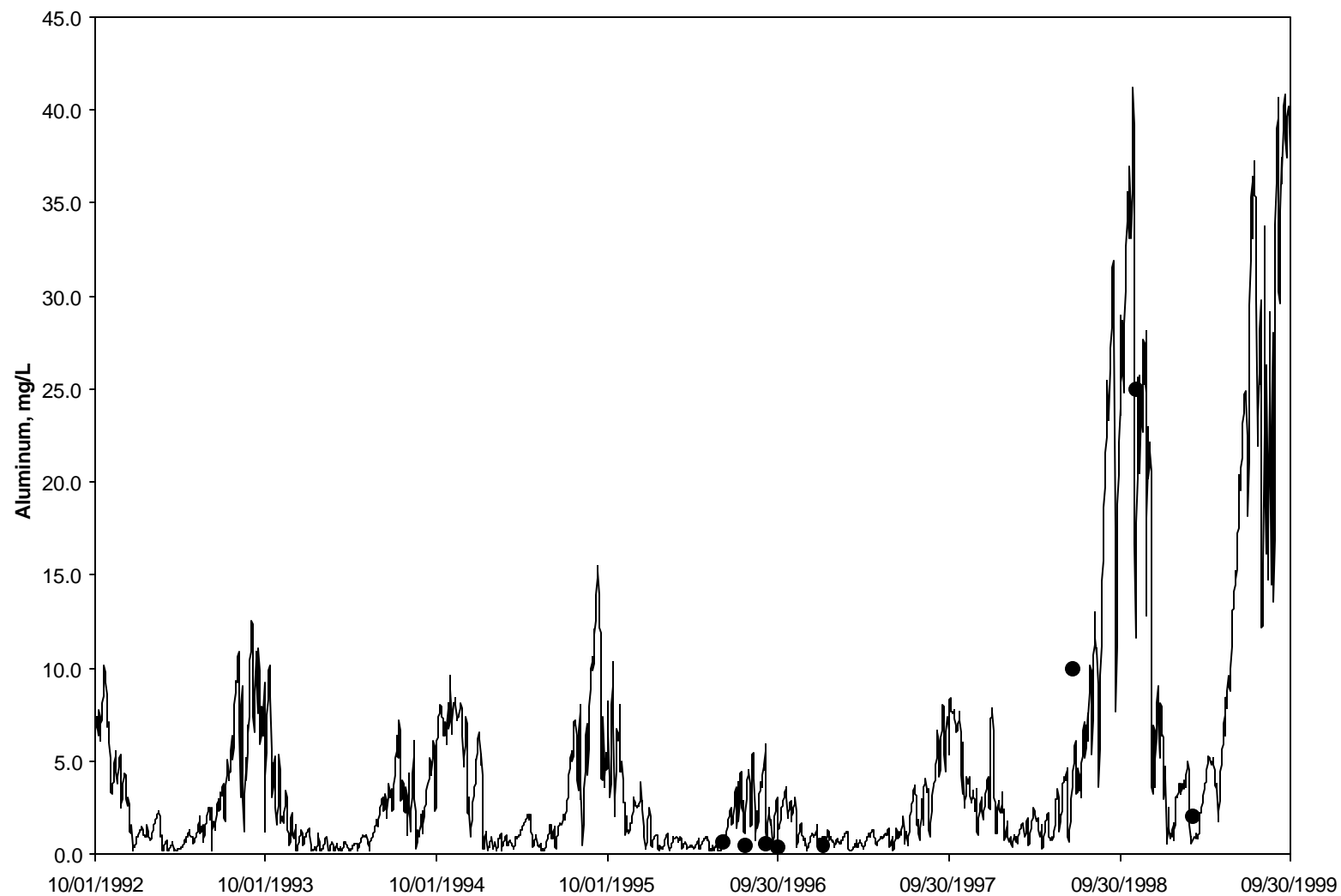


Figure D12. Time Series Plot of Observed and Calculated Aluminum for WVDEP-SRG Station PC073.

Table D1. Finite Difference and Hydraulic Parameters for Paint Creek TAMD Models.

SWS	Model Nodes	Node Spacing, m	Bottom Width, m	Inverse Side Slope	Mannings Roughness Coefficient	Channel Bottom Slope	Maximum Time Step, days
1	39	201.20	8.82	2.00	0.030	0.002200	1.000
2	25	208.15	1.35	2.00	0.530	0.062200	0.010
3	8	216.56	8.82	2.00	0.030	0.002000	1.000
4	39	200.37	3.28	2.00	0.110	0.062200	0.010
5	26	206.85	4.08	2.00	0.030	0.002130	1.000
6	10	200.21	0.91	2.00	0.160	0.081820	1.000
7	18	202.28	4.08	2.00	0.030	0.004950	1.000
8	16	206.61	0.20	2.00	0.190	0.110000	1.000
9	17	204.21	4.08	2.00	0.030	0.005220	1.000
10	34	200.58	0.91	2.00	0.410	0.061300	1.000
11	5	241.38	1.18	2.00	0.160	0.037400	1.000
12	13	213.62	1.01	2.00	0.070	0.118000	1.000
13	15	213.21	1.01	2.00	0.070	0.109000	1.000
14	8	206.76	14.50	2.00	0.190	0.012400	1.000
15	10	216.27	0.84	2.00	0.410	0.153500	1.000
16	12	203.54	14.50	2.00	0.190	0.004910	1.000
17	8	218.00	0.28	2.00	0.981	0.268000	0.010
18	5	86.98	14.50	2.00	0.190	0.005750	1.000
19	12	217.90	0.55	2.00	0.710	0.186900	0.010
20	5	223.58	14.50	2.00	0.190	0.023500	1.000
21	10	219.17	0.30	0.00	1.000	0.172400	1.000
22	10	213.31	14.50	2.00	0.190	0.002080	1.000
23	17	206.42	2.31	2.00	0.180	0.122900	1.000
24	7	225.01	14.50	2.00	0.190	0.017000	1.000
25	11	217.83	0.73	2.00	0.829	0.140500	0.001
26	9	221.64	14.50	2.00	0.190	0.018100	1.000
27	12	216.88	1.08	2.00	0.160	0.153400	0.010
28	10	205.78	14.50	2.00	0.190	0.018400	1.000
29	14	206.33	0.36	2.00	0.180	0.090220	1.000
30	7	217.57	14.50	2.00	0.190	0.006130	1.000
31	25	205.57	0.90	2.00	0.180	0.068102	1.000
32	8	202.32	14.50	2.00	0.190	0.023300	1.000
33	10	203.76	2.22	2.00	0.120	0.056167	0.001
34	17	207.56	14.50	2.00	0.190	0.001810	1.000
35	16	211.89	1.38	2.00	0.440	0.092816	0.010
36	5	87.75	14.50	2.00	0.190	0.025600	1.000
37	19	210.21	0.42	2.00	0.040	0.005810	1.000
38	25	205.74	1.72	2.00	0.270	0.030378	1.000
39	38	201.26	1.03	2.00	1.380	0.017323	1.000
40	25	202.50	1.72	2.00	0.270	0.012600	1.000
41	13	208.27	0.90	2.00	0.160	0.014404	1.000
42	13	200.12	1.72	2.00	0.270	0.000416	1.000
43	18	203.54	0.93	2.00	0.350	0.031790	1.000
44	5	201.83	1.72	2.00	0.270	0.001240	1.000
45	20	205.83	0.61	2.00	0.210	0.084896	1.000
46	18	205.98	1.72	2.00	0.270	0.002000	1.000

Table D1. Finite Difference and Hydraulic Parameters for Paint Creek TAMDLC Models.

SWS	Model Nodes	Node Spacing, m	Bottom Width, m	Inverse Side Slope	Mannings Roughness Coefficient	Channel Bottom Slope	Maximum Time Step, days
47	26	203.89	0.59	2.00	0.160	0.009809	1.000
48	21	209.39	1.72	2.00	0.270	0.001430	1.000
49	18	207.59	0.49	2.00	0.460	0.066873	1.000
50	10	200.21	1.72	2.00	0.270	0.010500	1.000
51	9	208.87	6.30	2.00	0.100	0.017400	1.000
52	5	243.98	4.24	2.00	0.290	0.017400	1.000
53	26	101.25	1.18	2.00	0.210	0.015013	1.000
54	32	206.24	0.46	2.00	0.260	0.028936	1.000
55	32	206.24	0.34	2.00	0.620	0.031268	1.000
56	5	152.49	3.08	2.00	0.330	0.004920	1.000
57	11	202.72	0.09	2.00	0.370	0.084846	1.000
58	13	200.79	3.08	2.00	0.330	0.009960	1.000
59	15	209.03	1.03	2.00	0.500	0.068684	0.010
60	8	226.45	3.08	2.00	0.330	0.001260	1.000
61	19	207.57	0.11	2.00	0.070	0.034793	1.000
62	20	200.45	0.08	2.00	0.003	0.027570	1.000

Table D2. Water Quality Parameters for Paint Creek TAMDL Models.

SWS	Hydrodynamic Dispersion, m ² /day	Mean Sediment Diameter, m	First Net Acidity - pH Constant, SU/(mg/L) ² b	Second Net Acidity - pH Constant (b)	Empirical Aluminum Kinetic Coefficient	Empirical Manganese Kinetic Coefficient, 1/(mg/L) ⁴ /day
1	1.0E+05	4.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
2	1.0E+03	1.0E-06	6.136517	-0.027648	1.0E-18	1.0E+00
3	1.0E+00	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
4	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
5	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
6	1.0E+03	1.0E-06	6.485866	-0.023431	1.0E-18	1.0E+00
7	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
8	1.0E+00	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
9	1.0E+05	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
10	9.0E+05	2.0E-06	6.414215	-0.046430	1.0E-18	1.0E+00
11	1.0E+00	1.0E-06	5.844200	-0.020000	1.0E-18	1.0E+00
12	1.0E+03	1.0E-06	5.844246	-0.030484	1.0E-18	1.0E+00
13	1.0E+04	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
14	1.0E+03	1.0E-05	6.350000	-0.020000	1.0E-18	1.0E+00
15	1.0E+03	1.5E-06	5.400000	-0.020000	1.0E-18	1.0E+00
16	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
17	1.0E+03	1.0E-06	5.478805	-0.032232	1.0E-15	1.0E+00
18	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
19	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+04
20	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
21	1.0E+03	1.0E-06	6.006690	-0.032960	1.0E-18	1.0E+00
22	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
23	1.0E+03	5.0E-06	5.327925	-0.034780	1.0E-18	1.0E+00
24	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
25	1.0E+03	1.0E-06	5.000000	-0.026332	1.0E-18	1.0E+00
26	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
27	1.0E+03	1.0E-06	5.283879	-0.045928	1.0E-20	1.0E+00
28	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
29	1.0E+03	1.0E-06	5.000000	-0.067674	1.0E-16	1.0E+00
30	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
31	1.0E+03	1.0E-06	5.000000	-0.050846	1.0E-16	1.0E+00
32	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
33	1.0E+03	1.0E-06	6.700000	-0.020000	1.0E-16	1.0E+00
34	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
35	1.0E+03	1.0E-06	5.508978	-0.036358	1.0E-18	1.0E+00
36	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
37	5.0E+03	2.0E-06	5.540300	-0.038241	1.0E-18	1.0E+00
38	1.0E+03	1.0E-06	5.023840	-0.051921	1.0E-16	1.0E+00
39	1.0E+03	1.0E-06	5.754938	-0.031614	1.0E-18	1.0E+00
40	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
41	1.0E+03	1.0E-06	6.626049	-0.018438	1.0E-16	1.0E+00
42	1.0E+03	1.0E-06	6.300000	-0.020000	1.0E-18	1.0E+00

Table D2. Water Quality Parameters for Paint Creek TAMDLC Models.

SWS	Hydrodynamic Dispersion, m ² /day	Mean Sediment Diameter, m	First Net Acidity - pH Constant, SU/(mg/L) ^{2b}	Second Net Acidity - pH Constant (b)	Empirical Aluminum Kinetic Coefficient	Empirical Manganese Kinetic Coefficient, 1/(mg/L) ⁴ /day
43	1.0E+03	1.0E-06	6.618670	-0.006549	1.0E-18	1.0E+00
44	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
45	1.0E+03	1.0E-06	5.000000	-0.047517	1.0E-18	1.0E+00
46	1.0E+03	1.0E-06	6.000000	-0.020000	1.0E-18	1.0E+00
47	1.0E+03	1.0E-06	5.000000	-0.049391	1.0E-18	1.0E+00
48	1.0E+03	1.0E-06	6.000000	-0.020000	1.0E-18	1.0E+00
49	1.0E+03	1.0E-06	6.465914	-0.009138	1.0E-15	1.0E+00
50	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+02
51	1.0E+03	5.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
52	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+00
53	1.0E+03	1.0E-06	6.401801	-0.020721	1.0E-18	1.0E+00
54	1.0E+03	1.0E-06	6.000000	-0.023832	1.0E-18	1.0E+00
55	1.0E+03	1.0E-06	6.000000	-0.022572	1.0E-18	1.0E+02
56	1.0E+03	5.0E-06	5.895000	-0.015000	1.0E-19	1.0E+00
57	1.0E+03	1.0E-06	7.000000	-0.011045	1.0E-18	1.0E+02
58	1.0E+03	1.0E-06	6.500000	-0.020000	1.0E-18	1.0E+02
59	1.0E+03	1.0E-06	7.000000	-0.006399	1.0E-18	1.0E+02
60	1.0E+03	1.0E-06	6.138000	-0.022615	1.0E-18	1.0E+00
61	1.0E+05	2.0E-06	6.703630	-0.011238	1.0E-15	1.0E+04
62	1.0E+03	1.0E-06	6.138009	-0.022615	1.0E-18	1.0E+00

Surface Heat Transfer Coefficient, calories/K
1.0E-05
1.0E-05
1.0E-05
1.0E-05
1.0E-05
1.0E-05
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1.0E-05

Table D3. Upstream Concentrations for Tributary Sub-Watersheds in Calibration and Baseline Models.					
SWS	Net Acidity, mg/L CaCO₃ equivalents	pH, SU	Iron, mg/L	Manganese, mg/L	Aluminum, mg/L
2	-10.66	6.99	0.08	0.03	0.39
4	26.71	5.70	0.13	0.71	1.83
6	4.09	6.07	0.14	0.00	0.50
8	0.00	6.50	0.09	0.01	0.57
10	0.00	6.41	0.02	0.03	0.38
12	30.00	4.75	0.00	0.25	0.30
13	0.00	6.50	2.60	1.00	6.00
15	-2999.27	7.44	0.16	0.03	2.18
17	-10.00	6.36	0.23	0.08	0.76
19	50.00	5.56	0.13	1.00	5.66
21	-50.00	7.77	0.05	0.10	0.10
23	-15.90	6.46	6.00	4.45	7.15
25	24.25	4.23	0.13	0.00	0.00
27	-15.50	6.80	0.24	0.16	0.95
29	-20.00	7.50	0.33	0.14	0.43
31	-36.00	7.20	0.26	0.24	1.05
33	0.00	6.70	0.07	0.03	0.55
35	-9.38	6.48	0.21	0.00	0.60
38	-14.88	6.65	0.16	0.01	0.53
39	-11.93	6.73	1.27	0.07	0.57
41	-1.59	6.74	0.13	0.06	0.17
43	15.44	6.39	0.05	0.04	0.30
45	-47.46	7.22	0.29	0.23	1.55
47	-28.47	6.96	0.06	0.02	0.28
49	-4.67	6.65	0.00	0.00	0.26
53	-0.24	6.40	0.08	0.00	0.31
54	-21.29	6.94	0.04	0.06	0.54
55	-16.53	6.81	0.60	0.37	0.55
57	9.62	6.66	0.00	0.00	0.10
59	5.16	6.85	0.06	0.01	0.16
61	-12.50	7.10	0.26	0.03	0.24
62	-20.17	7.03	0.05	0.01	0.40

Table D4. Upstream Concentrations for Tributary Sub-Watersheds in Calibration and Baseline Models.

SWS	Net Acidity, mg/L CaCO ₃ equivalents	pH, SU	Iron, mg/L	Manganese, mg/L	Aluminum, mg/L
2	-10.66	6.99	0.08	0.03	0.20
4	0.00	6.50	0.13	0.71	0.70
6	-5.00	6.99	0.14	0.00	0.25
8	-10.00	7.13	0.09	0.01	0.04
10	-1.50	6.66	0.02	0.03	0.10
12	-30.00	7.19	0.00	0.16	0.15
13	-10.00	7.13	0.36	0.20	0.10
15	-2999.27	7.44	0.02	0.03	0.04
17	-15.00	6.52	0.23	0.08	0.35
19	0.00	6.50	0.13	0.40	0.01
21	-50.00	7.77	0.05	0.10	0.02
23	-18.00	6.51	0.40	0.05	0.05
25	-200.00	6.61	0.01	0.00	0.00
27	-33.00	7.29	0.01	0.01	0.01
29	-20.00	7.50	0.33	0.14	0.43
31	-36.00	7.20	0.01	0.01	0.01
33	0.00	6.70	0.07	0.03	0.55
35	-20.00	6.85	0.21	0.00	0.00
38	-14.88	6.65	0.16	0.01	0.23
39	-11.93	6.73	1.00	0.07	0.57
41	-1.59	6.74	0.13	0.06	0.17
43	0.00	6.62	0.00	0.04	0.10
45	-50.00	7.25	0.03	0.03	0.05
47	-28.47	6.96	0.00	0.02	0.28
49	-4.67	6.65	0.00	0.00	0.26
53	-4.78	6.83	0.08	0.00	0.03
54	-21.29	6.94	0.04	0.06	0.03
55	-16.53	6.81	0.05	0.37	0.05
57	2.99	6.83	0.00	0.00	0.00
59	5.16	6.85	0.06	0.01	0.06
61	-12.50	7.10	0.03	0.03	0.04
62	-20.17	7.03	0.02	0.01	0.04

Appendix E: Wasteload Allocations

Table E1. Allocated Metal Effluent Concentrations for Mining Point Sources in Paint Creek Watershed.

NPDES Permit	Outlet	Sub-WS	Wasteload Allocations			Wasteload Allocations		
			Iron (mg/L)	Manganese (mg/L)	Aluminum (mg/L)	Iron (Mg/yr)	Manganese (Mg/yr)	Aluminum (Mg/yr)
WV0057011	18	8	3.20	2.00	1.25	0.091383	0.057114	0.035696
WV1009311	4	9	3.20	2.00	4.30	0.013124	0.008203	0.017635
WV1009311	5	9	3.20	2.00	4.30	0.016405	0.010253	0.022044
WV1015257	2	10	3.20	2.00	2.71	0.272940	0.170587	0.231146
WV1009320	2	11	3.20	2.00	4.30	0.003079	0.001924	0.004137
WV1009320	3	11	3.20	2.00	4.30	0.001539	0.000962	0.002069
WV1009311	6	15	2.00	1.13	0.86	0.017590	0.009938	0.007564
WV1014951	3	23	3.20	2.00	1.81	0.039962	0.024977	0.022604
WV1014951	4	23	3.20	2.00	1.81	0.001537	0.000961	0.000869
WV1014951	42	23	3.20	2.00	1.81	0.013833	0.008646	0.007824
WV1014951	5	23	3.20	2.00	1.81	0.015370	0.009606	0.008694
WV1014951	6	23	3.20	2.00	1.81	0.003074	0.001921	0.001739
WV1014951	7	23	3.20	2.00	1.81	0.015370	0.009606	0.008694
WV1014951	8	23	3.20	2.00	1.81	0.006148	0.003843	0.003478
WV1014951	9	26	3.20	2.00	4.30	0.009222	0.005764	0.012392
WV1014951	35	27	1.83	1.05	0.75	0.015826	0.009080	0.006486
WV1014951	36	27	1.83	1.05	0.75	0.000879	0.000504	0.000360
WV1014951	37	27	1.83	1.05	0.75	0.007913	0.004540	0.003243
WV1014951	14	27	1.83	1.05	0.75	0.005275	0.003027	0.002162
WV1014951	39	27	1.83	1.05	0.75	0.005275	0.003027	0.002162
WV1014951	40	27	1.83	1.05	0.75	0.001758	0.001009	0.000721
WV1014951	45	27	1.83	1.05	0.75	0.009671	0.005549	0.003964
WV1014951	32	27	1.83	1.05	0.75	0.008165	0.004685	0.003346
WV1014951	43	27	1.83	1.05	0.75	0.226799	0.130130	0.092950
WV1014951	10	27	1.83	1.05	0.75	0.010548	0.006052	0.004323
WV1014951	11	27	1.83	1.05	0.75	0.001758	0.001009	0.000720
WV1014951	12	27	1.83	1.05	0.75	0.014943	0.008574	0.006124
WV1014951	13	27	1.83	1.05	0.75	0.000879	0.000504	0.000360
WV1014951	41	27	1.83	1.05	0.75	0.004395	0.002522	0.001801
WV1002368	1	28	3.20	2.00	4.30	0.542277	0.338923	0.728685
WV1002368	4	28	3.20	2.00	4.30	0.136721	0.085451	0.183719
WV1002368	5	28	3.20	2.00	4.30	0.015362	0.009601	0.020643
WV1002368	2	28	3.20	2.00	4.30	0.044550	0.027844	0.059864
WV1002368	3	28	3.20	2.00	4.30	0.102925	0.064328	0.138306
WV0028452	9	31	1.86	1.05	0.91	0.001784	0.001007	0.000873
WV0028452	12	31	1.86	1.05	0.91	0.000892	0.000504	0.000436
WV0028452	13	31	1.86	1.05	0.91	0.000892	0.000504	0.000436
WV0028452	15	31	1.86	1.05	0.91	0.008922	0.005036	0.004365
WV0028452	8	31	1.86	1.05	0.91	0.001784	0.001007	0.000873
WV0028452	14	31	1.86	1.05	0.91	0.008922	0.005036	0.004365
WV0028452	3	31	1.86	1.05	0.91	0.005353	0.003022	0.002619
WV1014951	31	31	1.86	1.05	0.91	0.008029	0.004533	0.003928
WV1014951	33	31	1.86	1.05	0.91	0.008029	0.004533	0.003928
WV1014951	38	31	1.86	1.05	0.91	0.008029	0.004533	0.003928
WV0028452	5	31	1.86	1.05	0.91	0.000892	0.000504	0.000436
WV0028452	1	31	1.86	1.05	0.91	0.004461	0.002518	0.002182

Table E1. Allocated Metal Effluent Concentrations for Mining Point Sources in Paint Creek Watershed.

NPDES Permit	Outlet	Sub-WS	Wasteload Allocations			Wasteload Allocations		
			Iron (mg/L)	Manganese (mg/L)	Aluminum (mg/L)	Iron (Mg/yr)	Manganese (Mg/yr)	Aluminum (Mg/yr)
WV0028452	4	31	1.86	1.05	0.91	0.016059	0.009066	0.007857
WV0028452	6	31	1.86	1.05	0.91	0.003569	0.002015	0.001746
WV1012487	160	35	3.20	2.00	2.05	0.003440	0.002150	0.002204
WV1012487	161	35	3.20	2.00	2.05	0.055034	0.034396	0.035256
WV1012487	156	35	3.20	2.00	2.05	0.003440	0.002150	0.002204
WV1012487	157	35	3.20	2.00	2.05	0.030957	0.019348	0.019832
WV1012487	158	35	3.20	2.00	2.05	0.030957	0.019348	0.019832
WV1012487	159	35	3.20	2.00	2.05	0.008599	0.005374	0.005509
WV1012487	162	35	3.20	2.00	2.05	0.006879	0.004300	0.004407
WV1012487	163	35	3.20	2.00	2.05	0.001720	0.001075	0.001102
WV1012487	164	35	3.20	2.00	2.05	0.029237	0.018273	0.018730
WV1012487	165	35	3.20	2.00	2.05	0.010319	0.006449	0.006611
WV1012487	166	35	3.20	2.00	2.05	0.003440	0.002150	0.002204
WV1012487	151	35	3.20	2.00	2.05	0.008599	0.005374	0.005509
WV1012487	152	35	3.20	2.00	2.05	0.001720	0.001075	0.001102
WV1012487	153	35	3.20	2.00	2.05	0.001720	0.001075	0.001102
WV1012487	154	35	3.20	2.00	2.05	0.005159	0.003225	0.003305
WV1012487	9	40	1.57	2.00	4.30	0.012657	0.016123	0.034665
WV1012487	10	40	1.57	2.00	4.30	0.004219	0.005374	0.011555
WV1012487	64	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	7	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	8	40	1.57	2.00	4.30	0.002531	0.003225	0.006933
WV1012487	32	40	1.57	2.00	4.30	0.008438	0.010749	0.023110
WV1012487	33	40	1.57	2.00	4.30	0.024470	0.031172	0.067019
WV1012487	34	40	1.57	2.00	4.30	0.007594	0.009674	0.020799
WV1012487	65	40	1.57	2.00	4.30	0.001688	0.002150	0.004622
WV1012487	3	40	1.57	2.00	4.30	0.004219	0.005374	0.011555
WV1012487	4	40	1.57	2.00	4.30	0.001688	0.002150	0.004622
WV1012487	5	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	6	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	35	40	1.57	2.00	4.30	0.006750	0.008599	0.018488
WV1012487	36	40	1.57	2.00	4.30	0.004219	0.005374	0.011555
WV1012487	37	40	1.57	2.00	4.30	0.014344	0.018273	0.039287
WV1012487	38	40	1.57	2.00	4.30	0.003375	0.004300	0.009244
WV1012487	39	40	1.57	2.00	4.30	0.003375	0.004300	0.009244
WV1012487	40	40	1.57	2.00	4.30	0.004219	0.005374	0.011555
WV1012487	41	40	1.57	2.00	4.30	0.189008	0.240775	0.517665
WV1012487	42	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	66	40	1.57	2.00	4.30	0.001688	0.002150	0.004622
WV1012487	67	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	68	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	69	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	70	40	1.57	2.00	4.30	0.013501	0.017198	0.036976
WV1012487	71	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	72	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	73	40	1.57	2.00	4.30	0.019407	0.024722	0.053153
WV1012487	74	40	1.57	2.00	4.30	0.001688	0.002150	0.004622
WV1012487	75	40	1.57	2.00	4.30	0.000844	0.001075	0.002311

Table E1. Allocated Metal Effluent Concentrations for Mining Point Sources in Paint Creek Watershed.

NPDES Permit	Outlet	Sub-WS	Wasteload Allocations			Wasteload Allocations		
			Iron (mg/L)	Manganese (mg/L)	Aluminum (mg/L)	Iron (Mg/yr)	Manganese (Mg/yr)	Aluminum (Mg/yr)
WV1012487	76	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	77	40	1.57	2.00	4.30	0.001688	0.002150	0.004622
WV1012487	78	40	1.57	2.00	4.30	0.021938	0.027947	0.060086
WV1012487	79	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	80	40	1.57	2.00	4.30	0.006750	0.008599	0.018488
WV1012487	81	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	82	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	83	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	84	40	1.57	2.00	4.30	0.017720	0.022573	0.048531
WV1012487	85	40	1.57	2.00	4.30	0.004219	0.005374	0.011555
WV1012487	86	40	1.57	2.00	4.30	0.004219	0.005374	0.011555
WV1012487	87	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	88	40	1.57	2.00	4.30	0.001688	0.002150	0.004622
WV1012487	89	40	1.57	2.00	4.30	0.012657	0.016123	0.034665
WV1012487	90	40	1.57	2.00	4.30	0.007594	0.009674	0.020799
WV1012487	109	40	1.57	2.00	4.30	0.001688	0.002150	0.004622
WV1012487	110	40	1.57	2.00	4.30	0.000844	0.001075	0.002311
WV1012487	19	42	3.20	2.00	4.30	0.003988	0.002492	0.005359
WV1012487	20	42	3.20	2.00	4.30	0.005982	0.003739	0.008038
WV1012487	22	42	3.20	2.00	4.30	0.015951	0.009969	0.021434
WV1012487	57	42	3.20	2.00	4.30	0.003988	0.002492	0.005359
WV1012487	58	42	3.20	2.00	4.30	0.001994	0.001246	0.002679
WV1012487	59	42	3.20	2.00	4.30	0.039878	0.024924	0.053586
WV1012487	60	42	3.20	2.00	4.30	0.012039	0.007524	0.016177
WV1012487	104	42	3.20	2.00	4.30	0.005982	0.003739	0.008038
WV1012487	105	42	3.20	2.00	4.30	0.015951	0.009969	0.021434
WV1012487	106	42	3.20	2.00	4.30	0.030957	0.019348	0.041598
WV1012487	108	42	3.20	2.00	4.30	0.005159	0.003225	0.006933
WV1012487	115	42	3.20	2.00	4.30	0.257212	0.160758	0.345629
WV1012487	167	42	3.20	2.00	4.30	0.001994	0.001246	0.002679
WV1012487	21	42	3.20	2.00	4.30	0.001994	0.001246	0.002679
WV1012487	61	42	3.20	2.00	4.30	0.001720	0.001075	0.002311
WV1012487	62	42	3.20	2.00	4.30	0.001720	0.001075	0.002311
WV1012487	16	42	3.20	2.00	4.30	0.010319	0.006449	0.013866
WV1012487	17	42	3.20	2.00	4.30	0.001720	0.001075	0.002311
WV1012487	13	42	3.20	2.00	4.30	0.010319	0.006449	0.013866
WV1012487	15	42	3.20	2.00	4.30	0.013759	0.008599	0.018488
WV1012487	11	42	3.20	2.00	4.30	0.001720	0.001075	0.002311
WV1012487	12	42	3.20	2.00	4.30	0.015478	0.009674	0.020799
WV1012487	14	42	3.20	2.00	4.30	0.018918	0.011824	0.025421
WV1012487	63	42	3.20	2.00	4.30	0.013759	0.008599	0.018488
WV1012487	111	42	3.20	2.00	4.30	0.005159	0.003225	0.006933
WV1002066	1	45	1.80	1.04	0.79	0.011216	0.006480	0.004922
WV1002066	2	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1002066	3	45	1.80	1.04	0.79	0.009870	0.005703	0.004332
WV1002066	6	45	1.80	1.04	0.79	0.006729	0.003888	0.002953
WV1012487	43	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	145	45	1.80	1.04	0.79	0.005608	0.003240	0.002461

Table E1. Allocated Metal Effluent Concentrations for Mining Point Sources in Paint Creek Watershed.

NPDES Permit	Outlet	Sub-WS	Wasteload Allocations			Wasteload Allocations		
			Iron (mg/L)	Manganese (mg/L)	Aluminum (mg/L)	Iron (Mg/yr)	Manganese (Mg/yr)	Aluminum (Mg/yr)
WV1012487	149	45	1.80	1.04	0.79	0.011216	0.006480	0.004922
WV1012487	46	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	47	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1002066	4	45	1.80	1.04	0.79	0.006729	0.003888	0.002953
WV1002066	5	45	1.80	1.04	0.79	0.006729	0.003888	0.002953
WV1012487	4	45	1.80	1.04	0.79	0.001935	0.001118	0.000849
WV1012487	5	45	1.80	1.04	0.79	0.000967	0.000559	0.000425
WV1012487	31	45	1.80	1.04	0.79	0.004486	0.002592	0.001969
WV1012487	44	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1012487	45	45	1.80	1.04	0.79	0.003365	0.001944	0.001477
WV1012487	146	45	1.80	1.04	0.79	0.005608	0.003240	0.002461
WV1012487	150	45	1.80	1.04	0.79	0.003365	0.001944	0.001477
WV1012487	27	45	1.80	1.04	0.79	0.005608	0.003240	0.002461
WV1012487	48	45	1.80	1.04	0.79	0.008973	0.005184	0.003938
WV1012487	49	45	1.80	1.04	0.79	0.004486	0.002592	0.001969
WV1012487	91	45	1.80	1.04	0.79	0.011216	0.006480	0.004922
WV1012487	92	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1012487	93	45	1.80	1.04	0.79	0.006729	0.003888	0.002953
WV1012487	147	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	148	45	1.80	1.04	0.79	0.011216	0.006480	0.004922
WV1012487	29	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1012487	50	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	94	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	114	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	30	45	1.80	1.04	0.79	0.022431	0.012960	0.009845
WV1012487	51	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	52	45	1.80	1.04	0.79	0.007851	0.004536	0.003446
WV1012487	53	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	54	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1012487	95	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	96	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	97	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1012487	98	45	1.80	1.04	0.79	0.006729	0.003888	0.002953
WV1012487	99	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1012487	100	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1012487	103	45	1.80	1.04	0.79	0.003365	0.001944	0.001477
WV1012487	113	45	1.80	1.04	0.79	0.004486	0.002592	0.001969
WV1012487	25	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	55	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1012487	23	45	1.80	1.04	0.79	0.007851	0.004536	0.003446
WV1012487	24	45	1.80	1.04	0.79	0.010094	0.005832	0.004430
WV1012487	26	45	1.80	1.04	0.79	0.002243	0.001296	0.000984
WV1012487	56	45	1.80	1.04	0.79	0.001122	0.000648	0.000492
WV1012487	112	45	1.80	1.04	0.79	0.003365	0.001944	0.001477
WV1002074	26	46	3.20	2.00	4.30	0.012235	0.007647	0.016440
WV1002074	27	46	3.20	2.00	4.30	0.002719	0.001699	0.003653
WV1002074	28	46	3.20	2.00	4.30	0.032626	0.020391	0.043841
WV1002074	29	46	3.20	2.00	4.30	0.271881	0.169926	0.365340

Table E1. Allocated Metal Effluent Concentrations for Mining Point Sources in Paint Creek Watershed.

NPDES Permit	Outlet	Sub-WS	Wasteload Allocations			Wasteload Allocations		
			Iron (mg/L)	Manganese (mg/L)	Aluminum (mg/L)	Iron (Mg/yr)	Manganese (Mg/yr)	Aluminum (Mg/yr)
WV1002074	32	46	3.20	2.00	4.30	0.004078	0.002549	0.005480
WV1019317	1	46	3.20	2.00	4.30	0.004078	0.002549	0.005480
WV1002074	30	46	3.20	2.00	4.30	0.259647	0.162279	0.348900
WV1002074	31	46	3.20	2.00	4.30	0.014953	0.009346	0.020094
WV1014820	1	48	3.20	2.00	4.30	0.010875	0.006797	0.014614
WV0092142	2	52	2.43	1.54	1.15	0.032871	0.020832	0.015556
WV0092142	1	52	2.43	1.54	1.15	0.002529	0.001602	0.001197

Appendix F: Load Allocations.					
Table F1. Non-Point Source Acid Load Allocated for each Sub-Watershed.					
Stream	Sub-	Baseline	Allocated	Relative	
Name	WS	Acid Load	Acid Load	Acid Load	
	Number	(Mg/yr)	(Mg/yr)	Reduction	
Paint Creek below Banner Hollow	1	0.0000	0.0000	0%	
Banner Hollow	2	-26.8459	-26.8459	0%	
Paint Creek above Banner Hollow and below Fourmile Fork	3	0.0000	0.0000	0%	
Fourmile Fork	4	-16.9135	-16.9228	0%	
Paint Creek above Fourmile Fork and below Ash Branch	5	0.0000	0.0000	0%	
Ash Branch	6	-65.7423	-65.7450	0%	
Paint Creek above Ash Branch and below Toms Branch	7	0.0000	0.0000	0%	
Toms Branch	8	0.0000	0.0000	0%	
Paint Creek above Toms Branch and below Tenmile Branch	9	0.0000	0.0000	0%	
Long Branch	10	52.0795	0.0547	100%	
Tenmile Branch above Long Branch and below Unnamed Tributary	11	71.0703	0.0000	100%	
Unnamed Tributary of Tenmile Branch	12	-19.5621	-19.5621	0%	
Tenmile Branch above Unnamed Tributary	13	-20.7396	-20.7396	0%	
Paint Creek above Tenmile Branch and below Laurel Branch	14	0.0000	0.0000	0%	
Laurel Branch	15	1829.8900	591.6960	68%	
Paint Creek above Laurel Branch and below Unnamed Branch	16	0.0000	0.0000	0%	
Unnamed Branch of Paint Creek	17	-0.7872	-0.7872	0%	
Paint Creek above Unnamed Branch and below Hickory Camp Branch	18	0.0000	0.0000	0%	
Hickory Camp Branch	19	-1.4327	-1.4399	1%	
Paint Creek above Hickory Camp Branch and below Cedar Creek	20	0.0000	0.0000	0%	
Cedar Creek	21	6.8210	0.6322	91%	
Paint Creek above Cedar Creek and below Fifteenmile Creek	22	116.3010	0.0000	100%	
Fifteenmile Creek	23	-10.1211	-10.1211	0%	
Paint Creek above Fifteenmile Creek and below Unnamed Tributary	24	0.0000	0.0000	0%	
Spring Branch	25	7.3589	0.1004	99%	
Paint Creek above Unnamed Tributary and below Skitter Creek	26	0.0000	0.0000	0%	
Skitter Creek	27	-7.1882	-7.1882	0%	
Paint Creek above Skitter Creek and below Rattlesnake Run	28	0.0000	0.0000	0%	
Rattlesnake Run	29	0.0000	0.0000	0%	
Paint Creek above Rattlesnake Run and below Milburn Creek	30	0.0000	0.0000	0%	
Milburn Creek	31	-0.1830	-0.1830	0%	

Table F1. Non-Point Source Acid Load Allocated for each Sub-Watershed.					
Stream	Sub-WS	Baseline Acid Load	Allocated Acid Load	Relative Acid Load	
Name	Number	(Mg/yr)	(Mg/yr)	Reduction	
Paint Creek above Milburn Creek and below Lykins Creek	32	0.0000	0.0000	0%	
Lykins Creek	33	0.0000	0.0000	0%	
Paint Creek above Lykins Creek and below Bishop Fork	34	0.0000	0.0000	0%	
Bishop Fork	35	-12.0386	-12.0387	0%	
Paint Creek above Bishop Fork and below Mossy Creek	36	0.0000	0.0000	0%	
Mossy Creek below Lick Fork	37	0.0000	0.0000	0%	
Lick Fork	38	-6.4823	-6.4823	0%	
Mossy Creek above Lick Fork	39	-31.4732	-31.4732	0%	
Paint Creek above Mossy Creek and below Plum Orchard Creek	40	0.0000	0.0000	0%	
Plum Orchard Creek	41	-22.7147	-22.7147	0%	
Paint Creek above Plum Orchard Creek and below Horse Creek	42	0.0000	0.0000	0%	
Horse Creek	43	-7.8379	-7.8421	0%	
Paint Creek above Horse Creek and below Town Creek	44	0.0000	0.0000	0%	
Town Creek	45	-5.0554	-5.0554	0%	
Paint Creek above Town Creek and below Packs Branch	46	0.0000	0.0000	0%	
Packs Branch	47	-21.5686	-21.5686	0%	
Paint Creek above Packs Branch and below Dixons Branch	48	0.0000	0.0000	0%	
Dixons Branch	49	-12.0616	-12.0616	0%	
Paint Creek above Dixons Branch and below Sand Branch	50	0.0000	0.0000	0%	
Sand Branch below North Sand Branch	51	0.0000	0.0000	0%	
North Sand Branch below Maple Fork	52	0.0000	0.0000	0%	
Maple Fork	53	-21.3956	-21.3956	0%	
North Sand Branch above Maple Fork	54	-58.9777	-58.9777	0%	
South Sand Branch	55	-41.2540	-41.2540	0%	
Paint Creek above Sand Branch and below Laurel Branch	56	0.0000	0.0000	0%	
Laurel Branch	57	-2.9969	-3.2137	7%	
Paint Creek above Laurel Branch and below Davis Branch	58	0.0000	0.0000	0%	
Davis Branch	59	-0.4560	-0.4560	0%	
Paint Creek above Davis Branch and below Lefthand Fork	60	0.0000	0.0000	0%	
Lefthand Fork	61	-14.9771	-14.9771	0%	
Paint Creek above Lefthand Fork	62	-148.4890	-148.4890	0%	

Table F2. Non-Point Source Aluminum Load Allocated for each Sub-Watershed.					
	Sub-	Baseline	Allocated	Relative	
Stream	WS	Aluminum Load	Aluminum Load	Aluminum Load	
Name	Number	(Mg/yr)	(Mg/yr)	Reduction	
Paint Creek below Banner Hollow	1	0.6293	0.5657	10%	
Banner Hollow	2	0.1462	0.0457	69%	
Paint Creek above Banner Hollow and below Fourmile Fork	3	0.0012	0.0012	0%	
Fourmile Fork	4	0.0106	0.0106	0%	
Paint Creek above Fourmile Fork and below Ash Branch	5	2.2317	2.2317	0%	
Ash Branch	6	0.5519	0.0063	99%	
Paint Creek above Ash Branch and below Toms Branch	7	0.2631	0.2631	0%	
Toms Branch	8	0.1637	0.0164	90%	
Paint Creek above Toms Branch and below Tenmile Branch	9	0.4725	0.3911	17%	
Long Branch	10	11.7132	0.0079	100%	
Tenmile Branch above Long Branch and below Unnamed Tributary	11	0.0226	0.0117	48%	
Unnamed Tributary of Tenmile Branch	12	0.4968	0.0165	97%	
Tenmile Branch above Unnamed Tributary	13	1.3679	0.0304	98%	
Paint Creek above Tenmile Branch and below Laurel Branch	14	0.0012	0.0012	0%	
Laurel Branch	15	1.3277	0.0133	99%	
Paint Creek above Laurel Branch and below Unnamed Branch	16	0.0027	0.0027	0%	
Unnamed Branch of Paint Creek	17	0.1820	0.0029	98%	
Paint Creek above Unnamed Branch and below Hickory Camp Branch	18	0.0001	0.0001	0%	
Hickory Camp Branch	19	0.5227	0.0044	99%	
Paint Creek above Hickory Camp Branch and below Cedar Creek	20	0.1293	0.1293	0%	
Cedar Creek	21	1.4855	0.0135	99%	
Paint Creek above Cedar Creek and below Fifteenmile Creek	22	4.0173	0.9552	76%	
Fifteenmile Creek	23	0.0039	0.0039	0%	
Paint Creek above Fifteenmile Creek and below Unnamed Tributary	24	0.1362	0.1362	0%	
Spring Branch	25	1.6824	0.0112	99%	
Paint Creek above Unnamed Tributary and below Skitter Creek	26	0.0023	0.0023	0%	
Skitter Creek	27	0.1014	0.0101	90%	
Paint Creek above Skitter Creek and below Rattlesnake Run	28	0.0042	0.0042	0%	
Rattlesnake Run	29	0.0029	0.0029	0%	
Paint Creek above Rattlesnake Run and below Milburn Creek	30	0.0017	0.0017	0%	
Milburn Creek	31	0.7586	0.0091	99%	

Table F2. Non-Point Source Aluminum Load Allocated for each Sub-Watershed.					
Stream	Sub-WS	Baseline Aluminum Load	Allocated Aluminum Load	Relative Aluminum Load	
Name	Number	(Mg/yr)	(Mg/yr)	Reduction	
Paint Creek above Milburn Creek and below Lykins Creek	32	0.0012	0.0012	0%	
Lykins Creek	33	0.0021	0.0021	0%	
Paint Creek above Lykins Creek and below Bishop Fork	34	0.0041	0.0041	0%	
Bishop Fork	35	0.8125	0.0041	99%	
Paint Creek above Bishop Fork and below Mossy Creek	36	0.0001	0.0001	0%	
Mossy Creek below Lick Fork	37	0.0146	0.0146	0%	
Lick Fork	38	0.2920	0.0092	97%	
Mossy Creek above Lick Fork	39	0.3587	0.0145	96%	
Paint Creek above Mossy Creek and below Plum Orchard Creek	40	0.0068	0.0068	0%	
Plum Orchard Creek	41	0.1739	0.1739	0%	
Paint Creek above Plum Orchard Creek and below Horse Creek	42	0.0031	0.0031	0%	
Horse Creek	43	0.0982	0.0066	93%	
Paint Creek above Horse Creek and below Town Creek	44	0.0004	0.0004	0%	
Town Creek	45	0.1612	0.0055	97%	
Paint Creek above Town Creek and below Packs Branch	46	0.2679	0.2679	0%	
Packs Branch	47	0.3545	0.0603	83%	
Paint Creek above Packs Branch and below Dixons Branch	48	0.0098	0.0098	0%	
Dixons Branch	49	0.1810	0.0067	96%	
Paint Creek above Dixons Branch and below Sand Branch	50	0.0032	0.0032	0%	
Sand Branch below North Sand Branch	51	0.0017	0.0017	0%	
North Sand Branch below Maple Fork	52	0.0007	0.0007	0%	
Maple Fork	53	0.2997	0.0093	97%	
North Sand Branch above Maple Fork	54	0.7000	0.0117	98%	
South Sand Branch	55	0.1944	0.0351	82%	
Paint Creek above Sand Branch and below Laurel Branch	56	0.0006	0.0006	0%	
Laurel Branch	57	0.1057	0.0040	96%	
Paint Creek above Laurel Branch and below Davis Branch	58	0.0019	0.0019	0%	
Davis Branch	59	0.0755	0.0055	93%	
Paint Creek above Davis Branch and below Lefthand Fork	60	0.0023	0.0023	0%	
Lefthand Fork	61	0.1188	0.0048	96%	
Paint Creek above Lefthand Fork	62	1.3090	0.0073	99%	

Table F3. Non-Point Source Iron Load Allocated for each Sub-Watershed.						
	Sub-	Baseline	Allocated	Relative		
Stream	WS	Iron Load	Iron Load	Iron Load		
Name	Number	(Mg/yr)	(Mg/yr)	Reduction		
Paint Creek below Banner Hollow	1	0.4210	0.4210	0%		
Banner Hollow	2	0.0016	0.0016	0%		
Paint Creek above Banner Hollow and below Fourmile Fork	3	0.0003	0.0003	0%		
Fourmile Fork	4	0.0022	0.0022	0%		
Paint Creek above Fourmile Fork and below Ash Branch	5	1.4644	1.4644	0%		
Ash Branch	6	0.0013	0.0013	0%		
Paint Creek above Ash Branch and below Toms Branch	7	0.0370	0.0370	0%		
Toms Branch	8	0.0008	0.0008	0%		
Paint Creek above Toms Branch and below Tenmile Branch	9	0.1726	0.0517	70%		
Long Branch	10	0.3620	0.3620	0%		
Tenmile Branch above Long Branch and below Unnamed Tributary	11	0.0168	0.0168	0%		
Unnamed Tributary of Tenmile Branch	12	0.3510	0.0694	80%		
Tenmile Branch above Unnamed Tributary	13	1.0326	0.0820	92%		
Paint Creek above Tenmile Branch and below Laurel Branch	14	0.5219	0.0052	99%		
Laurel Branch	15	0.0005	0.0005	0%		
Paint Creek above Laurel Branch and below Unnamed Branch	16	1.1239	0.0112	99%		
Unnamed Branch of Paint Creek	17	0.0002	0.0002	0%		
Paint Creek above Unnamed Branch and below Hickory Camp Branch	18	0.0595	0.0006	99%		
Hickory Camp Branch	19	0.0084	0.0084	0%		
Paint Creek above Hickory Camp Branch and below Cedar Creek	20	0.3926	0.0039	99%		
Cedar Creek	21	0.3094	0.0307	90%		
Paint Creek above Cedar Creek and below Fifteenmile Creek	22	3.2347	0.3409	89%		
Fifteenmile Creek	23	0.0012	0.0008	35%		
Paint Creek above Fifteenmile Creek and below Unnamed Tributary	24	0.6187	0.0062	99%		
Spring Branch	25	0.0823	0.0082	90%		
Paint Creek above Unnamed Tributary and below Skitter Creek	26	0.9533	0.0095	99%		
Skitter Creek	27	0.0190	0.0019	90%		
Paint Creek above Skitter Creek and below Rattlesnake Run	28	1.7645	0.0176	99%		
Rattlesnake Run	29	0.0006	0.0006	0%		
Paint Creek above Rattlesnake Run and below Milburn Creek	30	0.7163	0.0072	99%		
Milburn Creek	31	0.3602	0.0360	90%		

Table F3. Non-Point Source Iron Load Allocated for each Sub-Watershed.					
Stream	Sub-WS	Baseline Iron Load	Allocated Iron Load	Relative Iron Load	
Name	Number	(Mg/yr)	(Mg/yr)	Reduction	
Paint Creek above Milburn Creek and below Lykins Creek	32	0.4916	0.0049	99%	
Lykins Creek	33	0.0004	0.0004	0%	
Paint Creek above Lykins Creek and below Bishop Fork	34	1.7046	0.0170	99%	
Bishop Fork	35	0.0673	0.0673	0%	
Paint Creek above Bishop Fork and below Mossy Creek	36	0.0398	0.0004	99%	
Mossy Creek below Lick Fork	37	2.6755	0.0307	99%	
Lick Fork	38	0.0162	0.0162	0%	
Mossy Creek above Lick Fork	39	1.9549	0.0150	99%	
Paint Creek above Mossy Creek and below Plum Orchard Creek	40	2.8793	0.0288	99%	
Plum Orchard Creek	41	1.2028	0.0097	99%	
Paint Creek above Plum Orchard Creek and below Horse Creek	42	1.3197	0.0013	100%	
Horse Creek	43	0.3990	0.0070	98%	
Paint Creek above Horse Creek and below Town Creek	44	0.1571	0.0016	99%	
Town Creek	45	0.2238	0.0023	99%	
Paint Creek above Town Creek and below Packs Branch	46	4.7979	0.0023	100%	
Packs Branch	47	1.3108	0.1227	91%	
Paint Creek above Packs Branch and below Dixons Branch	48	4.1438	0.0020	100%	
Dixons Branch	49	2.3775	0.0066	100%	
Paint Creek above Dixons Branch and below Sand Branch	50	1.3565	0.0094	99%	
Sand Branch below North Sand Branch	51	0.0003	0.0003	0%	
North Sand Branch below Maple Fork	52	0.0001	0.0001	0%	
Maple Fork	53	0.9272	0.0019	100%	
North Sand Branch above Maple Fork	54	1.2595	0.0059	100%	
South Sand Branch	55	18.5526	0.1169	99%	
Paint Creek above Sand Branch and below Laurel Branch	56	0.0997	0.0010	99%	
Laurel Branch	57	0.4359	0.0080	98%	
Paint Creek above Laurel Branch and below Davis Branch	58	0.3291	0.0033	99%	
Davis Branch	59	0.3475	0.0548	84%	
Paint Creek above Davis Branch and below Lefthand Fork	60	0.4138	0.0005	100%	
Lefthand Fork	61	0.3816	0.0010	100%	
Paint Creek above Lefthand Fork	62	2.6677	0.0015	100%	

Appendix G

Holistic Watershed Approach Protocol for Integrated Watershed Characterization

Holistic Watershed Approach Protocol for Integrated Watershed Characterization

(Adapted from Vukovich S.M. and G.E. Adolfson. 2001. Holistic Watershed Approach Protocol for Integrated Watershed Characterizations. Integrated Decision Making for Watershed Management Symposium. Washington, D.C.)

Background

Integrated watershed characterizations produce better environmental data and information to make more informed decisions about where and how we invest our resources toward watershed management of mine drainage pollution and associated Total Maximum Daily Load (TMDL) implementation. Involving local, state, and federal agencies; industry; academia; and the public in planning and sampling for watershed characterizations, has led to effective protection, restoration, and enhancement of the ecological integrity of water quality and quantity. Time, costs, knowledge, skills, and abilities are some of the limiting factors when attempting to perform these tasks separately for the desired ecological integrity. Inconsistencies in planning, sampling, and data collection methodologies create quality assurance and quality control concerns. A standard operating procedure, or protocol, eliminates these inconsistencies. Implementation of a protocol, in an integrated fashion, reduces limitations and promotes outreach, education, and training, as well as improves knowledge, skills, and abilities. The West Virginia Division of Environmental Protection's Stream Restoration Group currently implements a Holistic Watershed Approach Protocol involving diverse stakeholders in planning and sampling for integrated watershed characterizations in six of West Virginia's thirty-two hydrologic regions. The Protocol is a dynamic document continually evolving to accommodate multiple applications and satisfy specific needs of diverse stakeholders.

Methodology

When a watershed is designated for watershed characterization to determine impairment from mine drainage pollution discharges, the *study area* watershed boundaries are determined and stakeholders are notified. Watersheds are defined based on the USGS-developed hydrologic unit cataloging (HUC) system. Stakeholder involvement, spearheaded by watershed organizations, is incorporated into all aspects of watershed characterizations, including: restoration, protection, and enhancement.

With the assistance of the stakeholders, a *comprehensive sampling network* is established, mapped, and staked. This *network* includes sampling locations that divide the main-stem into segments representing changes in water quality from

upstream to downstream. Sampling locations at the mouth of all main-stem tributaries along with extensive sampling locations throughout the tributary stream reach are also included. Water quality and quantity measurements are obtained three to six times, spanning a range of hydrologic and climatologic conditions. Benthic macroinvertebrate surveys and fish surveys at selected locations are also collected during this time period.

If the watershed is large and dendritic, additional sampling of a *streamlined sampling network* is conducted. This consists of sampling locations of the main-stem and all the main-stem tributaries at the mouth locations only.

The environmental data and information is reviewed and main-stem tributaries are prioritized according to degree of impairment. A *focus area sampling network* of a selected main-stem tributary is then established and mapped. The *network* consists of sampling locations at the pollution sources as well as at various locations throughout the main-stem tributary reach. Sampling locations are determined by researching existing data and field reviewing the area for all sources of mine drainage pollution discharges. As with the *comprehensive sampling network*, water quality and quantity measurements are obtained three to six times, spanning a range of hydrologic and climatologic conditions. Benthic macroinvertebrate surveys are also collected during this time period.

The data is reviewed and utilized for: establishing the impact of the mine drainage pollution sources to the *focus area* tributaries, selecting the most feasible pollution sources within the *focus area* to address, and identifying the best available technology for the abatement or treatment of the pollution sources.

Following mine drainage pollution remediation of selected project sites within the *focus area*, a *post construction sampling network* is established. It consists of the same *focus area* locations sampled prior to construction, in addition to the treated discharges resulting from the installation of any mine drainage pollution abatement technologies. All new sampling site coordinates are obtained and mapped. Three to six water quality and quantity sampling sweeps are conducted spanning a range of hydrologic and climatologic conditions. Benthic macroinvertebrate surveys are also collected during this time period.

This process continues until all *focus areas* in the initial *study area* have been addressed, and all feasible treatment or abatement technologies applied. At that time, three to six water quality and quantity sampling sweeps of the initial *comprehensive sampling network* are conducted spanning a range of hydrologic and climatologic conditions. Benthic macroinvertebrate surveys and fish surveys are also collected during this time period.

Results are analyzed and a report prepared evaluating the effect of the abatement or treatment technologies on the mine drainage pollution sources and their receiving streams.

Once implemented, the Protocol is a perpetual cycle with many overlapping process steps. The Protocol outline and a process flowchart is presented below:

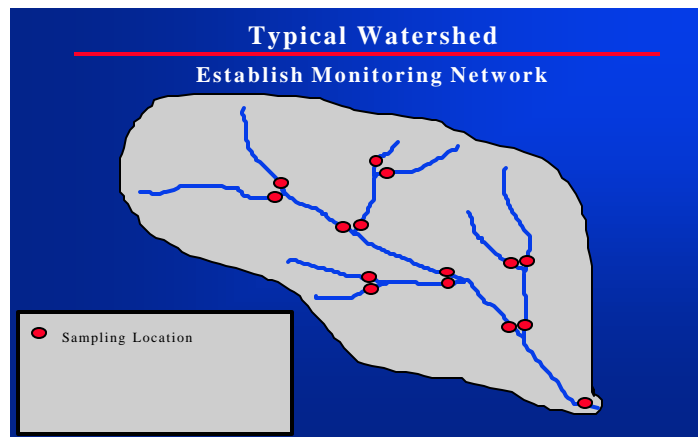
Holistic Watershed Approach Protocol

I. Define the *study area* and stakeholders.

- Select main-stem stream.
- Delineate watershed boundary.
- Foster Stakeholders.

II. Establish *comprehensive sampling network* within the *study area*.

- Select and number stream sampling stations utilizing USGS 7.5 Minute Topographic Quadrangle Maps and field reconnaissance.
 - Select main-stem stream sampling stations representing main-stem stream segments.
 - Select all main-stem tributary sampling stations at the mouth locations and at extensive locations throughout the main-stem tributary stream reach.



III. Geo-reference *comprehensive sampling network* for input into Geographical Information Systems (GIS).

IV. Implement sampling sweeps of the *comprehensive sampling network*.

- Conduct *Water Quality Study* sweeps three to six times spanning a range of hydrologic and climatologic conditions.
 - Perform water sample collection.
 - Collect stream water sample for laboratory analysis employing "grab" sample method.

- Perform field measurements.
 - Obtain *in situ* water quality measurements at all sampling stations.
 - Obtain stream flow.
- Conduct *Biological and Physical Study* one time between April and November.
 - Perform stream habitat assessments and qualitative benthic macroinvertebrate surveys at all stream sampling stations.
 - Perform fish survey at selective stream sampling stations only.

V. Review all data collected. (If watershed is large and dendritic, continue or otherwise skip to IX.)

- Analyze changes in tributary and main-stem stream segments and compare tributaries.
 - Represent *Water Quality Study* data graphically.
 - Compare *Biological and Physical Study* data.

VI. Establish *streamlined sampling network* within the *comprehensive sampling network*.

- Select and number stream sampling stations.
 - Select main-stem stream sampling stations representing main-stem stream segments.
 - Select all main-stem tributary sampling stations at the mouth locations only.

VII. Implement sampling sweeps of *streamlined sampling network*.

- Conduct *Water Quality Study* sweeps three to six times spanning a range of hydrologic and climatologic conditions.
 - Perform water sample collection.
 - Collect stream water sample for laboratory analysis employing “grab” sample method.
 - Perform field measurements.
 - Obtain *in situ* water quality measurements at all sampling stations.
 - Obtain stream flow.

VIII. Review all data collected.

- Analyze changes in tributary and main-stem stream segments and compare tributaries.
 - Represent *Water Quality Study* data graphically.
 - Compare *Biological and Physical Study* data.

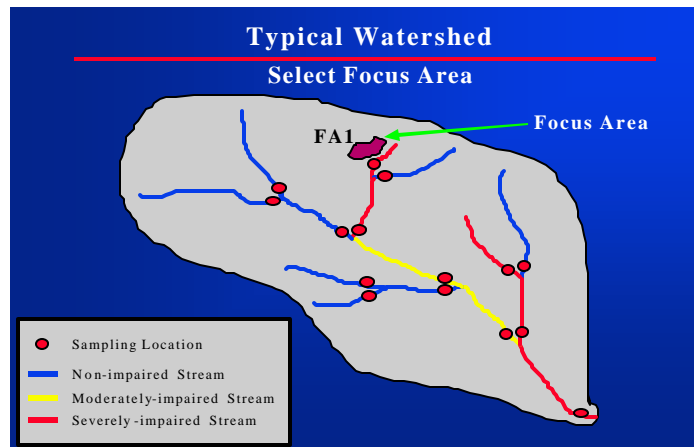
- Compare main-stem tributaries with respect to degree of impairment.

IX. Define *focus study area*.

- Select impaired tributary within *comprehensive sampling network* and determine watershed boundary.

X. Establish *focus area sampling network* within the *focus study area*.

- Locate mine drainage pollution discharge sampling stations within impaired tributary watershed.
 - Research existing data.
 - Field review entire impaired tributary watershed.
- Select impaired tributary sampling stations at mouth location and at extensive locations throughout the tributary stream reach, including stations upstream and downstream of mine drainage pollution discharge influx.
- Select receiving stream sampling stations upstream and downstream of the confluence with the impaired tributary.



XI. Geo-reference *focus area sampling network* for input into Geographical Information Systems (GIS).

XII. Implement sampling sweeps of *focus area sampling network*.

- Conduct *Water Quality Study* sweeps two to three times spanning a range of hydrologic and climatologic conditions.
 - Perform water sample collection.
 - Collect stream water sample for laboratory analysis employing “grab” sample method.

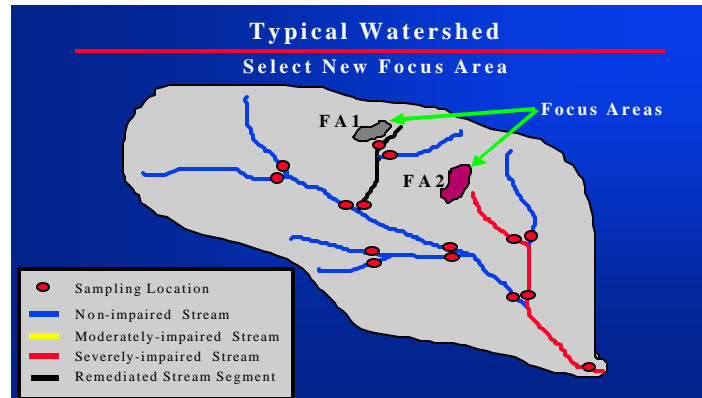
- Collect pollution source water sample at origin. (When several sources co-mingle, it is necessary to collect a sample of the combined discharge.)
- Perform field measurements.
 - Obtain *in situ* water quality measurements at all sampling stations.
 - Obtain stream flow.
- Conduct *Biological and Physical Study* one time between April and November.
 - Perform stream habitat assessments and qualitative benthic macroinvertebrate surveys upstream and downstream of mine drainage pollution discharge project areas.

XIII. Review all data collected.

- Analyze *focus area sampling* network data.
 - Determine extent of impairment mine drainage pollution discharge contributes to the *focus area* impaired tributaries.
 - Determine site-specific mine drainage pollution discharge treatment technology for the sources at each project area.
 - Evaluate chemical suitability of selected mine drainage pollution discharge treatment technology.
 - Evaluate physical suitability of selected mine drainage pollution discharge treatment technology.
 - Determine in-stream mine drainage pollution discharge treatment technology for stream benefits in addition to, or in lieu of site-specific pollution discharge treatment.

XIV. Modify *focus area sampling network*. [If additional data is or may be required to support pre construction design(s), repeat XII through XIII.]

- Cease sampling of any portion of project for which polluted water abatement appears infeasible.
- Incorporate sampling of any additional *focus area(s)* mine drainage pollution discharges found following completion of **XII**.



XV. Report findings.

- Prepare preliminary pre-design *Water Quality Study* report.

Implementation

XVI. Establish *post construction focus area sampling network* when mine drainage pollution discharge treatment is complete in the *focus study area*. (If initial *study area* contains other *focus study area(s)* that have not been addressed, repeat IX through XV, otherwise continue.)

- Locate constructed mine drainage pollution discharge treatment systems within treatment project boundaries.
 - Field review mine drainage pollution discharge treatment project site.
- Select and number stream sampling stations throughout *focus study area*.
 - Select the previously impaired tributary sampling stations at mouth location and at extensive locations throughout the tributary stream reach, including stations upstream and downstream of mine drainage pollution discharge treatment project influx.
 - Select receiving stream sampling stations upstream and downstream of the confluence with the previously impaired tributary.

XVII. Geo-reference *post construction focus area sampling network* for input into Geographical Information Systems (GIS).

XVIII. Implement sampling sweeps of *post construction focus area sampling network*.

- Conduct *Water Quality Study* sweeps monthly during the first year period; quarterly during the second year period; and semiannually

during the third and every subsequent year period spanning a range of hydrologic and climatologic conditions.

- Perform water sample collection.
 - Collect stream water sample for laboratory analysis employing “grab” sample method.
 - Collect untreated source water sample at origin if possible.
 - Collect treated source water sample at mine drainage pollution discharge treatment system outflow.
- Perform field measurements.
 - Obtain *in situ* water quality measurements at all sampling stations.
 - Obtain stream flow.
- Conduct *Biological and Physical Study* one time between April and November, at least one year after completion of project construction.
 - Perform stream habitat assessments and qualitative benthic macroinvertebrate surveys upstream and downstream of mine drainage pollution discharge treatment project influx.

XIX. Implement sampling sweeps of the *comprehensive sampling network*. (If mine drainage pollution discharge treatment is complete throughout initial *study area* continue.)

- Conduct *Water Quality Study* sweeps three to six times spanning a range of hydrologic and climatologic conditions.
 - Perform water sample collection.
 - Collect stream water sample for laboratory analysis employing “grab” sample method.
 - Perform field measurements.
 - Obtain *in situ* water quality measurements at all sampling stations.
 - Obtain stream flow.
- Conduct *Biological and Physical Study* one time between April and November.
 - Perform stream habitat assessments and qualitative benthic macroinvertebrate surveys at all stream sampling stations.
 - Perform fish survey at selective stream sampling stations only.

XX. Review all data collected.

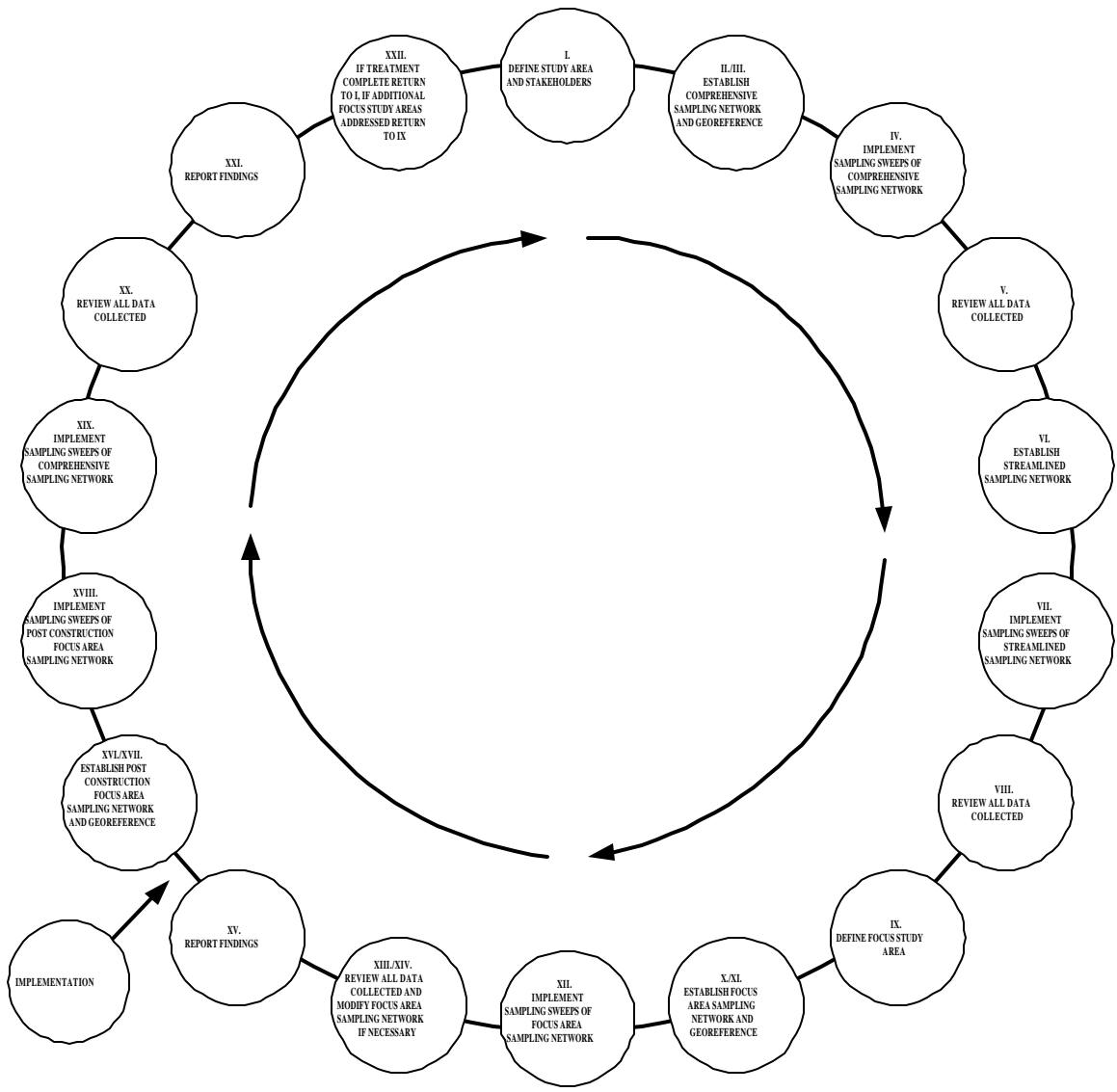
- Analyze changes in stream water quality.
- Analyze effectiveness and efficiency of constructed mine drainage pollution discharge treatment systems.
- Determine the effect of constructed mine drainage pollution discharge treatment systems on the mine drainage pollution discharges, *focus area sampling networks*, and *comprehensive sampling network*.

XXI. Report findings

- Prepare final post construction *Water Quality Study* report.

XXII. If mine drainage pollution discharge treatment is complete throughout the *study area*, return to I. If additional *focus study areas* will be addressed within the *study area*, return to IX.

Holistic Watershed Approach Protocol



Holistic Watershed Approach Protocol